



Water

Ambient Water Quality Criteria for

Pentachlorophenol - 1986



AMBIENT AQUATIC LIFE WATER QUALITY CRITERIA FOR
PENTACHLOROPHENOL

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FOREWORD

Section 304(a)(1) of the Clean Water Act of 1977 (P.L. 95-217) requires the Administrator of the Environmental Protection Agency to publish water quality criteria that accurately reflect the latest scientific knowledge on the kind and extent of all identifiable effects on health and welfare that might be expected from the presence of pollutants in any body of water, including ground water. This document is a revision of proposed criteria based upon consideration of comments received from other Federal agencies, State agencies, special interest groups, and individual scientists. Criteria contained in this document replace any previously published EPA aquatic life criteria for the same pollutant(s).

The term "water quality criteria" is used in two sections of the Clean Water Act, section 304(a)(1) and section 303(c)(2). The term has a different program impact in each section. In section 304, the term represents a non-regulatory, scientific assessment of ecological effects. Criteria presented in this document are such scientific assessments. If water quality criteria associated with specific stream uses are adopted by a State as water quality standards under section 303, they become enforceable maximum acceptable pollutant concentrations in ambient waters within that State. Water quality criteria adopted in State water quality standards could have the same numerical values as criteria developed under section 304. However, in many situations States might want to adjust water quality criteria developed under section 304 to reflect local environmental conditions and human exposure patterns before incorporation into water quality standards. It is not until their adoption as part of State water quality standards that criteria become regulatory.

Guidelines to assist States in the modification of criteria presented in this document, in the development of water quality standards, and in other water-related programs of this Agency, have been developed by EPA.

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Introduction*

Pentachlorophenol (PCP) and its sodium salt, sodium pentachlorophenate (NaPCP), are collectively the second most heavily used pesticide in the United States (Cirelli 1978). The principal uses of PCP and NaPCP are in the treatment of various wood products and as a wide-spectrum fungicide and bactericide. PCP** has been found in fresh and salt water at ng/L to mg/L concentrations (Buhler et al. 1973; Fontaine et al. 1976, Fox and Joshi 1984; Murray et al. 1981; Pierce 1978; Pierce et al. 1977; Renberg et al. 1983) with higher concentrations associated with point discharges. PCP has also been found in tissues of fish (De Vault 1985; Kuehl et al. 1980; Paasivirta et al. 1980,1983; Pierce 1978; Pierce et al. 1977; Veith et al. 1981; Zitko et al. 1974), in plankton, invertebrates, and sediment (DeLaune et al. 1983; Murray et al. 1980, Paasivirta et al. 1980; Pierce 1978; Pierce et al. 1977; Ray et al. 1983), and in humans (Bevenue et al. 1967; Dougherty 1978; Dougherty et al. 1980; Kuehl et al. 1980).

Several impurities are present in commercial-grade PCP, including lower chlorinated phenols (e.g., tetrachlorophenols) and chlorinated dibenzodioxins, dibenzofurans, diphenylethers, and 2-phenoxyphenols (Ahlborg and Thunberg 1980; Nilsson et al. 1978). The highly toxic 2,3,7,8-tetrachlorodibenzo-p-dioxin has not been found in PCP or NaPCP, and due to the methods of synthesis, is not expected to occur (Ahlborg

* An understanding of the "Guidelines for Deriving Numerical National Water Quality Criteria for the Protection of Aquatic Organisms and Their Uses" (Stephan et al. 1985), hereafter referred to as the Guidelines, and the response to public comment (U.S. EPA 1985a) is necessary in order to understand the following text, tables, and calculations.

** "PCP" is often used in the text to refer to the total amount of un-ionized pentachlorophenol and the pentachlorophenate ion that occurs in water regardless of whether PCP or NaPCP was initially added to water.

and Thunberg 1980). Cleveland et al. (1982) reported that a composite of three commercial PCP formulations contained much higher concentrations of hepta-, octa- and nonachlorophenoxyphenols, heptachlorodibenzodioxin, octachlorodibenzodioxin, and octachlorodibenzofuran, and was more toxic to fish, than either purified PCP (99% PCP) or Dowicide EC-7 (91% PCP). Dowicide EC-7 contained less of the hepta-, octa- and nonachlorophenoxyphenols than purified PCP, and was less toxic than purified PCP. Although the toxicities of the individual impurities are unknown and their concentrations apparently vary from batch to batch of PCP, Dowicide EC-7 seems to be an acceptable source of PCP for toxicity and bioconcentration tests. A more recent study has shown that ultrapure PCP is considerably less toxic than a mixture of the major impurities in commercial PCP, consisting mainly of octa- and nonachlorophenoxyphenols (Hamilton et al. 1986). Data concerning PCP should be reassessed as more information becomes available regarding the toxicities of various grades of PCP and their impurities to various species and the composition of the PCP that is produced, used, and discharged.

The toxicity of PCP to animals is due to the uncoupling of oxidative phosphorylation in the mitochondria (Ishak et al. 1970; Weinbach 1954,1956) and resultant reduced production of ATP. This is accompanied by an acceleration of metabolic rate and the utilization of tissue energy reserves, causing loss of weight (Cantelmo et al. 1978; Holmberg et al. 1972; Rao et al. 1979). Activities of various enzymes are also affected by PCP (Bostrom and Johansson 1972; Holmberg et al. 1972; Rao et al. 1979).

PCP is a weak acid with a pKa of about 10^{-5} (Blackman et al. 1955; Callahan et al. 1979; Cessna and Grover 1978; Mrak 1974). Consequently, its toxicity and potential for uptake by organisms are pH-dependent (Crandall and Goodnight 1959; Kobayashi and Kishino 1980; Saarikoski and

Viluksela 1981; Spenar et al. 1985). Both bioconcentration and toxicity increase as pH decreases, due to the greater penetration of cell membranes by un-ionized PCP molecules than by the pentachlorophenate ion. Transfer of PCP from water to fish is considered to be mainly by passive diffusion of the un-ionized form across the gill membrane (Kishino and Kobayashi 1980).

Unless otherwise noted, all concentrations reported herein are expressed as pentachlorophenol, not as the material tested. The criteria presented herein supersede previous aquatic life water quality criteria for PCP (U.S. EPA 1980) because these new criteria were derived using improved procedures and additional information. Whenever adequately justified, a national criterion may be replaced by a site-specific criterion (U.S. EPA 1983a), which may include not only site-specific criterion concentrations (U.S. EPA 1983b), but also site-specific durations of averaging periods and site-specific frequencies of allowed excursions (U.S. EPA 1985b). The latest comprehensive literature search for information for this document was conducted in July, 1986; some more recent information might have been included.

Acute Toxicity to Aquatic Animals

The acute toxicity of PCP to freshwater fish depends on the life stage of the fish and the pH, temperature, and concentration of dissolved oxygen in the water. Van Leeuwen et al. (1985) found that fry of rainbow trout were much more sensitive than embryos. PCP was more toxic to the fathead minnow (Crandall and Goodnight 1959) and to a nonresident fish, Notopterus notopterus, (Gupta et al. 1983b) at higher temperatures. Also, Goodnight (1942) observed that fish succumbed more rapidly at higher temperatures (28-30°C) than at cooler temperatures (9-24°C). Acute toxicity to Notopterus

notopterus increased with a decrease in the concentration of dissolved oxygen (Gupta et al. 1983a).

Of the variables studied, however, pH is the only one for which quantitative data are available for a variety of freshwater species. The acute value increases and toxicity decreases as pH increases. An analysis of covariance (Dixon and Brown 1979; Neter and Wasserman 1974) was performed using the natural logarithm of the acute value as the dependent variable, species as the treatment or grouping variable, and pH as the covariate or independent variable. This analysis of covariance model was fit to the data in Table 1 for the five freshwater species for which acute values are available over a range of pH. The slopes for all five species are between 0.67 and 1.2 (see end of Table 1). An F-test showed that, under the assumption of equality of slopes, the probability of obtaining five slopes as dissimilar as these is $P = 0.28$. This was interpreted as indicating that it is not unreasonable to assume that the slopes for the five species are the same.

The pooled slope of 1.005 was used with the data in Table 1 to adjust the acute values to $\text{pH} = 6.5$, where possible. Species Mean Acute Values were calculated as geometric means of the adjusted acute values, and Genus Mean Acute Values at $\text{pH} = 6.5$ (Table 3) were then calculated as geometric means of the available freshwater Species Mean Acute Values. Acute values are available for more than one species in each of three genera and the range of Species Mean Acute Values within each genus is less than a factor of 2.3. Of the 33 genera for which acute values are available, the most sensitive genus, Cyprinus, was over 10,000 times more sensitive than the most resistant, Orconectes. The freshwater Final Acute Value (FAV) for PCP at $\text{pH} = 6.5$ was calculated to be 10.97 $\mu\text{g/L}$ using the procedure

described in the Guidelines and the Genus Mean Acute Values in Table 3. The only Species Mean Acute Value that is lower than the FAV is that for the common carp, Cyprinus carpio. The freshwater Criterion Maximum Concentration (in $\mu\text{g/L}$) = $e^{[1.005(\text{pH})-4.830]}$.

Tests of the acute toxicity of PCP to resident North American saltwater animals have been performed with 13 species of invertebrates and five species of fish (Table 1). The range of acute values for invertebrates extends from 36.95 $\mu\text{g/L}$ for embryos of the American oyster, Crassostrea virginica (Borthwick and Schimmel 1978) to 18,000 $\mu\text{g/L}$ for the adult stage of the blue mussel, Mytilus edulis (Adema and Vink 1981). The range of acute values for saltwater fish is narrower, extending from 22.63 $\mu\text{g/L}$ for late yolk-sac larvae of the Pacific herring, Clupea harengus pallasi (Vigers et al. 1978) to 442 $\mu\text{g/L}$ for juvenile sheepshead minnow, Cyprinodon variegatus (Parrish et al. 1978). Fish appear to be generally more sensitive than invertebrates to PCP.

Embryos and larvae of the polychaete worm, Ophryotrocha diadema, and the mussel, Mytilus edulis, were more sensitive to PCP than adults of the same species (Adema and Vink 1981; Woelke 1972). Tests with Pacific herring, Clupea harengus pallasi, (Vigers et al. 1978) and the nonresident plaice, Pleuronectes platessa, (Adema and Vink 1981) revealed that sensitivity to PCP increases between the newly hatched yolk-sac larval stage and the late larval premetamorphosis stage. Juveniles and adults are slightly less sensitive than larvae.

Environmental factors, such as temperature, pH, and salinity, might have an effect on the acute toxicity of PCP to some saltwater animals. With the oligochaete worms, Limnodriloides verrucosus and Monophelephorus cuticulatus, sensitivity to PCP increased somewhat between 1°C and 10°C

at pH = 7, and between pH = 6 and pH = 8 at a temperature of 10°C (Chapman et al. 1982b). Larvae of the blue mussel, Mytilus edulis, were more sensitive to PCP at a salinity of 24 g/kg than at a salinity of 28 g/kg (Dimick and Breese 1965; Woelke 1972).

With some saltwater crustaceans, the stage of the molt cycle also seems to affect sensitivity to PCP. Late premolt and molting grass shrimp, Palaemonetes pugio, were more than five times as sensitive to PCP as intermolt animals (Conklin and Rao 1978a,b; Rao and Doughtie 1984). This effect was attributed to the greater permeability of the integument of the shrimp during molting.

Of the 17 saltwater genera for which acute values are available, the most sensitive genus, Clupea, is more than 247 times more sensitive than the most resistant, Crepidula (Table 3). The four most sensitive genera include two fish and two invertebrates and the range of the four Genus Mean Acute Values is a factor of 3.5. The saltwater Final Acute Value calculated from the values in Table 3 is 25.05 µg/L, which is very close to the acute value for the most sensitive tested saltwater species.

Chronic Toxicity to Aquatic Animals

Of the freshwater species with which chronic tests have been conducted on PCP, the cladoceran, Ceriodaphnia reticulata, is the most sensitive (Table 2), with a reduction in offspring occurring at 4.1 µg/L, the lowest concentration tested (Hedtke et al. 1986). Production of embryos by the snail, Physa gyrina, was reduced at a PCP concentration of 26 µg/L, the lowest concentration tested (Hedtke et al. 1986). A chronic value of 240 µg/L for Daphnia magna was based upon mortality (Adema 1978). Similarly, survival was the endpoint that determined the chronic values of 177 and 221 µg/L for Simocephalus vetulus (Hedtke et al. 1986).

Significant increases in mortality and decreases in length and weight occurred among rainbow trout at 19 $\mu\text{g/L}$ after a 72-day exposure (Dominguez and Chapman 1984), but the trout were not affected at 11 $\mu\text{g/L}$. Dominguez and Chapman used a pure form of PCP (99+%) and found that yolk sac edema and cranial malformations were rare as compared to their more common incidences in studies with technical-grade PCP.

Several studies with the fathead minnow in the pH range from 6.5 to 8.5 resulted in chronic values from 24 to 144 $\mu\text{g/L}$ (Table 2). Spehar et al. (1985) studied the relationship between pH and chronic toxicity, and obtained chronic values of 24, 40, 49, and 89 $\mu\text{g/L}$ at pH = 6.5, 7.5, 8.0, and 8.5, respectively. A linear regression of $\ln(\text{chronic value})$ on pH resulted in a slope of 0.6174. This slope is similar to the slope of 0.6782 obtained by the same investigators in an acute toxicity study with the fathead minnow. However, because data are available for chronic toxicity versus pH for only one species, the acute pooled slope of 1.005 was also applied to the freshwater chronic data in Table 2, where possible, to adjust the chronic values to pH = 6.5, to allow a comparison of freshwater chronic values at a common pH.

The long-term sublethal toxicity of PCP to fish has been shown to be affected by temperature and concentration of dissolved oxygen (Table 6). Newly hatched rainbow trout were affected more at 6°C than at 10°C, whereas after yolk sac sorption, a greater effect on growth occurred at 20°C than at 12°C (Hodson and Blunt 1981). They concluded that, if these laboratory data were applied in the field to wild trout populations, temperature effects on PCP toxicity would be greatest both during embryo development and subsequent growth of young trout.

Decreased concentrations of dissolved oxygen in the water increased the toxicity of PCP to young rainbow trout (Chapman and Shumway 1978). After a 24-day exposure to 37 $\mu\text{g/L}$, alevins reared at dissolved oxygen concentrations of 3, 5, and 10 mg/L suffered 76, 20, and 7% mortality, respectively. Also, the time required for trout to attain maximum growth was prolonged at the two lower concentrations of dissolved oxygen.

Usable data, according to the Guidelines, on the chronic toxicity of PCP are available for only two saltwater species, the polychaete worm, Ophryotrocha diadema, (Table 6) and the sheepshead minnow, Cyprinodon variegatus, (Table 2). Test populations of Ophryotrocha diadema were exposed to PCP in a renewal life-cycle test extending from 2 to 3-day-old larvae to two-week-old second generation larvae (Hooftman and Vink 1980). The most sensitive effect was an apparent inhibition of reproduction at 11 $\mu\text{g/L}$. This value was not used in Table 2 because the authors did not report any statistical analyses of the data. A complete life-cycle test was conducted with the sheepshead minnow, Cyprinodon variegatus (Parrish et al. 1978). The most sensitive effect was decreased long-term survival of the first generation fish at 88 $\mu\text{g/L}$. At 195 $\mu\text{g/L}$, survival of second generation embryos and juveniles was reduced.

The seven available Species Mean Acute-Chronic Ratios range from 0.8945 to over 15.79 (Table 3), but the two highest values are both "greater than" values and were obtained with two freshwater species that are much more acutely resistant than the other four freshwater species. Therefore, the Final Acute-Chronic Ratio of 3.166 was calculated as the geometric mean of the five Species Mean Acute-Chronic Ratios ranging from 0.8945 to 6.873. When this ratio is used with the freshwater Final Acute Value and the pooled slope for the pH-toxicity relationship (Table 3), the resulting

freshwater Final Chronic Value (in $\mu\text{g/L}$) = $e^{[1.005(\text{pH})-5.290]}$. Division of the saltwater Final Acute Value by the Final Acute-Chronic Ratio results in a saltwater Final Chronic Value of $7.912 \mu\text{g/L}$. This is about a factor of 8 below the only available saltwater chronic value, but no chronic test on PCP has been conducted with any of the 13 acutely most sensitive saltwater species.

Toxicity to Aquatic Plants

Freshwater plants are sensitive to PCP over a range of concentrations from 7.5 to $3,200 \mu\text{g/L}$ (Tables 4 and 6). Blackman et al. (1955) observed that the degree of chlorosis in a 48-hr exposure of Lemna minor was very sensitive to small changes in PCP concentrations. Only a slight concentration increase was required to induce 50% chlorosis in all of the duckweed fronds as compared to 50% induction in none of the fronds. In a prolonged exposure of the vascular plant, Elodea canadensis, significant biomass reductions occurred at progressively lower concentrations of PCP as duration of exposure increased (Hedtke et al. 1986). Biomass reductions were detected at $1,440 \mu\text{g/L}$ after 7 days, $810 \mu\text{g/L}$ after 14 days, and $380 \mu\text{g/L}$ after 21 days (Table 6). Biomass was not affected at any time by the next lower concentration of $230 \mu\text{g/L}$. In contrast, neither reduced frond production nor chlorosis was observed in Lemna minor at the highest PCP concentration of $1,440 \mu\text{g/L}$ (Hedtke et al. 1986). The river water used in both of these tests might have had some effect upon the results.

Complete destruction of chlorophyll in Chlorella pyrenoidosa occurred at a PCP concentration of $7.5 \mu\text{g/L}$ (Huang and Gloyna 1967). In a 7-day study with Selenastrum capricornutum, cell numbers and population growth rates were significantly reduced by $50 \mu\text{g/L}$, but were not affected by 10

µg/L (Adams et al. 1985). Concentrations near 300 µg/L resulted in 50% reductions of Selenastrum populations in two separate studies (Crossland and Wolff 1985; Richter 1982). The 4-day EC50 for Scenedesmus subspicatus was 90 µg/L (Geyer et al. 1985).

Usable data on the toxicity of PCP to saltwater plants are available for eight phytoplankters, one macroalgal species, and one vascular plant (Table 4). The 96-hr EC50s, based on reduction of cell population growth, range from 17.40 µg/L for the diatom, Skeletonema costatum (Walsh et al. 1982) to 3,600 µg/L for the green alga, Dunaliella tertiolecta (Adema and Vink 1981). Giant kelp, Macrocystis pyrifera, and seagrass, Thalassia testudinum, are about as sensitive as the phytoplankters. The range of sensitivities of saltwater plants to PCP is similar to that for saltwater animals and, therefore, a criterion that protects saltwater animals will probably also protect saltwater plants.

Bioaccumulation

Bioconcentration of PCP from water, like toxicity, has been shown to be inversely related to pH (Kobayashi and Kishino 1980; Spehar et al. 1985). PCP bioconcentrated in the tissues of fish from 7.3 to 1,066 times (Table 5), with test durations from 16 to 115 days. The gall bladder concentrated the highest levels of PCP (Glickman et al. 1977, Kobayashi and Akitake 1975b, McKim et al., Manuscript), whereas muscle and skin contained the lowest concentrations of PCP in rainbow trout exposed to 0.78 to 1.15 µg/L (McKim et al., Manuscript). The lowest bioconcentration factor (BCF) of 7.3 was obtained in bluegill muscle (Pierce 1978; Pruitt et al. 1977). BCFs of 320 and 378 were estimated from uptake and depuration rates using the rainbow trout (McKim et al., Manuscript) and the non-resident killifish, Oryzias latipes (Sugiura et al. 1984).

Residues of PCP in fish drop quite rapidly upon termination of exposure. Ninety-six percent of whole body ¹⁴C-labelled PCP was eliminated by fathead minnows within 3.5 days (Huckins and Petty 1983), whereas about 85 percent of the PCP residues in bluegill muscle were eliminated in 4 days (Pruitt et al. 1977). A first-order simulation model developed from empirical data indicated a half-life of 2.7 days in rainbow trout, with 95% elimination in 11.7 days (McKim et al., Manuscript).

McKim et al. (1985) studied the efficiency of chemical uptake by rainbow trout gills using 14 different chemicals including PCP. They found that PCP was in a group of chemicals, all with log n-octanol/water partition coefficients (log P) between 2.84 and 6.18, that were taken up more efficiently by gills than chemicals with either lower or higher partition coefficients. The rate of elimination of the chemicals in this log P range largely determined their BCFs. In this range, the elimination rates decreased as log P increased. PCP, with a log P of 3.32, would be expected to be eliminated quite readily. Indeed, this has been shown, as indicated above.

Several studies have shown that PCP is conjugated with glucuronic acid in fish (Huckins and Petty 1983; Kobayashi 1978,1979; Kobayashi and Nakamura 1979b; Kobayashi et al. 1977; Lech et al. 1978). Reduced glucuronidation occurred in fathead minnows exposed to industrial PCP, as compared to purified PCP, and it was suggested that this might play a role in the elevated toxicity of the impure form (Huckins and Petty 1983). Pentachlorophenylsulfate has been found in goldfish exposed to PCP (Akitake and Kobayashi 1975; Kobayashi 1978,1979; Kobayashi and Nakamura 1979a,b; Kobayashi et al. 1984), but not in rainbow trout (Lech et al. 1978) or fathead minnows (Huckins and Petty 1983).

Probable steady-state BCFs for PCP are available for the eastern oyster, Crassostrea virginica (Schimmel and Garnas 1985; Schimmel et al. 1978), the sheepshead minnow, Cyprinodon variegatus (Parrish et al. 1978) and the longnose killifish, Fundulus similis (Trujillo et al. 1982). In tests with oysters, the steady-state BCF ranged from 34 to 82, and was reached in 14 to 96 hours. In a life-cycle test with the sheepshead minnow, BCFs were 5 to 27 with adult fish, 13 to 22 with embryos, and 16 to 48 with 28-day-old young (Schimmel et al. 1978). A steady-state BCF of 64 was obtained in a 7-day test with the longnose killifish (Trujillo et al. 1982). In four-day acute toxicity tests, Schimmel et al. (1978) found BCFs ranging from 0.26 for brown shrimp to 38.0 for striped mullet, Mugil cephalus. Using radio-labeled PCP, Carr and Neff (1981) reported a 14-day BCF of 280 with the sand worm, Nereis virens.

No U.S. FDA action level or other maximum acceptable concentration in tissue is available for pentachlorophenol, and, therefore, no Final Residue Value can be calculated.

Other Data

Cleveland et al. (1982) and Hamilton et al. (1986) studied the effects of four samples of PCP on the survival and growth of fathead minnows for 90 days. In the range of 60 to 142 µg/L, both a purified PCP and an ultrapure PCP reduced growth by 10 to 25%, whereas Dowicide EC-7 increased growth by 18 to 21%. An industrial composite PCP was much more toxic; 13 µg/L reduced growth by 20%, 27 µg/L reduced growth by 40%, and 67 µg/L killed all the fish. Analyses of the four PCP samples for fourteen polychlorinated impurities and a test on the effects of a mixture of some of the impurities indicated that some of the effects seen in the tests on PCP were probably due to the measured impurities.

Webb and Brett (1973) found that growth and food conversion of sockeye salmon (Oncorhynchus nerka) were affected in a 56-day exposure to PCP concentrations of 1.74 and 1.80 $\mu\text{g/L}$, respectively, at a pH of 6.8 (Table 6). These are the lowest effect concentrations that have been reported for any species. Growth of salmonids was reduced by 10 to 27% at PCP concentrations ranging from 3.2 to 28 $\mu\text{g/L}$ (Chapman 1969; Chapman and Shumway 1978; Matida et al. 1976). The thresholds for reduction of growth of rainbow trout exposed to PCP through 4-week post swim-up were about 10 and 20 $\mu\text{g/L}$ in warm and cold regimes, respectively (Hodson and Blunt 1981). Reduced growth of largemouth bass, Micropterus salmoides, was observed at 50.4 $\mu\text{g/L}$ after 7 days of exposure (Mathers et al. 1985), of the guppy at 320 $\mu\text{g/L}$ after 28 days of exposure and of the amphibian, Xenopus laevis, at 100 $\mu\text{g/L}$ after 100 days of exposure (Slooff and Canton 1983).

Whitley (1968) and Kobayashi and Kishino (1980) reported that an increase in pH decreased the 24-hr LC50s for both tubificid worms and goldfish. Some reduction in toxicity appeared to result from the presence of sediment in acute toxicity tests conducted with several species of tubificid worms (Chapman et al. 1982a). Similarly, Mississippi River water afforded some protection to various invertebrate species when their LC50s were compared with values obtained in Lake Superior water (Hedtke et al. 1986). The tubificids, Limnodrilus hoffmeisteri and Tubifex tubifex, were more resistant to PCP when tested as mixed species than when each species was tested individually (Chapman et al. 1982a,b). Schauerte et al. (1982) found that 1,000 $\mu\text{g/L}$ decreased daphnids and autotrophic phytoplankton in compartments in a pond, and Yount and Richter

(1986) found that periphyton biomass in experimental streams was inversely related to pentachlorophenol in the range from 48 to 432 $\mu\text{g/L}$.

A wide variety of acute, chronic, and sublethal effects of PCP have been reported for saltwater organisms (Table 6). Several effects of PCP on polychaete worms have been reported such as reduced feeding activity by Arenicola cristata (Rubinstein 1978). depletion of glycogen reserves, increased tissue ascorbic acid concentrations, and disruption of osmoregulation Nereis virens (Carr and Neff 1981); and reproductive impairment in Ophryotrocha diadema (Hooftman and Vink 1980). PCP induced developmental abnormalities in embryos and larvae of the mussel, Mytilus edulis, (Dimick and Breese 1965) and oyster, Crassostrea virginica, (Davis and Hidu 1969) and inhibited shell growth of juvenile oysters (Schimmel et al. 1978). Exposure for 18 weeks to 460 $\mu\text{g/L}$ resulted in reduced resistance of the quahog clam, Mercenaria mercenaria, to bacterial infection (Anderson et al. 1981).

The grass shrimp, Palaemonetes pugio, exhibited a variety of sublethal physiological effects during exposure to PCP concentrations of 400 to 4,600 $\mu\text{g/L}$ (Brannon and Conklin 1978; Cantelmo et al. 1978; Doughtie and Rao 1978; Rao and Doughtie 1984; Rao et al. 1978, 1979, 1981) such as histological damage to epithelia of the gills, gut, and hepatopancreas, increased metabolic rate, decreased rate of limb regeneration, and increased weight of the molted exoskeleton.

Saltwater fish exposed to 50 to 200 $\mu\text{g/L}$ exhibited elevation of the concentration of acid-soluble thiols in liver of winter flounder (Thomas and Wofford 1984) and elevation of blood cortisol, leading to hyperglycemia, depletion of liver glycogen reserves, and an increase in liver ascorbic acid concentration in tullet (Thomas et al. 1981).

In a series of benthic colonization studies conducted in the laboratory under flow-through conditions, PCP concentrations of 55 to 140 $\mu\text{g/L}$ reduced both species richness and total faunal abundances of benthic macrofauna (Hansen and Tagatz 1980; Tagatz et al. 1977, 1980, 1981). At a concentration of 15.8 $\mu\text{g/L}$, PCP, administered as Dowicide G-ST, reduced total faunal abundances, but not species richness. Molluscs were generally most sensitive, but other phyla were also affected.

Unused Data

Some data on the effects of PCP on aquatic organisms were not used because the studies were conducted with species that are not resident in North America (e.g., Adema and Vink 1981; Dalela et al. 1980a,b,c, Dave 1984; Hanumante and Kulkarni 1979; Goel and Prasad 1978; Gupta 1983; Gupta and Dalela 1986; Gupta and Durve 1984a,b, Gupta and Rao 1982; Gupta et al. 1982b, 1983a,b; Hattori et al. 1984; Kaila and Saarikoski 1977, Khangarot et al. 1985; Kobayashi et al. 1969, Nagendran and Shakuntala 1979; Rao et al. 1983; Shim and Self 1973; Slooff 1978; Slooff et al. 1983b; Tomiyama et al. 1962; Van Dijk et al. 1977; Verma et al. 1980, 1981a,b, 1982, 1984). Results (e.g., Adema and Vink 1981) of tests conducted with brine shrimp, Artemia sp., were not used because these species are from a unique saltwater environment. Adelman et al. (1976b), Ahlborg and Thurnberg (1980), Alexander et al. (1983), Bevenue and Beckman (1967), Buikema et al. (1979), Conklin and Fox (1978), Hall and Kier (1984a,b), Koch (1982), Kozak et al. (1979), National Research Council of Canada (1982), Rao et al. (1979), von Rumker et al. (1974), and Strufe (1968) only contain data that have been published elsewhere.

Results were not used if either the test procedures or the test material was not adequately described (e.g., Benoit-Guyod et al. 1984a;

Canton and Slooff 1979; Clemens and Sneed 1959; Hashimoto and Nishiuchi 1983; Klein et al. 1984; Knie et al. 1983; Konemann and Musch 1981; Wong 1984), or if PCP was a component of a mixture (Hermens et al. 1985; Statham and Lech 1975). Tests were not used if PCP comprised only eight percent of the formulation (Batte and Swanson 1952; Inglis and Davis 1972). Data were not used if PCP was a component of a sediment (e.g., D'Asaro and Wilkes 1982) or if the organisms were exposed by injection (e.g., Bose and Fujiwara 1978; Tripp et al. 1984) or in food (e.g., Niimi and Cho 1983). Anderson et al. (1984), Bols et al. (1985), Cantelmo and Rao (1978a), Fox and Rao (1978), and Kwasniewska et al. (1979) only exposed enzymes, excised or homogenized tissue, cell cultures, or sewage bacteria.

Tests conducted with too few test organisms (e.g., Coglianese and Neff 1982; McLeese et al. 1979) and tests in which the concentrations fluctuated widely (e.g., Thomas et al. 1981) were not used. The 60-day test reported by Verma et al. (1981c) was not used because there were no replicate test chambers. The early life-stage toxicity data of Johansen et al. (1985) were not used due to high mortality of control fish and an interruption in the exposure. Results were not used if organisms were not cultured and tested in the same dilution water (Berglund and Dave 1984).

Studies with physiological endpoints only were not used (e.g., Bostrom and Johansson 1972; Chowdary et al. 1979; Gupta et al. 1983c,d; Hanke et al. 1983; Holmberg et al. 1972; Huber et al. 1982; Jayaweera et al. 1982; Kaila 1982; Kaila and Saarikoski 1980,1981; LeBlanc and Cochrane 1985; Liu 1981; Oikari and Nitryla 1985; Oikari et al. 1985; Peer et al. 1983; Saarikoski and Kaila 1977; Sarojini et al. 1983; Sloley et al. 1986; Tiedge et al. 1986; Verma 1981b; Verma et al. 1982; Yousri and Hanke 1985). A study of histological effects on bluegills was not used (Owen and Rosso

1981). Toxicity data were not used if they were only qualitative (Palmer and Maloney 1955) or were presented in graphic form (Norup 1972). A study by Anderson and Weber (1975) was not used because acute toxicity results for the guppy were presented only as linear regressions on body size.

A study on the uptake and metabolism of PCP in rice plants (Weiss et al. 1982) was not used, nor were uptake studies in which exposure was via the food (Niimi and Cho 1983) or by gavage (Niimi and Palazzo 1985). Bioconcentration studies were not used if the test was not flow-through or renewal (e.g., Ernst 1979) or if the exposures were of insufficient duration for steady state to have been achieved (Glickman et al. 1977; Kobasyahi and Akitake 1975a,b; Kobayashi et al. 1979; Kuehl et al. 1983; Lech et al. 1978; McKim et al., Manuscript). Reports of the concentrations of PCP in wild aquatic organisms (e.g., Butte et al. 1983; Faas and Moore 1979; Folke et al. 1983; Fox and Joshi 1984; Kuehl and Dougherty 1980; Murray et al. 1980,1981; Ray et al. 1983) were not used if the number of measurements of the concentration in water was too small or if the range of the measured concentrations in water was too great. Studies of the concentration or accumulation of PCP in organisms were not used if the data provided were considered insufficient for determination of a bioconcentration or bioaccumulation factor (Gotham and Rhee 1982; Hallas 1973; Klein et al. 1984; Korte et al. 1978; Oikari 1986; Oikari and Anas 1985; Paasivirta et al. 1981), if the data were being reported secondarily (Branson 1980; Davies and Dobbs 1984), if water and biota samples were not collected at the same times (Metcalf et al. 1984), or if the measurements were of total radiolabel rather than PCP itself (Fisher and Wadleigh 1986, Freitag et al. 1985; Geyer et al. 1981,1984; Gluth et al. 1985). Bioaccumulation and fate data

from model ecosystems or microcosm studies were not used (e.g., Brockway et al. 1984; Knowlton and Huckins 1983; Lu and Metcalf 1975; Robinson-Wilson et al. 1983; Tomizawa and Kazano 1979).

Summary

The acute and chronic toxicity of PCP to freshwater animals increased as pH and dissolved oxygen concentration of the water decreased. Generally, toxicity also increased with increased temperature. The estimated acute sensitivities of 36 species at pH = 6.5 ranged from 4.355 $\mu\text{g/L}$ for larval common carp to >43,920 $\mu\text{g/L}$ for a crayfish. At pH = 6.5, the lowest and highest estimated chronic values of <1.835 and 79.66 $\mu\text{g/L}$, respectively, were obtained with different cladoceran species. Chronic toxicity to fish was affected by the presence of impurities, with industrial-grade PCP being more toxic than purified samples. Mean acute-chronic ratios for six freshwater species ranged from 0.8945 to >15.79, but the mean ratios for the four most acutely sensitive species only ranged from 0.8945 to 5.035. Freshwater algae were affected by concentrations as low as 7.5 $\mu\text{g/L}$, whereas vascular plants were affected at 189 $\mu\text{g/L}$ and above. Bioconcentration factors ranged from 7.3 to 1,066 for three species of fish.

Acute toxicity values from tests with 18 species of salt-water animals, representing 17 genera, range from 22.63 $\mu\text{g/L}$ for late yolk-sac larvae of the Pacific herring, Clupea harengus pallasii, to 18,000 $\mu\text{g/L}$ for adult blue mussels, Mytilus edulis. The embryo and larval stages of invertebrates and the late larval premetamorphosis stage of fish appear to be the most sensitive life stages to PCP. With few exceptions, fish are more sensitive than invertebrates to PCP. Salinity, temperature, and pH have a slight effect on the toxicity of PCP to some saltwater animals.

Life-cycle toxicity tests have been conducted with the sheepshead minnow, Cyprinodon variegatus, and the polychaete worm, Ophryotrocha diadema. The chronic value for the minnow is 64.31 $\mu\text{g/L}$ and the acute-chronic ratio is 6.873. Unfortunately, no statistical analysis of the data from the test with the worm is available.

The EC50s for saltwater plants range from 17.40 $\mu\text{g/L}$ for the diatom, Skeletonema costatum, to 3,600 $\mu\text{g/L}$ for the green algae, Dunaliella tertiolecta. Apparent steady-state BCFs are available for the eastern oyster, Crassostrea virginica, and two saltwater fishes and range from 10 to 82.

National Criteria

The procedures described in the "Guidelines for Deriving Numerical National Water Quality Criteria for the Protection of Aquatic Organisms and Their Uses" indicate that, except possibly where a locally important species is very sensitive, freshwater aquatic organisms and their uses should not be affected unacceptably if the four-day average concentration (in $\mu\text{g/L}$) of pentachlorophenol does not exceed the numerical value given by $e^{[1.005(\text{pH})-5.290]}$ more than once every three years on the average and if the one-hour average concentration (in $\mu\text{g/L}$) does not exceed the numerical value given by $e^{[1.005(\text{pH})-4.830]}$ more than once every three years on the average. For example, at pH = 6.5, 7.8, and 9.0 the four-day average concentrations of pentachlorophenol are 3.5, 13, and 43 $\mu\text{g/L}$, respectively, and the one-hour average concentrations are 5.5, 20, and 68 $\mu\text{g/L}$. At pH = 6.8, a pentachlorophenol concentration of 1.74 $\mu\text{g/L}$ caused a 50% reduction in the growth of yearling sockeye salmon in a 56-day test.

The procedures described in the "Guidelines for Deriving Numerical National Water Quality Criteria for the Protection of Aquatic Organisms and Their Uses" indicate that, except possibly where a locally important species is very sensitive, saltwater aquatic organisms and their uses should not be affected unacceptably if the four-day average concentration of pentachlorophenol does not exceed 7.9 $\mu\text{g/L}$ more than once every three years on the average and if the one-hour average concentration does not exceed 13 $\mu\text{g/L}$ more than once every three years on the average.

Three years is the Agency's best scientific judgment of the average amount of time aquatic ecosystems should be provided between excursions (U.S. EPA 1985b). The resiliencies of ecosystems and their abilities to recover differ greatly, however, and site-specific allowed excursion frequencies may be established if adequate justification is provided.

Use of criteria for developing water quality-based permit limits and for designing waste treatment facilities requires selection of an appropriate wasteload allocation model. Dynamic models are preferred for the application of these criteria (U.S. EPA 1985b). Limited data or other considerations might make their use impractical, in which case one must rely on a steady-state model (U.S. EPA 1986).

Table 1. Acute Toxicity of Pentachlorophenol to Aquatic Animals

Species	Method [#]	Chemical ^{**}	pH	LC50 or EC50 ($\mu\text{g/L}$) ^{***}	Adjusted LC50 or EC50 ($\mu\text{g/L}$) ^{****}	Species Mean Acute Value ($\mu\text{g/L}$) ^{*****}	Reference
<u>FRESHWATER SPECIES</u>							
<u>Hydra</u> , <u>Hydra oligactis</u>	S, U	PCP (>98%)	-	730	-	-	Slooff et al. 1983a
Tubificid worm, <u>Branchiura sowerbyi</u>	R, U	NaPCP	7.0	257.7	155.9	155.9	Chapman et al. 1982a
Tubificid worm, <u>Limnodrilus hoffmeisteri</u>	R, U	NaPCP	7.0 (10°C)	303.8	183.8	-	Chapman et al. 1982a,b,c
Tubificid worm, <u>Limnodrilus hoffmeisteri</u>	R, U	NaPCP	7.0 (1°C)	478.7 [†]	289.6	-	Chapman et al. 1982b
Tubificid worm, <u>Limnodrilus hoffmeisteri</u>	R, U	NaPCP	7.0 (20°C)	451.0 [†]	272.9	-	Chapman et al. 1982b
Tubificid worm, <u>Limnodrilus hoffmeisteri</u>	R, U	NaPCP	6.0 (10°C)	303.8 [†]	502.1 ^{††}	-	Chapman et al. 1982b
Tubificid worm, <u>Limnodrilus hoffmeisteri</u>	R, U	NaPCP	8.0 (10°C)	345.2 [†]	76.45	182.5	Chapman et al. 1982b
Tubificid worm, <u>Quistadrilus multisetosus</u>	R, U	NaPCP	7.0	524.7	317.5	317.5	Chapman et al. 1982a,b
Tubificid worm, <u>Rhyacodrilus montana</u>	R, U	NaPCP	7.0	690.4	417.7	417.7	Chapman et al. 1982a
Tubificid worm, <u>Spirosperma ferox</u>	R, U	NaPCP	7.0	395.8	239.5	239.5	Chapman et al. 1982a
Tubificid worm, <u>Spirosperma nikolskyi</u>	R, U	NaPCP	7.0	902.1	545.8	545.8	Chapman et al. 1982a
Tubificid worm, <u>Stylodrilus heringianus</u>	R, U	NaPCP	7.0	579.9	350.8	-	Chapman et al. 1982a,b
Tubificid worm, <u>Stylodrilus heringianus</u>	R, U	NaPCP	7.0 (1°C)	1,638 [†]	991.0	-	Chapman et al. 1982b
Tubificid worm, <u>Stylodrilus heringianus</u>	R, U	NaPCP	7.0 (20°C)	681.2 [†]	412.1	-	Chapman et al. 1982b

Table i. (Continued)

Species	Method*	Chemical**	pH	LC50 or EC50 (µg/L)***	Adjusted LC50 or EC50 (µg/L)****	Species Mean Acute Value (µg/L)*****	Reference
<u>Tubificid worm, Stylodrilus heringianus</u>	R, U	NaPCP	6.0 (10°C)	690.4 [†]	1,141 ^{††}	-	Chapman et al. 1982b
<u>Tubificid worm, Stylodrilus heringianus</u>	R, U	NaPCP	8.0 (10°C)	874.5 [†]	193.7	408.2	Chapman et al. 1982b
<u>Tubificid worm, Tubifex tubifex</u>	R, U	NaPCP	7.0 (10°C)	349.8	211.6	-	Chapman et al. 1982a,b,c
<u>Tubificid worm, Tubifex tubifex</u>	R, U	NaPCP	7.0 (1°C)	607.5 [†]	367.6	-	Chapman et al. 1982b
<u>Tubificid worm, Tubifex tubifex</u>	R, U	NaPCP	7.0 (2°C)	405.0 [†]	245.0	-	Chapman et al. 1982b
<u>Tubificid worm, Tubifex tubifex</u>	R, U	NaPCP	6.0 (10°C)	340.6 [†]	563.0 ^{††}	-	Chapman et al. 1982b
<u>Tubificid worm, Tubifex tubifex</u>	R, U	NaPCP	8.0 (10°C)	598.3 [†]	132.5	224.2	Chapman et al. 1982b
<u>Tubificid worm, Varichaeta pacifica</u>	R, U	NaPCP	7.0	96.65	58.47	58.47	Chapman et al. 1982a
<u>Snail (adult), Aplexa hypnorum</u>	F, M	Dowicide EC-7 (88% PCP)	7.1- 7.8	157	63.54	-	Phipps and Holcombe 1985
<u>Snail (adult), Aplexa hypnorum</u>	F, M	Dowicide EC-7 (88% PCP)	7.1- 7.8	142	57.47	60.43	Phipps and Holcombe 1985
<u>Snail, Physa gyrina</u>	F, M	Dowicide EC-7 (95.7% PCP)	7.2	267	132.1	132.1	Hedtke et al. 1986
<u>Snail, Lymnaea stagnalis</u>	-	PCP	-	240	-	-	Adema and Vink 1981
<u>Snail, Gillia atrilis</u>	S, U	PCP (99%)	6.7	810	662.5	-	Stuart and Robertson 1985
<u>Snail, Gillia atrilis</u>	R, U	PCP (99%)	6.7	300	245.4	403.2	Stuart and Robertson 1985

<u>Species</u>	<u>Method*</u>	<u>Chemical**</u>	<u>pH</u>	<u>LC50 or EC50 (µg/L)***</u>	<u>Adjusted LC50 or EC50 (µg/L)****</u>	<u>Species Mean Acute Value (µg/L)*****</u>	<u>Reference</u>
<u>Cladoceran (<4 hr), Ceriodaphnia reticulata</u>	S, U	Dowicide EC-7 (93.7% PCP)	7.2- 7.4	153.7	68.79	-	Mount and Norberg 1984
<u>Cladoceran (<6 hr), Ceriodaphnia reticulata</u>	S, U	NaPCP	7.7	260	77.84	-	Hall et al. 1986
<u>Cladoceran (<6 hr), Ceriodaphnia reticulata</u>	S, U	NaPCP	7.7	510	152.7	-	Hall et al. 1986
<u>Cladoceran (<6 hr), Ceriodaphnia reticulata</u>	S, U	NaPCP	7.8	290	78.52	-	Hall et al. 1986
<u>Cladoceran (<6 hr), Ceriodaphnia reticulata</u>	S, U	NaPCP	7.8	410	111.0	-	Hall et al. 1986
<u>Cladoceran (<24 hr), Ceriodaphnia reticulata</u>	S, U	PCP (reagent grade)	8.0	900	199.3	-	Elnabarawy et al. 1986
<u>Cladoceran (<24 hr), Ceriodaphnia reticulata</u>	F, M	Dowicide EC-7 (93.7% PCP)	7.3	150	67.13	67.13	Hedtke et al. 1986
<u>Cladoceran (<24 hr), Daphnia magna</u>	S, U	-	7.4- 9.4	680	100.7	-	LeBlanc 1980
<u>Cladoceran (<24 hr), Daphnia magna</u>	S, U	PCP	-	250	-	-	Canton and Adema 1978
<u>Cladoceran (<24 hr), Daphnia magna</u>	S, U	PCP	-	400	-	-	Canton and Adema 1978
<u>Cladoceran (<24 hr), Daphnia magna</u>	S, U	PCP	-	800	-	-	Canton and Adema 1978
<u>Cladoceran (24 hr), Daphnia magna</u>	S, M	PCP	-	600	-	-	Adema 1978
<u>Cladoceran (larva), Daphnia magna</u>	-	PCP	-	1,050	-	-	Adema and Vink 1981
<u>Cladoceran (adult), Daphnia magna</u>	-	PCP	-	1,400	-	-	Adema and Vink 1981

Table 1. (Continued)

<u>Species</u>	<u>Method^a</u>	<u>Chemical^b</u>	<u>pH</u>	<u>LC50 or EC50 (µg/L)^{***}</u>	<u>Adjusted LC50 or EC50 (µg/L)^{****}</u>	<u>Species Mean Acute Value (µg/L)^{*****}</u>	<u>Reference</u>
<u>Cladoceran (<24 hr), Daphnia magna</u>	S, U	-	-	280	-	-	Hermens et al. 1984
<u>Cladoceran (<24 hr), Daphnia magna</u>	S, U	Dowicide EC-7 (93.7% PCP)	7.2- 7.4	134.0	59.97	-	Mount and Norberg 1984
<u>Cladoceran (<24 hr), Daphnia magna</u>	S, M	PCP (99%)	8.58	145	17.93	-	Thurston et al. 1985
<u>Cladoceran (<24 hr), Daphnia magna</u>	S, U	PCP	8.5	300	40.20	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia magna</u>	S, U	PCP	8.5	350	46.90	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia magna</u>	S, U	PCP	8.5	380	50.92	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia magna</u>	S, U	PCP	8.5	300	40.20	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia magna</u>	S, U	PCP	8.5	350	46.90	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia magna</u>	S, U	PCP	8.5	300	40.20	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia magna</u>	S, U	PCP	8.5	280	37.52	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia magna</u>	S, U	PCP	8.5	310	41.54	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia magna</u>	S, U	PCP	8.5	290	38.86	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia magna</u>	S, U	PCP	8.5	370	49.58	-	Lewis and Weber 1985

Table 1. (Continued)

<u>Species</u>	<u>Method^a</u>	<u>Chemical^b</u>	<u>pH</u>	<u>LC50 or EC50 (µg/L)^{***}</u>	<u>Adjusted LC50 or EC50 (µg/L)^{****}</u>	<u>Species Mean Acute Value (µg/L)^{*****}</u>	<u>Reference</u>
Cladoceran (<24 hr), <u>Daphnia magna</u>	S, U	PCP	8.5	350	46.90	-	Lewis and Weber 1985
Cladoceran (<24 hr), <u>Daphnia magna</u>	S, U	PCP	8.5	370	49.58	-	Lewis and Weber 1985
Cladoceran (<24 hr), <u>Daphnia magna</u>	S, U	PCP	8.5	340	45.56	-	Lewis and Weber 1985
Cladoceran (<24 hr), <u>Daphnia magna</u>	S, U	PCP	8.5	510	68.33	-	Lewis and Weber 1985
Cladoceran (<24 hr), <u>Daphnia magna</u>	S, U	PCP	8.5	840	112.6	-	Lewis and Weber 1985
Cladoceran (<24 hr), <u>Daphnia magna</u>	S, U	PCP	8.5	510	68.33	-	Lewis and Weber 1985
Cladoceran (<24 hr), <u>Daphnia magna</u>	S, U	NaPCP	7.7	450	134.7	-	Hall et al. 1986
Cladoceran (<24 hr), <u>Daphnia magna</u>	S, U	NaPCP	7.7	1,030	308.4	-	Hall et al. 1986
Cladoceran (<24 hr), <u>Daphnia magna</u>	S, U	NaPCP	7.8	960	259.9	-	Hall et al. 1986
Cladoceran (<24 hr), <u>Daphnia magna</u>	S, U	NaPCP	7.8	830	224.7	-	Hall et al. 1986
Cladoceran (<24 hr), <u>Daphnia magna</u>	S, U	PCP (reagent grade)	8.0	1,000	221.5	-	Elnabarawy et al. 1986
Cladoceran (6 d), <u>Daphnia magna</u>	S, M	PCP (99+%)	7.5	183	66.99	55.03	Brooke et al. Manuscript
Cladoceran (<24 hr), <u>Daphnia pulex</u>	S, U	PCP	-	2,000	-	-	Canton and Adema 1978

Table 1. (Continued)

<u>Species</u>	<u>Method^a</u>	<u>Chemical^{ab}</u>	<u>pH</u>	<u>LC50 or EC50 (µg/L)^{***}</u>	<u>Adjusted LC50 or EC50 (µg/L)^{****}</u>	<u>Species Mean Acute Value (µg/L)^{*****}</u>	<u>Reference</u>
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP	8.2	260	47.10	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP	8.2	490	88.76	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP	8.2	480	86.95	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP	8.2	470	85.13	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP	8.2	290	52.53	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP	8.2	170	30.79	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP	8.2	250	45.28	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP	8.2	390	70.64	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP	8.2	190	34.42	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP	8.2	330	59.78	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP	8.2	560	101.4	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP	8.2	550	99.63	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP	8.2	560	101.4	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP	8.2	440	79.70	-	Lewis and Weber 1985

Table 1. (Continued)

<u>Species</u>	<u>Method*</u>	<u>Chemical**</u>	<u>pH</u>	<u>LC50 or EC50 (µg/L)***</u>	<u>Adjusted LC50 or EC50 (µg/L)****</u>	<u>Species Mean Acute Value (µg/L)*****</u>	<u>Reference</u>
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP	8.2	590	106.9	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP	8.2	550	99.63	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP	8.2	680	123.2	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP	8.2	350	63.40	-	Lewis and Weber 1985
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	NaPCP	7.7	1,000	299.4	-	Hall et al. 1986
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	NaPCP	7.8	960	259.9	-	Hall et al. 1986
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	NaPCP	7.8	1,050	284.3	-	Hall et al. 1986
<u>Cladoceran (<24 hr), Daphnia pulex</u>	S, U	PCP (reagent grade)	8.0	1,100	243.6	90.83	Elnabarawy et al. 1986
<u>Cladoceran (<24 hr), Simocephalus vetulus</u>	S, U	Dowicide EC-7 (93.7% EC50)	7.2- 7.4	203.3	91.00	-	Mount and Norberg 1984
<u>Cladoceran (<24 hr), Simocephalus vetulus</u>	F, M	Dowicide EC-7 (93.7% PCP)	7.3	160	71.61	-	Hedtke et al. 1986
<u>Cladoceran (<24 hr), Simocephalus vetulus</u>	F, M	Dowicide EC-7 (93.7% PCP)	7.7	250	74.85	-	Hedtke et al. 1986
<u>Cladoceran (<24 hr), Simocephalus vetulus</u>	F, M	Dowicide EC-7 (93.7% PCP)	8.0	255	56.47	-	Hedtke et al. 1986
<u>Cladoceran (<24 hr), Simocephalus vetulus</u>	F, M	Dowicide EC-7 (93.7% PCP)	8.3	364	59.63	-	Hedtke et al. 1986
<u>Cladoceran (<24 hr), Simocephalus vetulus</u>	F, M	Dowicide EC-7 (93.7% PCP)	7.9- 8.4	196	35.50	57.72	Hedtke et al. 1986
<u>Amphipod (juvenile), Crangonyx pseudogracilis</u>	F, M	Dowicide EC-7 (88% PCP)	6.5	139	139.0	-	Spehar et al. 1985

Table 1. (Continued)

Species	Method ^a	Chemical ^b	pH	LC50 or EC50 ($\mu\text{g/L}$) ^{***}	Adjusted LC50 or EC50 ($\mu\text{g/L}$) ^{****}	Species Mean Acute Value ($\mu\text{g/L}$) ^{*****}	Reference
<u>Amphipod (juvenile), Crangonyx pseudogracilis</u>	F, M	Dowicide EC-7 (88% PCP)	7.5	465	170.2	-	Spehar et al. 1985
<u>Amphipod (juvenile), Crangonyx pseudogracilis</u>	F, M	Dowicide EC-7 (88% PCP)	8.0	929	205.7	-	Spehar et al. 1985
<u>Amphipod (juvenile), Crangonyx pseudogracilis</u>	F, M	Dowicide EC-7 (88% PCP)	8.5	1,344	180.1	172.1	Spehar et al. 1985
<u>Amphipod (11 mm), Gammarus pseudolimnaeus</u>	S, M	PCP (99+%)	7.5	296	108.35	-	Brooke et al. Manuscript
<u>Amphipod (0.050 g), Gammarus pseudolimnaeus</u>	F, M	PCP (99+%)	7.2	280	138.56	-	Call et al. 1983
<u>Amphipod (juvenile), Gammarus pseudolimnaeus</u>	F, M	Dowicide EC-7 (88% PCP)	6.5	92	92.00	-	Spehar et al. 1985
<u>Amphipod (juvenile), Gammarus pseudolimnaeus</u>	F, M	Dowicide EC-7 (88% PCP)	7.5	121	44.29	-	Spehar et al. 1985
<u>Amphipod (juvenile), Gammarus pseudolimnaeus</u>	F, M	Dowicide EC-7 (88% PCP)	8.0	484	107.19	-	Spehar et al. 1985
<u>Amphipod (juvenile), Gammarus pseudolimnaeus</u>	F, M	Dowicide EC-7 (88% PCP)	8.5	790	105.85	122.1	Spehar et al. 1985
<u>Amphipod, Hyalella azteca</u>	S, M	PCP (99+%)	7.5	239	87.48	87.48	Brooke et al. Manuscript
<u>Crayfish (0.4-2.0 g), Orconectes limnoides</u>	F, M	PCP (99%)	7.92	>183,000	>43,920	>43,920	Thurston et al. 1985
<u>Mosquito (3rd Instar), Aedes aegypti</u>	S, U	PCP (>98%)	-	7,200	-	-	Slooff et al. 1983a
<u>Mosquito (3rd Instar), Culex pipiens</u>	S, U	PCP (>98%)	-	34,000	-	-	Slooff et al. 1983a
<u>Midge (3rd, 4th Instar), Tanytarsus dissimilis</u>	S, M	PCP (99%)	8.51	31,300	4,152	-	Thurston et al. 1985
<u>Midge (3rd, 4th Instar), Tanytarsus dissimilis</u>	S, M	PCP (99%)	8.55	19,000	2,421	-	Thurston et al. 1985

Table 1. (Continued)

<u>Species</u>	<u>Method^a</u>	<u>Chemical^{ab}</u>	<u>pH</u>	<u>LC50 or EC50 ($\mu\text{g/L}$)^{***}</u>	<u>Adjusted LC50 or EC50 ($\mu\text{g/L}$)^{****}</u>	<u>Species Mean Acute Value ($\mu\text{g/L}$)^{*****}</u>	<u>Reference</u>
<u>Midge (3rd, 4th instar), <i>Tanytarsus dissimilis</i></u>	F, M	PCP (99%)	7.9	46,000	11,260	11,260	Call et al. 1983
<u>Sciomyzid (1st instar), <i>Sepedon fuscipennis</i></u>	S, U	NaPCP (90%)	8	28,000	6,201	-	McCoy and Joy 1977
<u>Sciomyzid (1st instar), <i>Sepedon fuscipennis</i></u>	S, U	NaPCP (90%)	7	30,000	18,150	10,610	McCoy and Joy 1977
<u>Coho salmon (1-3 g), <i>Oncorhynchus kisutch</i></u>	S, U	NaPCP	7.01	89	53.31	-	Davis and Hoos 1975
<u>Coho salmon (1-3 g), <i>Oncorhynchus kisutch</i></u>	S, U	NaPCP	7.01	34	20.36	-	Davis and Hoos 1975
<u>Coho salmon (2.7 g, 6.2 cm), <i>Oncorhynchus kisutch</i></u>	F, U	NaPCP	6.9- 7.5	60	29.69	31.82	Iwama and Greer 1980
<u>Sockeye salmon (yearling), <i>Oncorhynchus nerka</i></u>	F, U	NaPCP	6.8	58	42.90	-	Webb and Brett 1973
<u>Sockeye salmon (1-3 g), <i>Oncorhynchus nerka</i></u>	S, U	NaPCP	7.19	46	22.99	-	Davis and Hoos 1975
<u>Sockeye salmon (1-3 g), <i>Oncorhynchus nerka</i></u>	S, U	NaPCP	7.70	120	35.93	32.85	Davis and Hoos 1975
<u>Chinook salmon (1 g), <i>Oncorhynchus tshawytscha</i></u>	S, U	PCP (95%)	7.2- 7.5	65.28	27.78	-	Johnson and Finley 1980
<u>Chinook salmon (1 g), <i>Oncorhynchus tshawytscha</i></u>	S, U	NaPCP (90%)	7.2- 7.5	56.53	24.06	25.85	Johnson and Finley 1980
<u>Rainbow trout (1-3 g), <i>Salmo gairdneri</i></u>	S, U	NaPCP	6.96	85	53.54	-	Davis and Hoos 1975
<u>Rainbow trout (1-3 g), <i>Salmo gairdneri</i></u>	S, U	NaPCP	7.0	89	53.85	-	Davis and Hoos 1975
<u>Rainbow trout (1-3 g), <i>Salmo gairdneri</i></u>	S, U	NaPCP	7.0	46	27.83	-	Davis and Hoos 1975
<u>Rainbow trout (1-3 g), <i>Salmo gairdneri</i></u>	S, U	NaPCP	7.02	92	54.55	-	Davis and Hoos 1975

Table 1. (Continued)

<u>Species</u>	<u>Method^a</u>	<u>Chemical^{b,c}</u>	<u>pH</u>	<u>LC50 or EC50 ($\mu\text{g/L}$)^d</u>	<u>Adjusted LC50 or EC50 ($\mu\text{g/L}$)^e</u>	<u>Species Mean Acute Value ($\mu\text{g/L}$)^f</u>	<u>Reference</u>
<u>Rainbow trout (1-3 g), <i>Salmo gairdneri</i></u>	S, U	NaPCP	5.7	44	98.32	-	Davis and Hoos 1975
<u>Rainbow trout (1-3 g), <i>Salmo gairdneri</i></u>	S, U	NaPCP	7.0	69	41.75	-	Davis and Hoos 1975
<u>Rainbow trout (1.0 g), <i>Salmo gairdneri</i></u>	S, U	-	7.0	75	45.38	-	Bentley et al. 1975
<u>Rainbow trout (1.0 g), <i>Salmo gairdneri</i></u>	S, U	-	7.5	92	33.68	-	Bentley et al. 1975
<u>Rainbow trout (0.3-0.4 g), <i>Salmo gairdneri</i></u>	S, U	NaPCP	6.2- 6.8	83	83.00	-	Vigers and Maynard 1977
<u>Rainbow trout (1 g), <i>Salmo gairdneri</i></u>	S, U	PCP (96%)	7.2- 7.5	49.92	21.25	-	Johnson and Finley 1980
<u>Rainbow trout (1 g), <i>Salmo gairdneri</i></u>	S, U	NaPCP (90%)	7.2- 7.5	45.72	19.46	-	Johnson and Finley 1980
<u>Rainbow trout (10 wk), <i>Salmo gairdneri</i></u>	F, U	PCP (99+%)	7.4	66	26.71	-	Dominguez and Chapman 1984
<u>Rainbow trout (embryo; 0 h), <i>Salmo gairdneri</i></u>	R, U	PCP (97%)	7.2	3,000	1,485	-	Van Leeuwen et al. 1985
<u>Rainbow trout (embryo; 24 h), <i>Salmo gairdneri</i></u>	R, U	PCP (97%)	7.2	1,300	643.3	-	Van Leeuwen et al. 1985
<u>Rainbow trout (early eyed embryo; 14 d), <i>Salmo gairdneri</i></u>	R, U	PCP (97%)	7.2	3,000	1,485	-	Van Leeuwen et al. 1985
<u>Rainbow trout (late eyed embryo; 28 d), <i>Salmo gairdneri</i></u>	R, U	PCP (97%)	7.2	480	237.5	-	Van Leeuwen et al. 1985
<u>Rainbow trout (sac fry; 42 d), <i>Salmo gairdneri</i></u>	R, U	PCP (97%)	7.2	32	15.84	-	Van Leeuwen et al. 1985

Table 1. (Continued)

<u>Species</u>	<u>Method^a</u>	<u>Chemical^b</u>	<u>pH</u>	<u>LC50 or EC50 ($\mu\text{g/L}$)^{***}</u>	<u>Adjusted LC50 or EC50 ($\mu\text{g/L}$)^{****}</u>	<u>Species Mean Acute Value ($\mu\text{g/L}$)^{*****}</u>	<u>Reference</u>
Rainbow trout (early fry; 77 d), <u>Salmo gairdneri</u>	R, U	PCP (97%)	7.2	18	8.907	-	Van Leeuwen et al. 1985
Rainbow trout (0.81 g), <u>Salmo gairdneri</u>	S, M	PCP (99+%)	7.5	47.2	17.28	-	Brook et al. Manuscript
Rainbow trout (1.2-7.9 g), <u>Salmo gairdneri</u>	F, M	NaPCP	8.0- 8.3	210	38.04	-	Fogels and Sprague 1977
Rainbow trout (1-4 g), <u>Salmo gairdneri</u>	F, M	PCP	7.6- 8.2	160	39.18	-	Hodson et al. 1984; McCarty et al. 1985
Rainbow trout (0.6-8.0 g), <u>Salmo gairdneri</u>	F, M	PCP (99%)	7.85	115	29.61	35.34	Thurston et al. 1985
Brook trout (adult), <u>Salvelinus fontinalis</u>	F, M	PCP (99+%)	7.89	138	34.13	34.13	Cardwell et al. 1976
Goldfish (2.36 g), <u>Carassius auratus</u>	F, M	NaPCP	7.81	210	56.29	-	Adelman and Smith 1976
Goldfish (2.57 g), <u>Carassius auratus</u>	F, M	NaPCP	7.78	220	60.78	-	Adelman and Smith 1976
Goldfish (1.46 g), <u>Carassius auratus</u>	F, M	NaPCP	7.77	230	64.18	-	Adelman and Smith 1976
Goldfish (1.50 g), <u>Carassius auratus</u>	F, M	NaPCP	7.75	210	59.79	-	Adelman and Smith 1976
Goldfish (1.55 g), <u>Carassius auratus</u>	F, M	NaPCP	7.62	170	55.16	-	Adelman and Smith 1976
Goldfish (1.40 g), <u>Carassius auratus</u>	F, M	NaPCP	7.68	170	51.93	-	Adelman and Smith 1976
Goldfish (2.46 g), <u>Carassius auratus</u>	F, M	NaPCP	7.54	220	77.36	-	Adelman and Smith 1976
Goldfish (2.70 g), <u>Carassius auratus</u>	F, M	NaPCP	7.59	230	76.91	-	Adelman and Smith 1976
Goldfish (1.66 g), <u>Carassius auratus</u>	F, M	NaPCP	7.58	240	81.06	-	Adelman and Smith 1976

Table 1. (Continued)

Species	Method ^a	Chemical ^{ab}	pH	LC50 or EC50 ($\mu\text{g/L}$) ^{***}	Adjusted LC50 or EC50 ($\mu\text{g/L}$) ^{****}	Species Mean Acute Value ($\mu\text{g/L}$) ^{*****}	Reference
<u>Goldfish (1.74), Carassius auratus</u>	F, M	NaPCP	7.59	240	80.25	-	Adelman and Smith 1976
<u>Goldfish (1.69 g), Carassius auratus</u>	F, M	NaPCP	7.58	200	67.55	-	Adelman and Smith 1976
<u>Goldfish (1.65 g), Carassius auratus</u>	F, M	NaPCP	7.60	190	62.90	-	Adelman and Smith 1976
<u>Goldfish (2.31 g), Carassius auratus</u>	F, M	NaPCP	7.83	290	76.19	-	Adelman and Smith 1976
<u>Goldfish (2.31 g), Carassius auratus</u>	F, M	NaPCP	7.84	300	78.03	-	Adelman and Smith 1976
<u>Goldfish (1.76 g), Carassius auratus</u>	F, M	NaPCP	7.73	200	58.10	-	Adelman and Smith 1976
<u>Goldfish (1.54 g), Carassius auratus</u>	F, M	NaPCP	7.76	250	70.47	-	Adelman and Smith 1976
<u>Goldfish (1-2 g), Carassius auratus</u>	F, M	NaPCP	-	247 ^{†††}	-	-	Adelman and Smith 1976; Adelman et al. 1976a
<u>Goldfish (2-4 g), Carassius auratus</u>	F, M	NaPCP	-	190 ^{†††}	-	-	Adelman and Smith 1976; Adelman et al. 1976a
<u>Goldfish (1.9 g), Carassius auratus</u>	F, M	Dowicide EC-7 (88% PCP)	7.1- 7.8	117	47.36	-	Phipps and Holcombe 1985
<u>Goldfish (1.2 g), Carassius auratus</u>	F, M	Dowicide EC-7 (88% PCP)	7.1- 7.8	156	63.14	-	Phipps and Holcombe 1985
<u>Goldfish (1.7 g), Carassius auratus</u>	F, M	Dowicide EC-7 (88% PCP)	7.1- 7.8	191	77.31	-	Phipps and Holcombe 1985
<u>Goldfish (1.0-4.0 g), Carassius auratus</u>	F, M	PCP ^c (99%)	7.94	328	77.15	-	Thurston et al. 1985
<u>Goldfish (1.0-4.0 g), Carassius auratus</u>	F, M	PCP (99%)	7.84	200	52.02	65.53	Thurston et al. 1985
<u>Common carp (8 mm), Cyprinus carpio</u>	R, U	NaPCP	7.2	8.8	4.355	4.355	Verma et al. 1981c

Table 1. (Continued)

<u>Species</u>	<u>Method[#]</u>	<u>Chemical^{**}</u>	<u>pH</u>	<u>LC50 or EC50 ($\mu\text{g/L}$)^{***}</u>	<u>Adjusted LC50 or EC50 ($\mu\text{g/L}$)^{****}</u>	<u>Species Mean Acute Value ($\mu\text{g/L}$)^{*****}</u>	<u>Reference</u>
<u>Fathead minnow (4-8 wk), Pimephales promelas</u>	S, U	PCP	-	600	-	-	Mattson et al. 1976
<u>Fathead minnow (1.1 g), Pimephales promelas</u>	S, U	PCP (96%)	7.2- 7.5	196.8	83.76	-	Johnson and Finley 1980
<u>Fathead minnow (40 d), Pimephales promelas</u>	F, U	PCP (99%)	7.4	470	190.2	-	Cleveland et al. 1982
<u>Fathead minnow (14-30 d), Pimephales promelas</u>	S, U	NaPCP	7.7	18	5.389	-	Hall et al. 1986
<u>Fathead minnow (adult), Pimephales promelas</u>	F, M	-	7.92- 8.20	194 (15°C)	40.45	-	Ruesink and Smith 1975
<u>Fathead minnow (adult), Pimephales promelas</u>	F, M	-	7.81- 8.20	314 (25°C)	69.54	-	Ruesink and Smith 1975
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	7.82	200	53.08	-	Adelman and Smith 1976
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	7.83	180	47.29	-	Adelman and Smith 1976
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	7.72	220	64.56	-	Adelman and Smith 1976
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	7.72	180	52.82	-	Adelman and Smith 1976
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	7.69	190	57.46	-	Adelman and Smith 1976
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	7.68	210	64.15	-	Adelman and Smith 1976
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	7.86	220	56.08	-	Adelman and Smith 1976
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	7.78	180	49.73	-	Adelman and Smith 1976
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	7.59	190	63.53	-	Adelman and Smith 1976

Table 1. (Continued)

Species	Method ^a	Chemical ^b	pH	LC50 or EC50 (µg/L) ^{***}	Adjusted LC50 or EC50 (µg/L) ^{****}	Species Mean Acute Value (µg/L) ^{*****}	Reference
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	7.62	190	61.65	-	Adelman and Smith 1976
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	7.65	240	75.56	-	Adelman and Smith 1976
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	7.65	200	62.96	-	Adelman and Smith 1976
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	7.63	200	64.24	-	Adelman and Smith 1976
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	7.58	190	64.18	-	Adelman and Smith 1976
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	7.83	270	70.94	-	Adelman and Smith 1976
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	7.82	230	61.04	-	Adelman and Smith 1976
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	-	227 ^{†††}	-	-	Adelman and Smith 1976; Adelman et al. 1976a
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	-	203 ^{†††}	-	-	Adelman and Smith 1976; Adelman et al. 1976a
<u>Fathead minnow (4 wk), Pimephales promelas</u>	F, M	NaPCP	-	198 ^{††††}	-	-	Adelman and Smith 1976; Adelman et al. 1976a
<u>Fathead minnow (7 wk), Pimephales promelas</u>	F, M	NaPCP	-	230 ^{††††}	-	-	Adelman and Smith 1976; Adelman et al. 1976a
<u>Fathead minnow (11 wk), Pimephales promelas</u>	F, M	NaPCP	-	222 ^{††††}	-	-	Adelman and Smith 1976; Adelman et al. 1976a
<u>Fathead minnow (14 wk), Pimephales promelas</u>	F, M	NaPCP	-	190 ^{††††}	-	-	Adelman and Smith 1976; Adelman et al. 1976a
<u>Fathead minnow (3 mo), Pimephales promelas</u>	F, M	PCP (99+%)	7.83	285	74.88	-	Cardwell et al. 1976
<u>Fathead minnow (30-35 d), Pimephales promelas</u>	F, M	-	7.41- 8.33	220	55.52	-	Phipps et al. 1981

Table 1. (Continued)

Species	Method ^a	Chemical ^{b,c}	pH	LC50 or EC50 ($\mu\text{g/L}$) ^d	Adjusted LC50 or EC50 ($\mu\text{g/L}$) ^e	Species Mean Acute Value ($\mu\text{g/L}$) ^f	Reference
<u>Fathead minnow (30-35 d), Pimephales promelas</u>	F, M	-	7.41- 8.33	230	58.05	-	Phipps et al. 1981
<u>Fathead minnow (30-35 d), Pimephales promelas</u>	F, M	-	7.41- 8.33	232	58.55	-	Hall et al. 1984
<u>Fathead minnow (juvenile), Pimephales promelas</u>	F, M	Dowicide EC-7 (88% PCP)	6.5	95	95.00	-	Spehar et al. 1985
<u>Fathead minnow (juvenile), Pimephales promelas</u>	F, M	Dowicide EC-7 (88% PCP)	7.5	218	79.80	-	Spehar et al. 1985
<u>Fathead minnow (juvenile), Pimephales promelas</u>	F, M	Dowicide EC-7 (88% PCP)	8.0	261	57.80	-	Spehar et al. 1985
<u>Fathead minnow (juvenile), Pimephales promelas</u>	F, M	Dowicide EC-7 (88% PCP)	8.5	378	50.65	-	Spehar et al. 1985
<u>Fathead minnow (0.5 g), Pimephales promelas</u>	F, M	Dowicide EC-7 (88% PCP)	7.1- 7.8	223	90.26	-	Phipps and Holcombe 1985
<u>Fathead minnow (0.4 g), Pimephales promelas</u>	F, M	Dowicide EC-7 (88% PCP)	7.1- 7.8	286	115.76	-	Phipps and Holcombe 1985
<u>Fathead minnow (0.4 g), Pimephales promelas</u>	F, M	Dowicide EC-7 (88% PCP)	7.1- 7.8	244	98.76	-	Phipps and Holcombe 1985
<u>Fathead minnow (0.2-1.0 g), Pimephales promelas</u>	F, M	PCP (99%)	8.01	266	58.32	63.11	Thurston et al. 1985
<u>Channel catfish (0.8 g), Ictalurus punctatus</u>	S, U	PCP (96%)	7.2- 7.5	65.28	27.78	-	Johnson and Finley 1980
<u>Channel catfish (0.8 g), Ictalurus punctatus</u>	S, U	NaPCP (90%)	7.2- 7.5	64.01	27.24	-	Johnson and Finley 1980
<u>Channel catfish (1.4 g), Ictalurus punctatus</u>	F, M	Dowicide EC-7 (88% PCP)	7.1- 7.8	<53	<21.45 ^{††}	-	Phipps and Holcombe 1985
<u>Channel catfish (1.4 g), Ictalurus punctatus</u>	F, M	Dowicide EC-7 (88% PCP)	7.1- 7.8	54	21.86	-	Phipps and Holcombe 1985
<u>Channel catfish (1.9 g), Ictalurus punctatus</u>	F, M	Dowicide EC-7 (88% PCP)	7.1- 7.8	54	21.86	-	Phipps and Holcombe 1985

Table 1. (Continued)

<u>Species</u>	<u>Method^a</u>	<u>Chemical^{ab}</u>	<u>pH</u>	<u>LC50 or EC50 ($\mu\text{g/L}$)^{abc}</u>	<u>Adjusted LC50 or EC50 ($\mu\text{g/L}$)^{abcd}</u>	<u>Species Mean Acute Value ($\mu\text{g/L}$)^{abcde}</u>	<u>Reference</u>
<u>Channel catfish (0.3-4.0 g), <i>Ictalurus punctatus</i></u>	F, M	PCP (99%)	7.71	132	39.12	26.54	Thurston et al. 1985
<u>Flagfish (0.1-0.3 g), <i>Jordanella floridae</i></u>	F, M	NaPCP (79%)	8.0- 8.3	1,610	291.6	291.6	Fogels and Sprague 1977
<u>Mosquitofish (0.1-1.0 g), <i>Gambusia affinis</i></u>	F, M	PCP (99%)	8.05	288	60.62	-	Thurston et al. 1985
<u>Mosquitofish (0.1-1.0 g), <i>Gambusia affinis</i></u>	F, M	PCP (99%)	8.02	278	60.34	60.50	Thurston et al. 1985
<u>Guppy (young), <i>Poecilia reticulata</i></u>	-	PCP	-	720	-	-	Adema and Vink 1981
<u>Guppy (young), <i>Poecilia reticulata</i></u>	-	PCP	-	880	-	-	Adema and Vink 1981
<u>Guppy (adult), <i>Poecilia reticulata</i></u>	-	PCP	-	450	-	-	Adema and Vink 1981
<u>Guppy, <i>Poecilia reticulata</i></u>	R, M	-	7.0	400	242.0	-	Salkinoja-Salonen et al. 1981
<u>Guppy (0.04-0.06 g), <i>Poecilia reticulata</i></u>	R, U	PCP	5	42.6	192.4	-	Saarikoski and Viluksela 1981,1982
<u>Guppy (0.04-0.06 g), <i>Poecilia reticulata</i></u>	R, U	PCP	6	117	193.4	-	Saarikoski and Viluksela 1981,1982
<u>Guppy (0.04-0.06 g), <i>Poecilia reticulata</i></u>	R, U	PCP	7	442	267.4	-	Saarikoski and Viluksela 1981,1982
<u>Guppy (0.04-0.06 g), <i>Poecilia reticulata</i></u>	R, U	PCP	8	911	201.8	-	Saarikoski and Viluksela 1981, 1982
<u>Guppy (0.0875 g), <i>Poecilia reticulata</i></u>	R, U	PCP	8.1	970	194.3	-	Gupta et al. 1982a
<u>Guppy (0.0875 g), <i>Poecilia reticulata</i></u>	R, U	NaPCP	8.1	711	142.4	-	Gupta et al. 1982a
<u>Guppy, <i>Poecilia reticulata</i></u>	R, U	PCP	7.7	204	61.08	-	Khargarot 1983

Table 1. (Continued)

<u>Species</u>	<u>Method^a</u>	<u>Chemical^b</u>	<u>pH</u>	<u>LC50 or EC50 (μg/L)^{***}</u>	<u>Adjusted LC50 or EC50 (μg/L)^{****}</u>	<u>Species Mean Acute Value (μg/L)^{*****}</u>	<u>Reference</u>
<u>Guppy (juvenile), Poecilia reticulata</u>	R, U	PCP	7.2	1,020	504.7	195.4	Brown et al. 1985
<u>Bluegill (1.1 g), Lepomis macrochirus</u>	S, U	-	7.0	60	36.30	-	Bentley et al. 1975
<u>Bluegill (1.1 g), Lepomis macrochirus</u>	S, U	-	7.5	77	28.19	-	Bentley et al. 1975
<u>Bluegill (6 mo), Lepomis macrochirus</u>	R, U	PCP	7.2- 7.7	260	105.23	-	Pruitt et al. 1977; Pierce 1978
<u>Bluegill (6 mo), Lepomis macrochirus</u>	R, U	PCP	7.2- 7.7	305	123.5	-	Pruitt et al. 1977; Pierce 1978
<u>Bluegill (0.4 g), Lepomis macrochirus</u>	S, U	PCP (96%)	7.2- 7.5	30.72	13.07	-	Johnson and Finley 1980
<u>Bluegill (0.4 g), Lepomis macrochirus</u>	S, U	NaPCP (90%)	7.2- 7.5	36.58	15.57	-	Johnson and Finley 1980
<u>Bluegill (0.2 g), Lepomis macrochirus</u>	F, M	Dowicide EC-7 (88% PCP)	7.1- 7.8	150	60.71	-	Phipps and Holcombe 1985
<u>Bluegill (0.3 g), Lepomis macrochirus</u>	F, M	Dowicide EC-7 (88% PCP)	7.1- 7.8	152	61.52	-	Phipps and Holcombe 1985
<u>Bluegill (0.4 g), Lepomis macrochirus</u>	F, M	Dowicide EC-7 (88% PCP)	7.1- 7.8	115	46.55	-	Phipps and Holcombe 1985
<u>Bluegill (0.3-2.0 g), Lepomis macrochirus</u>	F, M	PCP (99%)	8.03	202	43.41	56.41	Thurston et al. 1985
<u>Largemouth bass (14 d), Micropterus salmoides</u>	R, U	PCP (99%)	7.2	287	142.0	-	Johansen et al. 1985
<u>Largemouth bass (28 d), Micropterus salmoides</u>	R, U	PCP (99%)	7.2	275	136.1	-	Johansen et al. 1985
<u>Largemouth bass (49 d), Micropterus salmoides</u>	R, U	PCP (99%)	7.2	136	67.30	-	Johansen et al. 1985
<u>Largemouth bass (84 d), Micropterus salmoides</u>	R, U	PCP (99%)	7.2	189	93.53	105.0	Johansen et al. 1985
<u>Bullfrog (tadpole), Rana catesbeiana</u>	F, M	PCP (99%)	8.3	207	44.48	44.48	Thurston et al. 1985

Table 1. (Continued)

<u>Species</u>	<u>Method</u> ^a	<u>Chemical</u> ^b	<u>Salinity</u> (g/kg)	<u>LC50</u> or <u>EC50</u> (μ g/L) ^c	<u>Species Mean</u> <u>Acute Value</u> (μ g/L)	<u>Reference</u>
<u>SALTWATER SPECIES</u>						
<u>Polychaete worm (adult),</u> <u>Nereis arenaceodentata</u>	S, U	PCP (98%)	28	435	435	Rubenstein 1981
<u>Polychaete worm</u> (3-d larva), <u>Ophryotrocha diadema</u>	R, M	PCP	33	620	-	Adema and Vink 1981; Hooftman and Vink 1980
<u>Polychaete worm (adult),</u> <u>Ophryotrocha diadema</u>	R, M	PCP	33	1,200	862.6	Adema and Vink 1981; Hooftman and Vink 1980
<u>Oligochaete worm (adult),</u> <u>Limnodriloides verrucosus</u>	R, U	NaPCP	20 (10°C, pH 6)	64.43 ^{†,††}	-	Chapman et al. 1982b
<u>Oligochaete worm (adult),</u> <u>Limnodriloides verrucosus</u>	R, U	NaPCP	20 (10°C, pH 7)	230.1	397.2	Chapman et al. 1982a,b
<u>Oligochaete worm (adult),</u> <u>Limnodriloides verrucosus</u>	R, U	NaPCP	20 (10°C, pH 8)	441.8 [†]	-	Chapman et al. 1982b
<u>Oligochaete worm (adult),</u> <u>Limnodriloides verrucosus</u>	R, U	NaPCP	20 (1°C, pH 7)	644.4 [†]	-	Chapman et al. 1982b
<u>Oligochaete worm,</u> <u>Monophelephorus cuticulatus</u>	R, U	NaPCP	20 (1°C, pH 7)	1,473 [†]	-	Chapman et al. 1982b
<u>Oligochaete worm,</u> <u>Monophelephorus cuticulatus</u>	R, U	NaPCP	20 (10°C, pH 7)	487.9	-	Chapman et al. 1982a,b
<u>Oligochaete worm,</u> <u>Monophelephorus cuticulatus</u>	R, U	NaPCP	20 (20°C, pH 7)	492.5 [†]	-	Chapman et al. 1982b
<u>Oligochaete worm,</u> <u>Monophelephorus cuticulatus</u>	R, U	NaPCP	20 (10°C, pH 6)	423.4 ^{†,††}	-	Chapman et al. 1982b
<u>Oligochaete worm,</u> <u>Monophelephorus cuticulatus</u>	R, U	NaPCP	20 (10°C, pH 8)	635.6 [†]	-	Chapman et al. 1982b
<u>Oligochaete worm,</u> <u>Monophelephorus cuticulatus</u>	R, U	NaPCP	10 (10°C, pH 7)	340.6 [†]	598.2	Chapman et al. 1982b

Table 1. (Continued)

<u>Species</u>	<u>Method^a</u>	<u>Chemical^b</u>	<u>Salinity (g/kg)</u>	<u>LC50 or EC50 (µg/L)^c</u>	<u>Species Mean Acute Value (µg/L)</u>	<u>Reference</u>
<u>Oligochaete worm, Tubificoides gabriellae</u>	R, U	NaPCP	20 (10°C, pH 7)	423.4	423.4	Chapman et al. 1982a
<u>Common Atlantic slippershell (larva), Crepidula fornicata</u>	Averaged results from S, R, F; M	PCP	-	1,200	1,200	Adema and Vink 1981
<u>Blue mussel (adult), Mytilus edulis</u>	Averaged results from S, R, F; M	PCP	-	18,000 ^{††}	-	Adema and Vink 1981
<u>Blue mussel (embryo), Mytilus edulis</u>	S, U	NaPCP	124	328.8 ^{†††††}	328.8	Woelke 1972
<u>Pacific oyster (embryo), Crassostrea gigas</u>	S, U	NaPCP	-	40.85 ^{†††††}	40.85	Woelke 1972
<u>Eastern oyster (embryo), Crassostrea virginica</u>	S, U	NaPCP	17	36.95	-	Borthwick and Schimmel 1978
<u>Eastern oyster (embryo), Crassostrea virginica</u>	S, U	PCP (98%)	30	430	-	Zarogian 1981
<u>Eastern oyster (embryo), Crassostrea virginica</u>	S, U	PCP (98%)	30	49	-	Zarogian 1981
<u>Eastern oyster (embryo), Crassostrea virginica</u>	S, U	PCP (98%)	30	>180	-	Zarogian 1981
<u>Eastern oyster (embryo), Crassostrea virginica</u>	S, U	PCP (98%)	30	640	†††††	Zarogian 1981
<u>Copepod (adult), Pseudodiaptomus coronatus</u>	S, U	NaPCP	18	62.81	62.81	Hauch et al. 1980
<u>Copepod (adult), Temora longicornis</u>	Averaged results from S, R, F; M	PCP	-	170	170	Adema and Vink 1981
<u>Brown shrimp (adult), Penaeus aztecus</u>	F, M	NaPCP	26.5	>195	>195	Schimmel et al. 1978

Table 1. (Continued)

<u>Species</u>	<u>Method^a</u>	<u>Chemical^b</u>	<u>Salinity (g/kg)</u>	<u>LC50 or EC50 (μg/L)^{c,d}</u>	<u>Species Mean Acute Value (μg/L)</u>	<u>Reference</u>
<u>Pink shrimp (adult), <i>Penaeus duorarum</i></u>	S, U	PCP	25	5,600	5,600	Bentley et al. 1975
<u>Grass shrimp (adult intermolt), <i>Palaemonetes pugio</i></u>	R, U	NaPCP	10	2,431 ^{††}	-	Conklin and Rao 1978a
<u>Grass shrimp (adult early premolt), <i>Palaemonetes pugio</i></u>	R, U	NaPCP	10	2,534 ^{††}	-	Conklin and Rao 1978a
<u>Grass shrimp (adult late premolt), <i>Palaemonetes pugio</i></u>	R, U	NaPCP	10	402.7	-	Conklin and Rao 1978a,b
<u>Grass shrimp (24-hr larva), <i>Palaemonetes pugio</i></u>	S, U	NaPCP	24	599.5	-	Borthwick and Schimmel 1978
<u>Grass shrimp (juvenile), <i>Palaemonetes pugio</i></u>	F, M	NaPCP	24.3	>515.0 ^{††}	491.3	Schimmel et al. 1978
<u>Pacific herring (day 1 yolk sac larva), <i>Clupea harengus pallasii</i></u>	S, U	NaPCP	24-27	147.8 ^{†,††}	-	Vigers et al. 1978
<u>Pacific herring (day 12 larva), <i>Clupea harengus pallasii</i></u>	S, U	NaPCP	24-27	63.74 ^{†,††}	-	Vigers et al. 1978
<u>Pacific herring (day 52 larva), <i>Clupea harengus pallasii</i></u>	S, U	NaPCP	24-27	23.09 [†]	-	Vigers et al. 1978
<u>Pacific herring (day 180 juvenile), <i>Clupea harengus pallasii</i></u>	S, U	NaPCP	24-27	27.71 [†]	25.29	Vigers et al. 1978
<u>Sheepshead minnow (1-day juvenile), <i>Cyprinodon variegatus</i></u>	S, U	PCP	10	329	-	Borthwick and Schimmel 1978

Table 1. (Continued)

<u>Species</u>	<u>Method*</u>	<u>Chemical**</u>	<u>Salinity (g/kg)</u>	<u>LC50 or EC50 (µg/L)***</u>	<u>Species Mean Acute Value (µg/L)</u>	<u>Reference</u>
Sheepshead minnow (2-week juvenile), <u>Cyprinodon variegatus</u>	S, U	PCP	10	392	-	Borthwick and Schimmel 1978
Sheepshead minnow (4-week juvenile), <u>Cyprinodon variegatus</u>	S, U	PCP	10	240	-	Borthwick and Schimmel 1978
Sheepshead minnow (6-week juvenile), <u>Cyprinodon variegatus</u>	S, U	PCP	10	223	-	Borthwick and Schimmel 1978
Sheepshead minnow (2-week juvenile) <u>Cyprinodon variegatus</u>	S, U	Dowicide (79% NaPCP)	24	376.5	-	Borthwick and Schimmel 1978
Sheepshead minnow (juvenile), <u>Cyprinodon variegatus</u>	F, M	PCP	24	442	442	Parrish et al. 1978
Longnose killifish (juvenile), <u>Fundulus similis</u>	F, M	NaPCP	22.9	>306	>306	Schimmel et al. 1978
Pinfish (larva), <u>Lagodon rhomboides</u>	S, U	NaPCP	26	35.10	-	Borthwick and Schimmel 1978
Pinfish (larva), <u>Lagodon rhomboides</u>	S, U	Dowicide (79% NaPCP)	26	48.16	-	Borthwick and Schimmel 1978
Pinfish (adult), <u>Lagodon rhomboides</u>	F, M	NaPCP	20.8	53.2	53.2	Schimmel et al. 1978
Striped mullet (juvenile), <u>Muqil cephalus</u>	F, M	NaPCP	25.5	112.1	112.1	Schimmel et al. 1978

Table 1. (Continued)

- * S = static; R = renewal; F = flow-through; U = unmeasured, M = measured.
- ** PCP = pentachlorophenol; NaPCP = sodium pentachlorophenate. Percent purity is given in parentheses when available.
- *** Results are expressed as pentachlorophenol. If the concentrations were not measured and the published results were not reported to be adjusted for purity, the published results were multiplied by the purity if it was reported to be less than 97%.
- **** Freshwater LC50s and EC50s were adjusted to pH = 6.5 using the pooled slope of 1.005 (see text).
- ***** Freshwater Species Mean Acute Values are at pH = 6.5.
- † Interpolated from graph.
- †† Not used in calculation of Species Mean Acute Value.
- ††† Mean of three LC50s.
- †††† Mean of four LC50s.
- ††††† Calculated using moving average method and author's raw data.
- †††††† No Species Mean Acute Value calculated because acute values are too divergent for this species.

Results of Covariance Analysis of Freshwater Acute Toxicity versus pH

<u>Species</u>	<u>n</u>	<u>Slope</u>	<u>95% Confidence Limits</u>		<u>Degrees of Freedom</u>
<u>Simocephalus vetulus</u>	4	0.760	0.128,	1.393	2
<u>Crangonyx pseudo-gracilis</u>	4	1.164	0.844,	1.484	2
<u>Gammarus pseudolimnoides</u>	4	1.129	-0.319,	2.578	2
Fathead minnow	4	0.678	0.413,	0.943	2
Guppy	4	1.052	0.710,	1.394	2
All of above	20	1.005*	0.841,	1.169	14

* P = 0.28 for equality of slopes.

Table 2. Chronic Toxicity of Pentachlorophenol to Aquatic Animals

<u>Species</u>	<u>Test*</u>	<u>Chemical**</u>	<u>pH</u>	<u>Limits (µg/L)***</u>	<u>Chronic Value (µg/L)</u>	<u>Adjusted Chronic Value (µg/L)****</u>	<u>Reference</u>
FRESHWATER SPECIES							
<u>Snail, Physa gyrina</u>	LC	Dowicide EC-7 (93.7% PCP)	7.4- 7.7	<26 [†]	<26	<8.607	Hedtke et al. 1986
<u>Cladoceran, Ceriodaphnia reticulata</u>	LC	Dowicide EC-7 (93.7% PCP)	7.3	<22 [†]	<22	<9.846	Hedtke et al. 1986
<u>Cladoceran, Ceriodaphnia reticulata</u>	LC	Dowicide EC-7 (93.7% PCP)	7.3	<4.1 [†]	<4.1	<1.835	Hedtke et al. 1986
<u>Cladoceran, Daphnia magna</u>	LC	PCP	-	180- 320	240	-	Adema 1978
<u>Cladoceran, Simocephalus vetulus</u>	LC	Dowicide EC-7 (93.7% PCP)	7.3	119- 264	177.2	79.66	Hedtke et al. 1986
<u>Cladoceran, Simocephalus vetulus</u>	LC	Dowicide EC-7 (93.7% PCP)	7.9- 8.4 ^{††}	137- 357	221.2	40.03	Hedtke et al. 1986
<u>Rainbow trout, Salmo gairdneri</u>	ELS	PCP (99+%)	7.4	11- 19	14.46	5.666	Dominguez and Chapman 1984
<u>Fathead minnow, Pimephales promelas</u>	ELS	PCP (Reagent grade)	7.2- 7.9	44.9- 73.0	57.25	18.94	Holcombe et al. 1982
<u>Fathead minnow, Pimephales promelas</u>	ELS	Dowicide EC-7 (88% PCP)	6.5	16.5- 34.6	23.89	23.89	Spehar et al. 1985

Table 2. (Continued)

<u>Species</u>	<u>Test*</u>	<u>Chemical**</u>	<u>pH</u>	<u>Limits ($\mu\text{g/L}$)***</u>	<u>Chronic Value ($\mu\text{g/L}$)</u>	<u>Adjusted Chronic Value ($\mu\text{g/L}$)****</u>	<u>Reference</u>
<u>Fathead minnow, Pimephales promelas</u>	ELS	Dowicide EC-7 (88% PCP)	7.5	27.6- 58.2	40.08	14.64	Spehar et al. 1985
<u>Fathead minnow, Pimephales promelas</u>	ELS	Dowicide EC-7 (88% PCP)	8.0	32.0- 75.0	48.99	10.85	Spehar et al. 1985
<u>Fathead minnow, Pimephales promelas</u>	ELS	Dowicide EC-7 (88% PCP)	8.5	63.7- 125.0	89.23	11.92	Spehar et al. 1985
<u>SALTWATER SPECIES</u>							
<u>Sheepshead minnow, Cyprinodon variegatus</u>	LC	PCP	-	47-88	64.31	-	Parrish et al. 1978

* LC = life-cycle or partial life-cycle; ELS = early life-stage.

** PCP = pentachlorophenol. Percent purity is given in parentheses when available.

*** Results are based on measured concentrations of pentachlorophenol.

**** Freshwater chronic values were adjusted to pH = 6.5 using the pooled slope of 1.005 (see text).

† unacceptable effects occurred at all concentrations tested.

†† Controlled daily pH fluctuation consisting of 16 hours at 8.4 and 8 hours at 7.9.

Table 2. (Continued)

<u>Species</u>	<u>Acute-Chronic Ratio</u>			
	<u>pH</u>	<u>Acute Value</u> <u>($\mu\text{g/L}$)</u>	<u>Chronic Value</u> <u>($\mu\text{g/L}$)</u>	<u>Ratio</u>
Snail, <u>Physa gyrina</u>	7.2-7.7	267	<26	>10.27
Cladoceran, <u>Ceriodaphnia reticulata</u>	7.3	150	<22	>6.818
Cladoceran, <u>Ceriodaphnia reticulata</u>	7.3	150	<4.1	>36.59
Cladoceran, <u>Daphnia magna</u>	-	600	240	2.500
Cladoceran, <u>Simocephalus vetulus</u>	7.3	160	177.2	0.9029
Cladoceran, <u>Simocephalus vetulus</u>	7.9-8.4	196	221.2	0.8861
Rainbow trout, <u>Salmo gairdneri</u>	7.4	66	14.46	4.564
Fathead minnow, <u>Pimephales promelas</u>	7.1-8.3	224.9*	57.25	3.928
Fathead minnow, <u>Pimephales promelas</u>	6.5	95	23.89	3.977
Fathead minnow, <u>Pimephales promelas</u>	7.5	218	40.08	5.439
Fathead minnow, <u>Pimephales promelas</u>	8.0	261	48.99	5.328
Fathead minnow, <u>Pimephales promelas</u>	8.5	378	89.23	4.236
Sheepshead minnow, <u>Cyprinodon variegatus</u>	-	442	64.31	6.873

* Geometric mean of 2 acute values obtained by Philpps et al. (1981) (see Table 1).

Table 3. Ranked Genus Mean Acute Values with Species Mean Acute-Chronic Ratios

<u>Rank#</u>	<u>Genus Mean Acute Value (µg/L)</u>	<u>Species</u>	<u>Species Mean Acute Value (µg/L)**</u>	<u>Species Mean Acute-Chronic Ratio***</u>
<u>FRESHWATER SPECIES</u>				
33	>43,920	Crayfish, <u>Orconectes immunis</u>	>43,920	-
32	11,260	Midge, <u>Tanytarsus dissimilis</u>	11,260	-
31	10,610	Sclomyzid, <u>Sepedon fuscipennis</u>	10,610	-
30	417.7	Tubificid worm, <u>Rhyacodrilus montana</u>	417.7	-
29	408.2	Tubificid worm, <u>Stylodrilus heringianus</u>	408.2	-
28	403.2	Snail, <u>Gililla altilis</u>	403.2	-
27	361.6	Tubificid worm, <u>Spilosperma ferox</u>	239.5	-
		Tubificid worm, <u>Spilosperma nikolskyl</u>	545.8	-
26	317.5	Tubificid worm, <u>Quilstradrilus multisetosus</u>	317.5	-
25	291.6	Flagfish, <u>Jordanella floridae</u>	291.6	-
24	224.2	Tubificid worm, <u>Tubifex tubifex</u>	224.2	-
23	182.5	Tubificid worm, <u>Limnodrilus hoffmeisteri</u>	182.5	-
22	172.1	Amphipod, <u>Crangonyx pseudoqracilis</u>	172.1	-
21	195.4	Guppy, <u>Lebistes reticulata</u>	195.4	-

Table 3. (Continued)

Rank*	Genus Mean Acute Value ($\mu\text{g/L}$)	Species	Species Mean Acute Value ($\mu\text{g/L}$)**	Species Mean Acute-Chronic Ratio***
20	155.9	Tubificid worm, <u>Branchiura sowerbyi</u>	155.9	-
19	132.1	Snail, <u>Physa gyrina</u>	132.1	>10.27
18	121.1	Amphipod, <u>Gammarus pseudolimnaeus</u>	122.1	-
17	105.0	Largemouth bass, <u>Micropterus salmoides</u>	105.0	-
16	87.48	Amphipod, <u>Hyalella azteca</u>	87.48	-
15	78.10	Cladoceran, <u>Daphnia pulex</u>	90.83	-
		Cladoceran, <u>Daphnia magna</u>	67.15	2.500
14	67.13	Cladoceran, <u>Ceriodaphnia reticulata</u>	67.13	>15.79 [†]
13	65.53	Goldfish, <u>Carassius auratus</u>	65.53	-
12	63.11	Fathead minnow, <u>Pimephales promelas</u>	63.11	4.535 ^{††}
11	60.50	Mosquitofish, <u>Gambusia affinis</u>	60.50	-
10	60.43	Snail, <u>Aplexa hypnorum</u>	60.43	-
9	58.47	Tubificid worm, <u>Varichaeta pacifica</u>	58.47	-
8	57.72	Cladoceran, <u>Simocephalus vetulus</u>	57.72	0.8945 [†]

Table 3. (Continued)

Rank*	Genus Mean Acute Value (µg/L)	Species	Species Mean Acute Value (µg/L)**	Species Mean Acute-Chronic Ratio***
7	56.41	Bluegill, <u>Lepomis macrochirus</u>	56.41	-
6	44.48	Bullfrog, <u>Rana catesblana</u>	44.48	-
5	35.34	Rainbow trout, <u>Salmo gairdneri</u>	35.34	4.564
4	34.13	Brook trout, <u>Salvelinus fontinalis</u>	34.13	-
3	30.01	Coho salmon, <u>Oncorhynchus kisutch</u>	31.82	-
		Sockeye salmon, <u>Oncorhynchus nerka</u>	32.85	-
		Chinook salmon, <u>Oncorhynchus tshawytscha</u>	25.85	-
2	26.54	Channel catfish, <u>Ictalurus punctatus</u>	26.54	-
1	4.355	Common carp, <u>Cyprinus carpio</u>	4.355	-
<u>SALTWATER SPECIES</u>				
17	1,200	Common Atlantic slippershell, <u>Crepidula fornicata</u>	1,200	-
16	>1,045	Brown shrimp, <u>Penaeus aztecus</u>	>195	-
		Pink shrimp, <u>Penaeus duorarum</u>	5,600	-
15	862.6	Polychaete worm, <u>Ophryotrocha diadema</u>	862.6	-

Table 3. (Continued)

<u>Rank^a</u>	<u>Genus Mean Acute Value (µg/L)</u>	<u>Species</u>	<u>Species Mean Acute Value (µg/L)^{aa}</u>	<u>Species Mean Acute-Chronic Ratio^{aaa}</u>
14	598.2	Oligochaete worm, <u>Monopylephorus cuticulatus</u>	598.2	-
13	491.3	Grass shrimp, <u>Palaemonetes pugio</u>	491.3	-
12	442	Sheepshead minnow, <u>Cyprinodon variegatus</u>	442	6.873
11	435	Polychaete worm, <u>Nereis arenaceodentata</u>	435	-
10	423.4	Oligochaete worm, <u>Tubificoides gabriellae</u>	423.4	-
9	397.2	Oligochaete worm, <u>Limnodriloides verrucosus</u>	397.2	-
8	328.8	Blue mussel, <u>Mytilus edulis</u>	328.8	-
7	>306	Longnose killifish, <u>Fundulus similis</u>	>306	-
6	170	Copepod, <u>Temora longicornis</u>	170	-
5	112.1	Striped mullet, <u>Mugil cephalus</u>	112.1	-
4	62.81	Copepod, <u>Pseudodaptomus coronatus</u>	62.81	-
3	53.2	Pinfish, <u>Lagodon rhomboides</u>	53.2	-
2	40.83	Pacific oyster, <u>Crassostrea gigas</u>	40.83	-

Table 3. (Continued)

<u>Rank*</u>	<u>Genus Mean Acute Value (µg/L)</u>	<u>Species</u>	<u>Species Mean Acute Value (µg/L)**</u>	<u>Species Mean Acute-Chronic Ratio***</u>
1	25.29	Pacific herring, <u>Clupea harengus pallas</u>	25.29	-

* Ranked from most resistant to most sensitive based on Genus Mean Acute Value. Inclusion of "greater than" values does not necessarily imply a true ranking, but does allow use of all genera for which data are available so that the Final Acute Value is not unnecessarily lowered.

** From Table 1.

*** From Table 2.

† Geometric mean of two values in Table 2.

†† Geometric mean of five values in Table 2.

Table 3. (Continued)

Fresh water

$$\text{Final Acute Value} = 10.97 \text{ } \mu\text{g/L (at pH} = 6.5)$$

$$\text{Criterion Maximum Concentration} = (10.97 \text{ } \mu\text{g/L}) / 2 = 5.485 \text{ } \mu\text{g/L (at pH} = 6.5)$$

$$\text{Pooled Slope} = 1.005 \text{ (see Table 1)}$$

$$\ln(\text{Criterion Maximum Intercept}) = \ln(5.485) - (\text{slope} \times 6.5)$$

$$= 1.702 - (1.005 \times 6.5) = -4.830$$

$$\text{Criterion Maximum Concentration} = e^{[1.005(\text{pH})-4.830]}$$

$$\text{Final Acute-Chronic Ratio} = 3.166 \text{ (see text)}$$

$$\text{Final Chronic Value} = (10.97 \text{ } \mu\text{g/L}) / 3.166 = 3.465 \text{ } \mu\text{g/L (at pH} = 6.5)$$

$$\ln(\text{Final Chronic Intercept}) = \ln(3.465) - (\text{slope} \times 6.5)$$

$$= 1.243 - (1.005 \times 6.5) = -5.290$$

$$\text{Final Chronic Value} = e^{[1.005(\text{pH})-5.290]}$$

Salt water

$$\text{Final Acute Value} = 25.05 \text{ } \mu\text{g/L}$$

$$\text{Criterion Maximum Concentration} = (25.05 \text{ } \mu\text{g/L}) / 2 = 12.52 \text{ } \mu\text{g/L}$$

$$\text{Final Acute-Chronic Ratio} = 3.166 \text{ (see text)}$$

$$\text{Final Chronic Value} = (25.05 \text{ } \mu\text{g/L}) / 3.166 = 7.912 \text{ } \mu\text{g/L}$$

Table 4. Toxicity of Pentachlorophenol to Aquatic Plants

<u>Species</u>	<u>Chemical^a</u>	<u>pH</u>	<u>Duration (days)</u>	<u>Effect</u>	<u>Concentration (µg/L)^{b,c}</u>	<u>Reference</u>
<u>FRESHWATER SPECIES</u>						
<u>Alga, Chlorella vulgaris</u>	NaPCP (98%)	-	11	EC50	600	Burrell et al. 1985
<u>Alga, Chlorella pyrenoidosa</u>	-	7.0	3	EC100 (chlorosis)	7.5	Huang and Gloyna 1967
<u>Alga, Chlorella pyrenoidosa</u>	PCP	-	4	EC50 (growth)	7,000	Adema and Vink 1981
<u>Alga, Scenedesmus quadricauda</u>	PCP	-	4	EC50 (growth)	80	Adema and Vink 1981
<u>Alga, Scenedesmus subspicatus</u>	PCP (99%)	7.0	4	EC50	90	Geyer et al. 1985
<u>Alga, Selenastrum capricornutum</u>	NaPCP	-	7	Reduced growth	50	Adams et al. 1985
<u>Alga, Selenastrum capricornutum</u>	PCP	-	4	EC50 (growth)	290	Crossland and Wolff 1985
<u>Alga, Selenastrum capricornutum</u>	PCP	-	4	EC50 (growth)	312	Richter 1982
<u>Alga, Ankistrodesmus braunii</u>	NaPCP (98%)	-	11	EC50	830	Burrell et al. 1985
<u>Duckweed, Lemna minor</u>	-	5.1	2	EC50 (chlorosis)	189.1	Blackman et al. 1955
<u>SALTWATER SPECIES</u>						
<u>Green alga, Chlamydomonas sp.</u>	PCP	-	4	EC50 (population growth)	1,400	Adema and Vink 1981
<u>Green alga, Dunaliella tertiolecta</u>	PCP	30***	4	EC50 (population growth)	206	Walsh et al. 1982
<u>Green alga, Dunaliella tertiolecta</u>	PCP	30***	4	EC50 (population growth)	170	Walsh et al. 1982
<u>Green alga, Dunaliella sp.</u>	PCP	-	4	EC50 (population growth)	3,600	Adema and Vink 1981

Table 4. (Continued)

<u>Species</u>	<u>Chemical*</u>	<u>pH</u>	<u>Duration (days)</u>	<u>Effect</u>	<u>Concentration (µg/L)**</u>	<u>Reference</u>
Golden brown alga, <u>Monochrysis sp.</u>	PCP	-	4	EC50 (population growth)	200	Adema and Vink 1981
Diatom, <u>Skeletonema costatum</u>	PCP	30***	4	EC50 (population growth)	20.30	Walsh et al. 1982
Diatom, <u>Skeletonema costatum</u>	PCP	30***	4	EC50 (population growth)	17.40	Walsh et al. 1982
Diatom, <u>Skeletonema costatum</u>	PCP	30***	4	EC50 (population growth)	17.80	Walsh et al. 1982
Diatom, <u>Thalassiosira pseudonana</u>	PCP	30***	4	EC50 (population growth)	205	Walsh et al. 1982
Diatom, <u>Thalassiosira pseudonana</u>	PCP	30***	4	EC50 (population growth)	189	Walsh et al. 1982
Diatom, <u>Thalassiosira pseudonana</u>	PCP	30***	4	EC50 (population growth)	179	Walsh et al. 1982
Diatom, <u>Phaeodactylum tricornutum</u>	PCP	-	4	EC50 (population growth)	3,000	Adema and Vink 1981
Giant kelp (young fronds), <u>Macrocystis pyrifera</u>	NaPCP (Santobrite)	-	4	EC50 (photosynthetic activity)	277.1	Clendenning and North 1959

* PCP = pentachlorophenol; NaPCP = sodium pentachlorophenate. Percent purity is given in parentheses when available.

** Results are expressed as pentachlorophenol. If the concentrations were not measured and the published results were not reported to be adjusted for purity, the published results were multiplied by the purity if it was reported to be less than 97%.

*** Salinity (g/kg), not pH.

Table 5. Bioaccumulation of Pentachlorophenol by Aquatic Organisms

<u>Species</u>	<u>Chemical*</u>	<u>Concentration in Water (µg/L)**</u>	<u>pH</u>	<u>Duration (days)</u>	<u>Tissue</u>	<u>BCF or BAF***</u>	<u>Reference</u>
<u>FRESHWATER SPECIES</u>							
<u>Rainbow trout (400 g), Salmo gairdneri</u>	PCP (Reagent grade)	0.035	-	115	****	406	Nilmi and McFadden 1982
<u>Rainbow trout (400 g), Salmo gairdneri</u>	PCP (Reagent grade)	0.660	-	115	****	168	Nilmi and McFadden 1982
<u>Fathead minnow (6 mo), Pimephales promelas</u>	-	-	7.49	32	Whole body	770	Velth et al. 1979
<u>Fathead minnow (0.5-1.5 g), Pimephales promelas</u>	PCP (99%)	50	-	31	Whole body	163	Hucklins and Petty 1983
<u>Fathead minnow (0.5-1.5 g), Pimephales promelas</u>	PCP (88%)	50	-	31	Whole body	211	Hucklins and Petty 1983
<u>Fathead minnow (juvenile), Pimephales promelas</u>	Dowicide EC-7 (88% PCP)	3.1-34.6	6.5	32	Whole body	1,066	Spehar et al. 1985
<u>Fathead minnow (juvenile), Pimephales promelas</u>	Dowicide EC-7 (88% PCP)	5.6-58.2	7.5	32	Whole body	434	Spehar et al. 1985
<u>Fathead minnow (juvenile), Pimephales promelas</u>	Dowicide EC-7 (88% PCP)	12.7-161.0	8.0	32	Whole body	426	Spehar et al. 1985
<u>Fathead minnow (juvenile), Pimephales promelas</u>	Dowicide EC-7 (88% PCP)	29.3-125.0	8.5	32	Whole body	281	Spehar et al. 1985
<u>Bluegill (6 mo), Lepomis macrochirus</u>	PCP	100	7.2- 7.7	16	Muscle	7.3	Pruitt et al. 1977; Pierce 1978
<u>SALTWATER SPECIES</u>							
<u>Eastern oyster (adult), Crassostrea virginica</u>	NaPCP	25	-	28	Soft tissue	41	Schimmel et al. 1978
<u>Eastern oyster (adult), Crassostrea virginica</u>	NaPCP	2.5	-	28	Soft tissue	78	Schimmel et al. 1978

Table 5. (Continued)

<u>Species</u>	<u>Chemical*</u>	<u>Concentration in Water (µg/L)**</u>	<u>pH</u>	<u>Duration (days)</u>	<u>Tissue</u>	<u>BCF or BAF***</u>	<u>Reference</u>
<u>Eastern oyster (adult), Crassostrea virginica</u>	NaPCP	3.7±1.4	-	Steady state	Soft tissue	82	Schimmel and Garnas 1985
<u>Eastern oyster (adult), Crassostrea virginica</u>	NaPCP	9.5±3.4	-	Steady state	Soft tissue	34	Schimmel and Garnas 1985
<u>Eastern oyster (adult), Crassostrea virginica</u>	NaPCP	11.3±1.7	-	Steady state	Soft tissue	76	Schimmel and Garnas 1985
<u>Sheepshead minnow (embryo), Cyprinodon variegatus</u>	PCP	18-195	-	Parents exposed 133-142 days	Whole body	19.39 [†]	Parrish et al. 1978
<u>Sheepshead minnow (juvenile), Cyprinodon variegatus</u>	PCP	18-195	-	28	Whole body	31.56 [†]	Parrish et al. 1978
<u>Sheepshead minnow (adult), Cyprinodon variegatus</u>	PCP	18-195	-	151	Whole body	10.75 [†]	Parrish et al. 1978
<u>Longnose killifish, Fundulus similis</u>	PCP	57-120	-	7 (Steady state)	Whole body	64	Trujillo et al. 1982

* PCP = pentachlorophenol; NaPCP = sodium pentachlorophenolate. Percent purity is given in parentheses when available.

** Measured concentration of pentachlorophenol.

*** Bioconcentration factors (BCFs) and bioaccumulation factors (BAFs) are based on measured concentrations of pentachlorophenol in water and in tissue.

**** Whole body minus intestines, liver, and gall bladder.

† Geometric mean of values from four test concentrations.

Table 6. Other Data on Effects of Pentachlorophenol on Aquatic Organisms

<u>Species</u>	<u>Chemical*</u>	<u>pH</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration (µg/L)**</u>	<u>Reference</u>
<u>FRESHWATER SPECIES</u>						
<u>Alga,</u> <u>Scenedesmus pannonicus</u>	PCP	-	4 days	Reduced biomass	100-320	Slooff and Canton 1983
Pond phytoplankton community	PCP	-	10 days	Community structural change	1,000	Boyle et al. 1984
<u>Vascular plant,</u> <u>Elodea canadensis</u>	Dowicide EC-7 (93.7% PCP)	7.7-8.2	21 days	Reduced growth (river water)	380	Hedtke et al. 1986
<u>Duckweed,</u> <u>Lemna minor</u>	PCP	-	7 days	Reduced specific growth rate	1,000-3,200	Slooff and Canton 1983
<u>Duckweed,</u> <u>Lemna minor</u>	Dowicide EC-7 (93.7% PCP)	7.7-8.2	21 days	Reduced growth (river water)	>1,440	Hedtke et al. 1986
<u>Bacterium,</u> <u>Pseudomonas fluorescens</u>	PCP	-	0.3 days	Reduced specific growth rate	1,000-3,200	Slooff and Canton 1983
<u>Cyanobacterium,</u> <u>Microcystis aeruginosa</u>	PCP	-	4 days	Reduced specific growth rate	1,000-3,200	Slooff and Canton 1983
Detritus mixed microbial community	Dowicide EC-7 (88% PCP)	-	56 days	Reduction in biomass and activity	8,131	Fairchild et al. 1984
<u>Amoeba,</u> <u>Amoeba proteus</u>	-	4.5	30 min	Reduced survival	1,332	Smith and Ord 1979
<u>Ciliate protozoan,</u> <u>Colpidium campylum</u>	-	-	43 hr	Minimal active dose	600	Dive et al. 1980
<u>Hydra,</u> <u>Hydra oligactis</u>	PCP (>98%)	-	48 hr	LC50	730	Slooff 1983
<u>Hydra,</u> <u>Hydra oligactis</u>	PCP	-	21 days	Reduced specific growth rate	32-100	Slooff and Canton 1983
<u>Planarian,</u> <u>Dugesia lugubris</u>	PCP (>98%)	-	48 hr	LC50	130	Slooff 1983

Table 6. (Continued)

<u>Species</u>	<u>Chemical^a</u>	<u>pH</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration ($\mu\text{g/L}$)^{b,c}</u>	<u>Reference</u>
<u>Rotifer, Brachionus rubens</u>	-	-	24 hr	LC50	160	Halbach et al. 1983
<u>Rotifer, Brachionus rubens</u>	PCP	-	5 days	LC100	200	Halbach 1984
<u>Rotifer, Brachionus rubens</u>	PCP	-	23 days	Reduced popula- tion density	50	Halbach 1984
<u>Tubificid worm, Branchiura sowerbyi</u>	NaPCP	7.0	96 hr	LC50 (with sediment)	515.5	Chapman et al. 1982a
<u>Tubificid worm, Limnodrilus hoffmeisteri</u>	NaPCP	7.0	96 hr	LC50 (with sediment)	1,150	Chapman et al. 1982a
<u>Tubificid worm, Quilstradrillus multisetosus</u>	NaPCP	7.0	96 hr	LC50 (with sediment)	846.9	Chapman et al. 1982a
<u>Tubificid worm, Spirosperma nikolskyl</u>	NaPCP	7.0	96 hr	LC50 (with sediment)	3,314	Chapman et al. 1982a
<u>Tubificid worm, Stylodrilus heringianus</u>	NaPCP	7.0	96 hr	LC50 (with sediment)	1,243	Chapman et al. 1982a
<u>Tubificid worm, Tubifex tubifex</u>	NaPCP	7.0	96 hr	LC50 (with sediment)	754.8	Chapman et al. 1982a
<u>Mixed tubificid worms, Tubifex tubifex and Limnodrilus hoffmeisteri</u>	NaPCP	7.0	96 hr	LC50	533.9	Chapman et al. 1982c
<u>Mixed tubificid worms, Tubifex tubifex and Limnodrilus hoffmeisteri</u>	NaPCP	7.5	24 hr	LC50	290	Whitley 1968
<u>Mixed tubificid worms, Tubifex tubifex and Limnodrilus hoffmeisteri</u>	NaPCP	8.5	24 hr	LC50	620	Whitley 1968
<u>Mixed tubificid worms, Tubifex tubifex and Limnodrilus hoffmeisteri</u>	NaPCP	9.5	24 hr	LC50	1,290	Whitley 1968

Table 6. (Continued)

<u>Species</u>	<u>Chemical^a</u>	<u>pH</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration ($\mu\text{g/L}$)^{b,c}</u>	<u>Reference</u>
Snail, <u>Lymnaea stagnalis</u>	PCP (>98%)	-	48 hr	LC50	560	Slooff 1983
Snail, <u>Lymnaea stagnalis</u>	PCP	-	40 days	Reduced hatching	3.2- 10	Slooff and Canton 1983
Snail, <u>Physa gyrina</u>	Dowicide EC-7 (93.7% PCP)	7.7- 7.9	96 hr	LC50 (river water)	1,380	Hedtke et al. 1986
Snail, <u>Physa gyrina</u>	Dowicide EC-7 (93.7% PCP)	7.8- 8.5	12 wk	Reduced popu- lation density	365	Zischke et al. 1985
Snail, <u>Physa gyrina</u>	Dowicide EC-7 (93.7% PCP)	7.6- 8.0	96 hr	LC50 (river water)	1,250	Hedtke et al. 1986
Snail, <u>Physa gyrina</u>	Dowicide EC-7 (93.7% PCP)	7.7- 8.1	96 hr	LC50 (river water)	730	Hedtke and Arthur 1985; Hedtke et al. 1986
Snail, <u>Physa gyrina</u>	Dowicide EC-7 (93.7% PCP)	8.2- 8.4	96 hr	LC50 (river water)	620	Hedtke et al. 1986
Snail, <u>Physa gyrina</u>	Dowicide EC-7 (93.7% PCP)	8.0- 8.2	96 hr	LC50 (river water)	220	Hedtke et al. 1986
Snail, <u>Physa gyrina</u>	Dowicide EC-7 (93.7% PCP)	7.9- 8.1	96 hr	LC50 (river water)	260	Hedtke et al. 1986
Snail, <u>Physa gyrina</u>	Dowicide EC-7 (93.7% PCP)	7.8- 7.9	96 hr	LC50 (river water)	580	Hedtke et al. 1986
Snail, <u>Physa gyrina</u>	Dowicide EC-7 (93.7% PCP)	7.7- 7.9	96 hr	LC50 (river water)	810	Hedtke et al. 1986
Cladoceran, <u>Ceriodaphnia affinis/dubia</u>	Dowicide EC-7 (93.7% PCP)	7.9- 8.0	7 days	MATC (river water)	110	Hedtke et al. 1986
Cladoceran, <u>Ceriodaphnia affinis/dubia</u>	Dowicide EC-7 (93.7% PCP)	8.0	48 hr	LC50 (river water)	307	Hedtke et al. 1986
Cladoceran, <u>Ceriodaphnia affinis/dubia</u>	Dowicide EC-7 (93.7% PCP)	7.9- 8.0	48 hr	LC50 (river water)	347	Hedtke et al. 1986

Table 6. (Continued)

<u>Species</u>	<u>Chemical*</u>	<u>pH</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration (µg/L)**</u>	<u>Reference</u>
<u>Cladoceran, Ceriodaphnia reticulata</u>	Dowicide EC-7 (93.7% PCP)	7.8- 7.9	48 hr	LC50 (river water)	240	Hedtke et al. 1986
<u>Cladoceran, Ceriodaphnia reticulata</u>	Dowicide EC-7 (93.7% PCP)	7.8- 8.0	48 hr	LC50 (river water)	700	Hedtke et al. 1986
<u>Cladoceran, Daphnia magna</u>	PCP	-	21 days	LC50	480	Adema 1978
<u>Cladoceran, Daphnia magna</u>	PCP	-	21 days	LC50	510	Adema 1978
<u>Cladoceran, Daphnia magna</u>	PCP	-	21 days	LC50	400	Adema 1978
<u>Cladoceran, Daphnia magna</u>	PCP	-	21 days	LC50	470	Adema 1978
<u>Cladoceran, Daphnia magna</u>	PCP	-	21 days	LC50	430	Adema 1978
<u>Cladoceran, Daphnia magna</u>	PCP	-	21 days	LC50	490	Adema 1978
<u>Cladoceran, Daphnia magna</u>	PCP	-	21 days	LC50	170	Adema 1978
<u>Cladoceran, Daphnia magna</u>	PCP	-	21 days	LC50	190	Adema 1978
<u>Cladoceran, Daphnia magna</u>	PCP	-	14 days	LC50	440	Adema 1978
<u>Cladoceran, Daphnia magna</u>	PCP	-	14 days	LC50	460	Adema 1978
<u>Cladoceran, Daphnia magna</u>	PCP	-	21 days	No effect on reproduction	340	Adema and Vink 1981
<u>Cladoceran, Daphnia magna</u>	PCP	-	21 days	Reduced survival and reproduction	100- 320	Slooff and Canton 1983

Table 6. (Continued)

<u>Species</u>	<u>Chemical[#]</u>	<u>pH</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration (µg/L)^{**}</u>	<u>Reference</u>
<u>Cladoceran (<24 hr), Daphnia magna</u>	-	-	16 days	EC50 (reproduction)	150	Hermens et al. 1984
<u>Cladoceran (6 d), Daphnia magna</u>	PCP (99+%)	7.9	48 hr	EC50 (river water)	254	Brooke et al. Manuscript
<u>Cladoceran (6 d), Daphnia magna</u>	PCP (99+%)	7.6	48 hr	EC50 (river water)	599	Brooke et al. Manuscript
<u>Cladoceran, Simcephalus vetulus</u>	Dowicide EC-7 (93.7% PCP)	7.8- 7.9	48 hr	LC50 (river water)	670	Hedtke and Arthur 1985; Hedtke et al. 1986
<u>Cladoceran, Simcephalus vetulus</u>	Dowicide EC-7 (93.7% PCP)	8.0	48 hr	LC50 (river water)	204	Hedtke et al. 1986
<u>Isopod, Asellus racovitzai</u>	Dowicide EC-7 (93.7% PCP)	7.7- 7.9	96 hr	LC50 (river water)	>7,770	Hedtke et al. 1986
<u>Isopod, Asellus racovitzai</u>	Dowicide EC-7 (93.7% PCP)	7.6- 8.0	96 hr	LC50 (river water)	4,320	Hedtke et al. 1986
<u>Isopod, Asellus racovitzai</u>	Dowicide EC-7 (93.7% PCP)	7.7- 8.1	96 hr	LC50 (river water)	2,370	Hedtke and Arthur 1985; Hedtke et al. 1986
<u>Isopod, Asellus racovitzai</u>	Dowicide EC-7 (93.7% PCP)	7.8- 7.9	96 hr	LC50 (river water)	3,400	Hedtke et al. 1986
<u>Amphipod, Cragonyx pseudogracilis</u>	Dowicide EC-7 (93.7% PCP)	7.7- 7.9	96 hr	LC50 (river water)	3,120	Hedtke et al. 1986
<u>Amphipod, Cragonyx pseudogracilis</u>	Dowicide EC-7 (93.7% PCP)	7.7- 8.1	96 hr	LC50 (river water)	2,770	Hedtke et al. 1986
<u>Amphipod, Cragonyx pseudogracilis</u>	Dowicide EC-7 (93.7% PCP)	7.7- 8.1	96 hr	LC50 (river water)	1,890	Hedtke and Arthur 1985; Hedtke et al. 1986
<u>Amphipod, Cragonyx pseudogracilis</u>	Dowicide EC-7 (93.7% PCP)	8.2- 8.4	96 hr	LC50 (river water)	500	Hedtke et al. 1986
<u>Amphipod, Cragonyx pseudogracilis</u>	Dowicide EC-7 (93.7% PCP)	8.0- 8.1	96 hr	LC50 (river water)	320	Hedtke et al. 1986

Table 6. (Continued)

<u>Species</u>	<u>Chemical^a</u>	<u>pH</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration ($\mu\text{g/L}$)^{b,c}</u>	<u>Reference</u>
<u>Amphipod,</u> <u>Crangonyx pseudogracilis</u>	Dowicide EC-7 (93.7% PCP)	8.0- 8.1	96 hr	LC50 (river water)	220	Hedtke et al. 1986
<u>Amphipod,</u> <u>Crangonyx pseudogracilis</u>	Dowicide EC-7 (93.7% PCP)	7.8- 7.9	96 hr	LC50 (river water)	1,550	Hedtke et al. 1985
<u>Amphipod,</u> <u>Crangonyx pseudogracilis</u>	Dowicide EC-7 (93.7% PCP)	7.7- 7.9	96 hr	LC50 (river water)	2,000	Hedtke et al. 1986
<u>Amphipod (11 mm),</u> <u>Gammarus pseudolimnaeus</u>	PCP (99+%)	7.9	96 hr	LC50 (river water)	451	Brooke et al. Manuscript
<u>Amphipod (11 mm),</u> <u>Gammarus pseudolimnaeus</u>	PCP (99+%)	7.6	96 hr	LC50 (river water)	450	Brooke et al. Manuscript
<u>Amphipod,</u> <u>Hyalofella azteca</u>	PCP (99+%)	7.9	96 hr	LC50 (river water)	286	Brooke et al. Manuscript
<u>Amphipod,</u> <u>Hyalofella azteca</u>	PCP (99+%)	7.6	96 hr	LC50 (river water)	353	Brooke et al. Manuscript
<u>Mayfly,</u> <u>Callibaetis skokianus</u>	Dowicide EC-7 (93.7% PCP)	7.5- 7.7	96 hr	LC50 (river water)	1,780	Hedtke and Arthur 1985; Hedtke et al. 1986
<u>Mayfly,</u> <u>Callibaetis skokianus</u>	Dowicide EC-7 (93.7% PCP)	7.8- 7.9	96 hr	LC50 (river water)	1,300	Hedtke et al. 1986
<u>Mayfly (larvae),</u> <u>Cloeon dipterum</u>	PCP (>98%)	-	48 hr	LC50	5,900	Slooff 1983
<u>Caddisfly,</u> <u>Philarctus quaeris</u>	Dowicide EC-7 (93.7% PCP)	7.5- 7.7	96 hr	LC50 (river water)	1,260	Hedtke and Arthur 1985; Hedtke et al. 1986
<u>Sclomyzid fly (1st instar),</u> <u>Sepedon fuscipennis</u>	NaPCP (90%)	-	96 hr	LC50 (fed)	2,200	McCoy and Joy 1977
<u>Mosquito (1st instar),</u> <u>Culex pipiens</u>	PCP	-	25 days	Reduced survival and development	3,200- 10,000	Slooff and Canton 1983

Table 6. (Continued)

<u>Species</u>	<u>Chemical^a</u>	<u>pH</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration ($\mu\text{g/L}$)^b</u>	<u>Reference</u>
<u>Midge (4th instar), Chironomus riparius</u>	PCP (97%)	7.0	24 hr (15°C) (25°C) (35°C)	LC50	1,176 1,556 631	Fisher 1986
<u>Midge (4th instar), Chironomus riparius</u>	PCP (reagent grade)	4.0 6.0 9.0	24 hr	LC50	384 465 1,948	Fisher and Wadleigh 1986
<u>Sockeye salmon (<1 yr), Oncorhynchus nerka</u>	NaPCP	6.8	21 days	LC50	57	Webb and Brett 1973
<u>Sockeye salmon (<1 yr), Oncorhynchus nerka</u>	NaPCP	6.8	42 days	10% growth inhibition	3.2	Webb and Brett 1973
<u>Sockeye salmon (<1 yr), Oncorhynchus nerka</u>	NaPCP	6.8	56 days	EC50 (growth)	1.74	Webb and Brett 1973
<u>Sockeye salmon (<1 yr), Oncorhynchus nerka</u>	NaPCP	6.8	56 days	EC50 (food conversion)	1.80	Webb and Brett 1973
<u>Chinook salmon (juvenile), Oncorhynchus tshawytscha</u>	NaPCP	7.0- 7.1	96 hr	LC50 (high loading)	72	Iwama and Greer 1979
<u>Chinook salmon (juvenile), Oncorhynchus tshawytscha</u>	-	6.5- 6.9	8 days	LC100 (with bac- terial disease injection)	36	Iwama and Greer 1982
<u>Rainbow trout (alevin), Salmo gairdneri</u>	Santobrite (>90% NaPCP)	-	20 days	11% growth inhibition	28	Chapman 1969
<u>Rainbow trout (alevin), Salmo gairdneri</u>	Santobrite (>90% NaPCP)	-	20 days	18% growth inhibition	28	Chapman 1969
<u>Rainbow trout (alevin), Salmo gairdneri</u>	Santobrite (>90% NaPCP)	-	21 days	18% growth inhibition	28	Chapman 1969

Table 6. (Continued)

<u>Species</u>	<u>Chemical^a</u>	<u>pH</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration ($\mu\text{g/L}$)^{b,c}</u>	<u>Reference</u>
Rainbow trout (alevin), <u>Salmo gairdneri</u>	Santobrite (>90% NaPCP)	-	28 days	12% growth inhibition	28	Chapman 1969
Rainbow trout (embryo and alevin), <u>Salmo gairdneri</u>	Santobrite (>90% NaPCP)	-	41 days	LC100	46	Chapman 1969
Rainbow trout, <u>Salmo gairdneri</u>	Santobrite (>90% NaPCP)	7.8	Fertilization and yolk sac resorption	25% mortality	18	Chapman and Shumway 1978
Rainbow trout (alevin), <u>Salmo gairdneri</u>	Santobrite (>90% NaPCP)	7.8	71 days	9.1% growth reduction	18	Chapman and Shumway 1978
Rainbow trout, <u>Salmo gairdneri</u>	-	-	28 days	27% growth inhibition	7.4	Matida et al. 1976
Rainbow trout (2.1-11.4 g), Nisqually strain, <u>Salmo gairdneri</u>	-	7.4- 7.7	8.8 hr	Median survival time	250	Alexander and Clarke 1978
Rainbow trout (2.1-11.4 g), Idaho strain, <u>Salmo gairdneri</u>	-	7.4- 7.7	7.1 hr	Median survival time	250	Alexander and Clarke 1978
Rainbow trout, <u>Salmo gairdneri</u>	-	8.0	24 hr	Increased respiration	70	Slooff 1979
Rainbow trout, <u>Salmo gairdneri</u>	NaPCP	7.78- 7.89	4 wk post swim-up	Biomass reduc- tion (lower temperature)	20- 80	Hodson and Blunt 1981
Rainbow trout, <u>Salmo gairdneri</u>	NaPCP	7.96- 8.08	4 wk post swim-up	Biomass reduc- tion (higher temperature)	>16	Hodson and Blunt 1981
Rainbow trout (0.81 g), <u>Salmo gairdneri</u>	PCP	7.9	96 hr	LC50 (river water)	70.1	Brooke et al. Manuscript

Table 6. (Continued)

<u>Species</u>	<u>Chemical*</u>	<u>pH</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration ($\mu\text{g/L}$)**</u>	<u>Reference</u>
<u>Rainbow trout (0.81 g), Salmo gairdneri</u>	PCP (99+%)	7.6	96 hr	LC50 (river water)	85.6	Brooke et al. Manuscript
<u>Atlantic salmon, Salmo salar</u>	NaPCP	-	24 hr	Altered temperature preference	46	Peterson 1976
<u>Brown trout (4.5 g), Salmo trutta</u>	PCP (99%)	-	24 hr	LC50	200	Hattula et al. 1981
<u>Central stoneroller (10 g), Campostoma anomalum</u>	NaPCP	7.6	4.3 hr	LC100	400	Goodnight 1942
<u>Goldfish (1.5 g), Carassius auratus</u>	PCP	-	24 hr	LC90	1,600	Lemma and Yau 1974
<u>Goldfish (108.0 g), Carassius auratus</u>	PCP	-	24 hr	LC90	1,600	Lemma and Yau 1974
<u>Goldfish (adult), Carassius auratus</u>	PCP (99+%)	7.59	46 hr	LC50	270	Cardwell et al. 1976
<u>Goldfish (juvenile), Carassius auratus</u>	NaPCP	7.59	21 hr	LC50	369	Cardwell et al. 1976
<u>Goldfish (juvenile), Carassius auratus</u>	NaPCP	7.59	120 hr	LC50	253	Cardwell et al. 1976
<u>Goldfish (juvenile), Carassius auratus</u>	NaPCP	7.59	336 hr	LC50	189	Cardwell et al. 1976
<u>Goldfish (2 g), Carassius auratus</u>	-	-	24 hr	LC50	270	Kobayashi et al. 1979
<u>Goldfish (1.3 g), Carassius auratus</u>	-	5.5	24 hr	LC50	52	Kobayashi and Kishino 1980
<u>Goldfish (1.3 g), Carassius auratus</u>	-	6	24 hr	LC50	60	Kobayashi and Kishino 1980
<u>Goldfish (1.3 g), Carassius auratus</u>	-	7	24 hr	LC50	82	Kobayashi and Kishino 1980

Table 6. (Continued)

<u>Species</u>	<u>Chemical*</u>	<u>pH</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration ($\mu\text{g/L}$)**</u>	<u>Reference</u>
<u>Goldfish (1.3 g), Carassius auratus</u>	-	8	24 hr	LC50	250	Kobayashi and Kishino 1980
<u>Goldfish (1.3 g), Carassius auratus</u>	-	9	24 hr	LC50	2,200	Kobayashi and Kishino 1980
<u>Goldfish (1.3 g), Carassius auratus</u>	-	10	24 hr	LC50	16,000	Kobayashi and Kishino 1980
<u>Common carp (eyed embryo), Cyprinus carpio</u>	PCP	-	24 hr	LC50	180	Hashimoto et al. 1982
<u>Common carp (1-3 days), Cyprinus carpio</u>	PCP	-	24 hr	LC50	140	Hashimoto et al. 1982
<u>Common carp (5-6 days), Cyprinus carpio</u>	PCP	-	24 hr	LC50	150	Hashimoto et al. 1982
<u>Common carp (9-10 days), Cyprinus carpio</u>	PCP	-	24 hr	LC50	130	Hashimoto et al. 1982
<u>Common carp (17-19 days), Cyprinus carpio</u>	PCP	-	24 hr	LC50	110	Hashimoto et al. 1982
<u>Common carp (25-33 days), Cyprinus carpio</u>	PCP	-	24 hr	LC50	100	Hashimoto et al. 1982
<u>Common carp (50-60 days), Cyprinus carpio</u>	PCP	-	24 hr	LC50	110	Hashimoto et al. 1982
<u>Common carp (70-80 days), Cyprinus carpio</u>	PCP	-	24 hr	LC50	110	Hashimoto et al. 1982
<u>Common carp (15.1-28.8 g), Cyprinus carpio</u>	PCP	-	50 min	LT50	3,000	Peer et al. 1983
<u>Silverjaw minnow (2 g), Epiplatys buccata</u>	NaPCP	7.6	6.25 hr	LC100	400	Goodnight 1942
<u>Redfin shiner (2 g), Notropis umbratilis</u>	NaPCP	7.6	2.67 hr	LC100	600	Goodnight 1942

Table 6. (Continued)

<u>Species</u>	<u>Chemical^a</u>	<u>pH</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration ($\mu\text{g/L}$)^{b,c}</u>	<u>Reference</u>
<u>Steelcolor shiner (2 g), Notropis atherinoides</u>	NaPCP	7.6	5.25 hr	LC100	400	Goodnight 1942
<u>Bluntnose minnow (3 g), Pimephales notatus</u>	NaPCP	7.6	7.45 hr	LC100	400	Goodnight 1942
<u>Fathead minnow (2 inches), Pimephales promelas</u>	NaPCP	7.4- 7.5	24 hr	LC50	300- 320	Crandall and Goodnight 1959
<u>Fathead minnow (juvenile), Pimephales promelas</u>	NaPCP	7.83	336 hr	LC50	153	Cardwell et al. 1976
<u>Fathead minnow, Pimephales promelas</u>	PCP	-	48 hr	LC50	210	Slooff 1982
<u>Fathead minnow, Pimephales promelas</u>	Dowicide EC-7 (93.7% PCP)	7.7- 8.4	12 wk	Reduced growth and larval drift	111	Zischke et al. 1985
<u>Fathead minnow (adult), Pimephales promelas</u>	Dowicide EC-7 (93.7% PCP)	8.0- 8.1	96 hr	LC50 (river water)	300	Hedtke et al. 1986
<u>Fathead minnow (adult), Pimephales promelas</u>	Dowicide EC-7 (93.7% PCP)	7.9- 8.2	96 hr	LC50 (river water)	190	Hedtke et al. 1986
<u>Fathead minnow (adult), Pimephales promelas</u>	Dowicide EC-7 (93.7% PCP)	7.9- 8.1	96 hr	LC50 (river water)	170	Hedtke et al. 1986
<u>Fathead minnow (adult), Pimephales promelas</u>	Dowicide EC-7 (93.7% PCP)	8.0- 8.2	96 hr	LC50 (river water)	160	Hedtke et al. 1986
<u>Fathead minnow (adult), Pimephales promelas</u>	Dowicide EC-7 (93.7% PCP)	7.7- 8.1	96 hr	LC50 (river water)	120	Hedtke et al. 1986
<u>Fathead minnow (adult), Pimephales promelas</u>	Dowicide EC-7 (93.7% PCP)	7.9- 8.1	96 hr	LC50 (river water)	208	Hedtke and Arthur 1985; Hedtke et al. 1986
<u>Fathead minnow (adult), Pimephales promelas</u>	Dowicide EC-7 (93.7% PCP)	7.4- 7.9	96 hr	LC50 (river water)	120	Hedtke et al. 1986
<u>Fathead minnow (juvenile), Pimephales promelas</u>	Dowicide EC-7 (93.7% PCP)	7.8- 8.2	96 hr	LC50 (river water)	396	Hedtke et al. 1986

Table 6. (Continued)

<u>Species</u>	<u>Chemical*</u>	<u>pH</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration (µg/L)**</u>	<u>Reference</u>
<u>Fathead minnow (juvenile), Pimephales promelas</u>	Dowicide EC-7 (93.7% PCP)	7.8- 8.1	96 hr	LC50 (river water)	510	Hedtke et al. 1986
<u>Fathead minnow (fry), Pimephales promelas</u>	Dowicide EC-7 (93.7% PCP)	7.9- 8.2	96 hr	LC50 (river water)	314	Hedtke et al. 1986
<u>Fathead minnow (embryo), Pimephales promelas</u>	Dowicide EC-7 (93.7% PCP)	7.8- 8.1	96 hr	LC50 (river water)	465	Hedtke et al. 1986
<u>Fathead minnow (embryo), Pimephales promelas</u>	Dowicide EC-7 (93.7% PCP)	8.0- 8.1	96 hr	LC50 (river water)	480	Hedtke et al. 1986
<u>Fathead minnow (embryo, fry), Pimephales promelas</u>	Dowicide EC-7 (93.7% PCP)	7.8- 8.1	32 days	Reduced survival or growth (river water)	118- 176	Hedtke et al. 1986
<u>Fathead minnow (<2 wk), Pimephales promelas</u>	PCP (purified; 99%)	7.4	90 days	Decreased growth 24% Decreased growth 18%	85 142	Cleveland et al. 1982
<u>Fathead minnow (<2 wk), Pimephales promelas</u>	Dowicide EC-7 (91%)	7.4	90 days	Increased growth 18% Increased growth 21%	60 139	Cleveland et al. 1982
<u>Fathead minnow (<2 wk), Pimephales promelas</u>	PCP (industrial composite)	7.4	90 days	Decreased growth 20% Decreased growth 40% 100% mortality	13 27 67	Cleveland et al. 1982
<u>Fathead minnow (7 day), Pimephales promelas</u>	PCP (ultrapure)	7.4	90 days	Reduced growth 10% Reduced growth 17%	66 130	Hamilton et al. 1986
<u>Creek chub (12 g), Semotilus atromaculatus</u>	NaPCP	7.6	3.92 hr	LC100	600	Goodnight 1942
<u>White sucker, Catostomus commersoni</u>	Dowicide EC-7 (93.7% PCP)	7.7- 8.2	96 hr	LC50 (river water)	85	Hedtke et al. 1986

Table 6. (Continued)

<u>Species</u>	<u>Chemical^a</u>	<u>pH</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration ($\mu\text{g/L}$)^b</u>	<u>Reference</u>
<u>Blackstripe topminnow (3 g), Fundulus notatus</u>	NaPCP	7.6	9.75 hr	LC100	800	Goodnight 1942
<u>Guppy, Poecilia reticulata</u>	NaPCP	8.4- 8.6	180 days	Damaged liver and kidney	462	Crandall and Goodnight 1962,1963
<u>Guppy, Poecilia reticulata</u>	PCP	-	60 days	No effect on growth	180	Adema and Vink 1981
<u>Guppy, Poecilia reticulata</u>	PCP	-	28 days	Reduced growth	100- 320	Slooff and Canton 1983
<u>Guppy (2-3 mo), Poecilia reticulata</u>	-	7.7	24 hr	LC50	40	Benoit-Guyod et al. 1984a
<u>Threespine stickleback (0.6 g), Gasterosteus aculeatus</u>	PCP	-	24 hr	LC90	370	Lemma and Yau 1974
<u>Orangespotted sunfish (2 g), Lepomis humilis</u>	NaPCP	7.6	6.75 hr	LC100	400	Goodnight 1942
<u>Bluegill (juvenile), Lepomis macrochirus</u>	PCP (99+%)	7.94	30 hr	LC50	303	Cardwell et al. 1976
<u>Bluegill (juvenile), Lepomis macrochirus</u>	PCP (99+%)	7.94	336 hr	LC50	215	Cardwell et al. 1976
<u>Bluegill, Lepomis macrochirus</u>	Dowicide EC-7 (93.7% PCP)	7.7- 7.9	96 hr	LC50 (river water)	200	Hedtke and Arthur 1985; Hedtke et al. 1986
<u>Bluegill, Lepomis macrochirus</u>	Dowicide EC-7 (93.7% PCP)	7.8- 7.9	96 hr	LC50 (river water)	270	Hedtke et al. 1986
<u>Bluegill, Lepomis macrochirus</u>	Dowicide EC-7 (93.7% PCP)	7.6- 8.6	12 wk	Reduced growth and larval drift	40	Zischke et al. 1985

Table 6. (Continued)

<u>Species</u>	<u>Chemical*</u>	<u>pH</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration ($\mu\text{g/L}$)**</u>	<u>Reference</u>
<u>Largemouth bass,</u> <u>Micropterus salmoides</u>	-	7.0	<4 hr	Threshold oper- cular rhythm response	47 50	Morgan 1976, 1977
<u>Largemouth bass (4.1 g),</u> <u>Micropterus salmoides</u>	PCP (99%)	7.2	7 days	Reduced growth rate and food conversion efficiency	50.4	Mathers et al. 1985
<u>Mozambique tilapia (2.0 g),</u> <u>Tilapia mossambica</u>	PCP	-	24 hr	LC90	800	Lenma and Yau 1974
<u>Leopard frog (tadpole),</u> <u>Rana pipiens</u>	NaPCP	7.6	9.5 hr	LC100	800	Goodnight 1942
<u>African clawed toad (3-4 wk),</u> <u>Xenopus laevis</u>	-	-	48 hr	LC50	260	Slooff 1982; Slooff and Baerelman 1980
<u>African clawed toad (<2 d),</u> <u>Xenopus laevis</u>	PCP	-	100 days	Reduced survi- val and growth	32- 100	Slooff and Canton 1983
<u>SALTWATER SPECIES</u>						
<u>Photoluminescent bacterium,</u> <u>Photobacterium phosphoreum</u>	PCP	-	5 min	50 % reduction in light output	80	Curtis et al. 1982
<u>Photoluminescent bacterium,</u> <u>Photobacterium phosphoreum</u>	PCP	-	5 min	50 % reduction in light output	924	Ribo and Kaiser 1983
<u>Microfungal populations</u> <u>in microcosms</u>	PCP	10-17***	8 wk	Successional change in micro- fungal species composition	140	Cook et al. 1980
<u>Golden brown alga,</u> <u>Monochrysis lutheri</u>	NaPCP	28.9***	12- 15 days	Decreased cell numbers	270,6	Woelke 1965
<u>Diatom,</u> <u>Skeletonema costatum</u>	PCP	25***	7 days	EC50 (cell division)	2,000	Erickson and Freeman 1978

Table 6. (Continued)

<u>Species</u>	<u>Chemical^a</u>	<u>Salinity (g/kg)</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration (µg/L)^{b,c}</u>	<u>Reference</u>
Diatom, <u>Thalassiosira pseudonana</u>	PCP	25	7 days	EC50 (cell division)	500	Erickson and Freeman 1978
Diatom, <u>Thalassiosira pseudonana</u>	PCP	26-29	24 hr	EC50 (cell division)	250	Erickson 1981
Diatom, <u>Thalassiosira pseudonana</u>	PCP	26-29	48 hr	EC50 (cell division)	300	Erickson 1981
Dinoflagellate, <u>Glenodinium halli</u>	PCP	25	7 days	EC50 (cell division)	1,000	Erickson and Freeman 1978
Microflagellate, <u>Isochrysis galbana</u>	PCP	25	7 days	EC50 (cell division)	250	Erickson and Freeman 1978
Sea grass, <u>Thalassia testudinum</u>	PCP	30	40 hr	EC50 (reduced oxygen evolu- tion)	740	Walsh et al. 1982
Giant kelp, <u>Macrocystis pyrifera</u>	PCP	-	4 days	Photosynthesis inhibition	2,660	Lammering and Burbank 1960; reported in Bulkema et al. 1979
Giant kelp, <u>Macrocystis pyrifera</u>	PCP	-	2 days	Photosynthesis inhibition	1,000	Lammering and Burbank 1960; reported in Bulkema et al. 1979
Nematodes, (Mesobenthic communities)	Dowicide G-ST	-	9-13 wk	Decrease in biomass and density; shift in species composition	≥161	Cantelmo and Rao 1978b,c
Polychaete worm, <u>Nereis virens</u>	PCP	32-35	48 hr	Significant decrease in coelomic fluid osmolality, coupled with mortality	720	Carr and Neff 1981
Polychaete worm, <u>Nereis virens</u>	PCP	32-35	14 days	BCF = 280	-	Carr and Neff 1981

Table 6. (Continued)

<u>Species</u>	<u>Chemical^a</u>	<u>Salinity (g/kg)</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration (µg/L)^{b,c}</u>	<u>Reference</u>
<u>Polychaete worm, Nereis virens</u>	PCP	32-35	2 mo	Significant increase in ascorbic acid levels and significant decrease in glycogen reserves	100	Carr and Neff 1981
<u>Polychaete worm, Ophryotrocha diadema</u>	PCP	33	37 days	Apparent reduc- tion in repro- duction	>98	Hooftman and Vink 1980
<u>Polychaete worm, Ophryotrocha diadema</u>	PCP	-	41 days	EC50 (reproduction)	23	Adema and Vink 1981
<u>Polychaete worm, Ophryotrocha diadema</u>	PCP	-	48 days	Apparent reduc- tion in repro- duction	>11	Adema and Vink 1981
<u>Polychaete worm (adult), Arenicola cristata</u>	Dowicide G-ST	22-24	6 days	Reduced feeding activity	>80	Rubinstein 1978
<u>Japanese littleneck, Tapes philippinarum</u>	PCP	-	120 days	Lethal	100	Tomiyama et al. 1962
<u>Japanese littleneck, Tapes philippinarum</u>	PCP	-	24 hr	BCF = 6 to 189 depending on tissue	-	Kobayashi et al. 1969
<u>Common Atlantic slippershell (larva), Crepidula fornicata</u>	PCP	-	7 days	LC50	460	Adema and Vink 1981
<u>Blue mussel (larva), Mytilus edulis</u>	NaPCP	28	48 hr	22.1% abnormal larvae	369.5	Dimick and Breese 1965; Woelke 1972
<u>Blue mussel (larva), Mytilus edulis</u>	NaPCP	24	48 hr	69.1% abnormal larvae	369.5	Dimick and Breese 1965; Woelke 1972

Table 6. (Continued)

<u>Species</u>	<u>Chemical*</u>	<u>Salinity (g/kg)</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration (µg/L)**</u>	<u>Reference</u>
<u>Blue mussel, Mytilus edulis</u>	PCP	-	14 days	LC50	750	Adema and Vink 1981
<u>Eastern oyster (embryo), Crassostrea virginica</u>	PCP	-	48 hr	100% abnormal larvae	250- 5,000	Davis and Hildu 1969
<u>Eastern oyster (larva), Crassostrea virginica</u>	PCP	-	14 days	100% mortality	100- 500	Davis and Hildu 1969
<u>Eastern oyster (adult), Crassostrea virginica</u>	NaPCP	20.3	192 hr	EC50 (growth)	76.50	Schimmel et al. 1978
<u>Quahog clam, Mercenaria mercenaria</u>	NaPCP	28-30	18 wk	Reduced resis- tance to bacter- ial infection	459.9	Anderson et al. 1981
<u>Quahog clam, Mercenaria mercenaria</u>	NaPCP	28-30	8 wk	BCF = 100	-	Anderson et al. 1981
<u>Quahog clam, Mercenaria mercenaria</u>	NaPCP	28-30	18 wk	BCF = 54	-	Anderson et al. 1981
<u>Copepod, Pseudodaptomus coronatus</u>	NaPCP	18	96 hr	Significant increase in feeding rate	74.82	Hauch et al. 1980
<u>Brown shrimp (adult), Penaeus aztecus</u>	NaPCP	26.5	96 hr	BCF = 0.2165 (geometric mean of 4 values)	-	Schimmel et al. 1978
<u>Grass shrimp (adult), Palaemonetes pugio</u>	NaPCP	10	12 days	Histological changes (gill necrosis; loss of microvilli and epithelial rupture of midgut and hepatopancreas; and mitochondria compartmentalization)	923.7	Doughtie and Rao 1978; Rao and Doughtie 1984

Table 6. (Continued)

<u>Species</u>	<u>Chemical[#]</u>	<u>Salinity (g/kg)</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration (µg/L)^{##}</u>	<u>Reference</u>
<u>Grass shrimp (adult), Palaemonetes pugio</u>	NaPCP	24.3	96 hr	BCF = 1.506 (geometric mean of 4 values)	-	Schimmel et al. 1978
<u>Grass shrimp (post molt), Palaemonetes pugio</u>	NaPCP	-	13 hr	100% mortality, 3 hrs after molt	4,618	Cantelmo et al. 1978
<u>Grass shrimp (adult), Palaemonetes pugio</u>	NaPCP	10	9 days	50% reduction in limb regen- eration	436.9 521.9	Rao et al. 1978, 1979, 1981
<u>Grass shrimp (adult), Palaemonetes pugio</u>	NaPCP	10	96 hr	Significant increase in exuvial dry weight	923	Brannon and Conklin 1978
<u>Grass shrimp (adult), Palaemonetes pugio</u>	NaPCP	-	24-36 hr	Significant increase or decrease in oxygen consump- tion followed by death	>9.237	Cantelmo et al. 1978
<u>Grass shrimp (adult, new molt), Palaemonetes pugio</u>	PCP	10	1 hr	BCF = 150	-	Rao et al. 1981
<u>Grass shrimp (adult, intermolt), Palaemonetes pugio</u>	PCP	10	1 hr	BCF = 30	-	Rao et al. 1981
<u>Longnose killifish, (juvenile), Fundulus similis</u>	NaPCP	22.9	96 hr	BCF = 27.03 (geometric mean of 4 values)	-	Schimmel et al. 1978
<u>Striped mullet, Mugil cephalus</u>	NaPCP	25.5	96 hr	BCF = 7.446 (geometric mean of lowest two values)	-	Schimmel et al. 1978

Table 6. (Continued)

<u>Species</u>	<u>Chemical*</u>	<u>Salinity (g/kg)</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration (µg/L)**</u>	<u>Reference</u>
<u>Striped mullet, Mugil cephalus</u>	PCP	33	1-120 hr	Elevation of plasma cortisol concentration, hyperglycemia, depletion of liver glycogen, increase in liver ascorbate concentration	100-200	Thomas et al. 1981
<u>Winter flounder, Pseudopleuronectes americanus</u>	PCP	-	15 days	Elevated levels of acid-soluble thiol in liver	50	Thomas and Wofford 1984
Benthic macrofauna	PCP	27	5 wk	Significant reduction in species richness and faunal numbers	55	Tagatz et al. 1983
Benthic macrofauna	PCP	16.7	9 wk	Significant reduction in species richness and faunal numbers	76	Hansen and Tagatz 1980; Tagatz et al. 1977,1980
Benthic macrofauna	PCP	16.7	9 wk	No significant effects on colonization	7	Tagatz et al. 1977,1980
Benthic macrofauna	PCP	18	1 wk	Significant reduction in species richness and total faunal numbers	140 141	Tagatz et al. 1981
Benthic macrofauna	Dielder G-ST	22	13 wk	Significant reduction in total faunal numbers	15.8	Hansen and Tagatz 1980; Tagatz et al. 1978,1980

Table 6. (Continued)

<u>Species</u>	<u>Chemical*</u>	<u>Salinity (g/kg)</u>	<u>Duration</u>	<u>Effect</u>	<u>Concentration (µg/L)**</u>	<u>Reference</u>
Benthic macrofauna	Dowicide G-ST	22	13 wk	No significant effect on colonization	1.8	Hansen and Tagatz 1980; Tagatz et al. 1978, 1980

* PCP = pentachlorophenol; NaPCP = sodium pentachlorophenate. Percent purity is given in parentheses when available.

** Results are expressed as pentachlorophenol. If the concentrations were not measured and the published results were not reported to be adjusted for purity, the published results were multiplied by the purity if it was reported to be less than 97%.

** Salinity (g/kg), not pH.

REFERENCES

Adams, N., K.H. Goulding and A.J. Dobbs. 1985. Toxicity of eight water-soluble organic chemicals to Selanstrum capricornutum: A study of methods for calculating toxic values using different growth parameters. Arch. Environ. Contam. Toxicol. 14:333-345.

Adelman, I.R. and L.L. Smith, Jr. 1976. Standard test fish development. Part 1. Fathead minnows (Pimephales promelas) and goldfish (Carassius auratus) as standard fish in bioassays and their reaction to potential reference toxicants. EPA-600/3-76-061a. National Technical Information Service, Springfield, VA.

Adelman, I.R., L.L. Smith, Jr. and G.D. Seisennop. 1976a. Effect of size or age of goldfish and fathead minnows on use of pentachlorophenol as a reference toxicant. Water Res. 10:685-687.

Adelman, I.R., L.L. Smith, Jr. and G.D. Seisennop. 1976b. Acute toxicity of sodium chloride, pentachlorophenol, Guthion, and hexavalent chromium to fathead minnows (Pimephales promelas) and goldfish (Carassius auratus). J. Fish. Res. Board Can. 33:203-208.

Adema, D.M.M. 1978. Daphnia magna as a test animal in acute and chronic toxicity tests. Hydrobiologia 59:125-134.

Adema, D.M.M. and G.J. Vink. 1981. A comparative study of the toxicity of 1,1,2-trichloroethane, dieldrin, pentachlorophenol and 3,4-dichloroaniline for marine and fresh water organisms. *Chemosphere* 10:533-554.

Ahlborg, U.G. and T.M. Thunberg. 1980. Chlorinated phenols: Occurrence, toxicity, metabolism, and environmental impact. *Crit. Rev. Toxicol.* 7:1-35.

Akitake, H. and K. Kobayashi. 1975. Studies on the metabolism of chlorophenols in fish-III. Isolation and identification of a conjugated PCP excreted by goldfish. *Bull. Jpn. Soc. Sci. Fish.* 41:321-327.

Alexander, D.G. and R.M.V. Clarke. 1978. The selection and limitations of phenol as a reference toxicant to detect differences in sensitivity among groups of rainbow trout (Salmo gairdneri). *Water Res.* 12:1085-1090.

Alexander, H.C., K.M. Bodner and M.A. Mayes. 1983. Evaluation of the OECD "fish prolonged toxicity study at least 14 days". *Chemosphere* 12:415-423.

Anderson, P.D. and L.J. Weber. 1975. Toxic response as a quantitative function of body size. *Toxicol. Appl. Pharmacol.* 33:471-483.

Anderson, R.S., C.S. Giam, L.E. Ray and M.R. Tripp. 1981. Effects of environmental pollutants on immunological competency of the clam Mercenaria mercenaria: Impaired bacterial clearance. *Aquat. Toxicol.* 1:187-195.

Anderson, R.S., C.S. Giam and L.E. Ray. 1984. Effects of hexachlorobenzene and pentachlorophenol on cellular and humoral immune parameters in Glycera dibranchiata. Mar. Environ. Res. 14:317-326.

Batte, E.G. and L.E. Swanson. 1952. Laboratory evaluation of organic compounds as molluscicides and ovicides, II. J. Parasitol. 38:65-68.

Benoit-Guyod, J.L., C. Andre and K. Clavel. 1984a. Chlorophenols: Degradation and toxicity. J. Francais Hydrol. 15:249-263.

Benoit-Guyod, J.L., C. Andre, G. Taillandier, J. Rochat and A. Boucherle. 1984b. Toxicity and QSAR of chlorophenols on Lebistes reticulatus. Ecotoxicol. Environ. Safety 8:227-235.

Bentley, R.E., T. Heitnuller, B.H. Sleight, III and P.R. Parrish. 1975. Acute toxicity of pentachlorophenol to bluegill (Lepomis macrochirus), rainbow trout (Salmo gairdneri), and pink shrimp (Penaeus duorarum). No. WA-6-99-1414-B. Criteria Branch, U.S. EPA, Washington, DC.

Berglind, R. and G. Dave. 1984. Acute toxicity of chromate, DDT, PCP, TPBs, and zinc to Daphnia magna cultured in hard and soft water. Bull. Environ. Contam. Toxicol. 33:63-68.

Bevenue, A. and H. Beckman. 1967. Pentachlorophenol: A discussion of its properties and its occurrence as a residue in human and animal tissues. Residue Rev. 19:83-134.

Bevenue, A., J. Wilson, L.J. Casarett and H.W. Klemmer. 1967. A survey of pentachlorophenol content in human urine. Bull. Environ. Contam. Toxicol. 2:319-333.

Blackman, G.E., M.H. Parke and G. Carton. 1955. The physiological activity of substituted phenols. I. Relationships between chemical structure and physiological activity. Arch. Biochem. Biophys. 54:45-54.

Bols, N.C., S.A. Boliska, D.G. Dixon, P.V. Hodson and K.L.E. Kaiser. 1985. The use of fish cell cultures as an indication of contaminant toxicity to fish. Aquat. Toxicol. 6:147-155.

Borthwick, P.W. and S.C. Schimmel. 1978. Toxicity of pentachlorophenol and related compounds to early life stages of selected estuarine animals. In: Pentachlorophenol: Chemistry, pharmacology and environmental toxicology. Rao, K.R. (Ed.). Plenum Publishing Company, New York, NY. pp. 141-146.

Bose, A.K. and H. Fujiwara. 1978. Fate of pentachlorophenol in the blue crab, Callinectes sapidus. In: Pentachlorophenol: Chemistry, pharmacology and environmental toxicology. Rao, K.R. (Ed.). Plenum Publishing Company, New York, NY. pp. 83-88.

Bostrom, S.L. and R.G. Johansson. 1972. Effects of pentachlorophenol on enzymes involved in energy metabolism in the liver of the eel. Comp. Biochem. Physiol. 41B:359-369.

Boyle, T.P., J. Sebaugh and E. Robinson-Wilson. 1984. A hierarchical approach to the measurement of changes in community structure induced by environmental stress. *J. Test. Eval.* 12:241-245.

Brannon, A.C. and P.J. Conklin. 1978. Effect of pentachlorophenate on exoskeletal calcium in the grass shrimp, Palaemonetes pugio. In: Pentachlorophenol: Chemistry, pharmacology and environmental toxicology. Rao, K.R. (Ed.). Plenum Publishing Company, New York, NY. pp. 205-211.

Branson, D.R. 1980. Prioritization of chemicals according to the degree of hazard in the aquatic environment. *Environ. Health Perspect.* 34:133-138.

Brockway, D.L., P.D. Smith and F.E. Stancil. 1984. Fate and effects of pentachlorophenol in hard- and soft-water microcosms. *Chemosphere* 13:1363-1377.

Brooke, L.T., D.J. Call, D.E. Hammermeister, A. Hoffman and C.E. Northcott. Manuscript. Acute toxicities of five chemicals in different natural waters. Center for Lake Superior Environmental Studies, University of Wisconsin-Superior, Superior, WI.

Brown, J.A., P.H. Johansen, P.W. Colgan and R.A. Mathers. 1985. Changes in the predator-avoidance behaviour of juvenile guppies (Poecilia reticulata) exposed to pentachlorophenol. *Can. J. Zool.* 63:2001-2005.

Buhler, D.R., M.E. Rasmusson and H.S. Nakaue. 1973. Occurrence of hexachlorophene and pentachlorophenol in sewage and water. Environ. Sci. Technol. 7:929-939.

Buikema, A.L., Jr., M.J. McGinnis and J. Cairns, Jr. 1979. Phenolics in aquatic ecosystems: A selected review of recent literature. Mar. Environ. Res. 2:87-101.

Burrell, R.E., W.E. Inniss and C.I. Mayfield. 1985. Detection and analysis of interactions between atrazine and sodium pentachlorophenate with single and multiple algal-bacterial populations. Arch. Environ. Contam. Toxicol. 14:167-177.

Butte, W., M. Kirsch and J. Denker. 1983. The determination of pentachlorophenol and tetrachlorophenols in Wadden sediment and clams (Mya arenaria) using triethylsulfonium hydroxide for extraction and pyrolytic ethylation. Int. J. Environ. Anal. Chem. 13:141-153.

Call, D.J., L.T. Brooke, N. Ahmad and J.E. Richter. 1983. Toxicity and metabolism studies with EPA priority pollutants and related chemicals in freshwater organisms. EPA-600/3-83-095 or PB83-263665. National Technical Information Service, Springfield, VA.

Callahan, M.A., M.W. Slimak, N.W. Gabel, I.P. May, C.F. Fowler, J.R. Freed, P. Jennings, R.L. Durfee, F.C. Whitmore, B. Maestri, W.R. Mabey, B.R. Holt and C. Gould. 1979. Water-related environmental fate of 129 priority pollutants. Vol II. EPA-440/4-79-029b. National Technical Information Service, Springfield, VA. pp. 87-1 to 87-13.

Cantelmo, A.C. and K.R. Rao. 1978a. The effects of pentachlorophenol (PCP) and 2,4-dinitrophenol (DNP) on the oxygen consumption of tissues from the blue crab, Callinectes sapidus, under different osmotic conditions. Comp. Biochem. Physiol. 60C:215-219.

Cantelmo, A.C. and K.R. Rao. 1978b. Effects of pentachlorophenol on the meiobenthic nematodes in an experimental system. In: Pentachlorophenol: Chemistry, pharmacology and environmental toxicology. Rao, K.R. (Ed.). Plenum Publishing Company, New York, NY. pp. 165-174.

Cantelmo, F.R. and K.R. Rao. 1978c. Effect of pentachlorophenol (PCP) on meiobenthic communities established in an experimental system. Mar. Biol. (Berl.) 46:17-22.

Cantelmo, A.C., P.J. Conklin, F.R. Fox and K.R. Rao. 1978. Effects of sodium pentachlorophenate and 2,4-dinitrophenol on respiration in crustaceans. In: Pentachlorophenol: Chemistry, pharmacology and environmental toxicology. Rao, K.R. (Ed.). Plenum Publishing Company, New York, NY. pp. 251-263.

Canton, J.H. and D.M.M. Adema. 1978. Reproducibility of short-term and reproduction toxicity experiments with Daphnia magna and comparison of Daphnia magna with Daphnia pulex and Daphnia cucullata in short-term experiments. *Hydrobiologia* 59:135-140.

Canton, J.H. and W. Slooff. 1979. A proposal to classify compounds and to establish water quality criteria based on laboratory data. *Ecotoxicol. Environ. Safety* 3:126-132.

Cardwell, R.D., D.G. Foreman, I.R. Payne and D.J. Wilbur. 1976. Acute toxicity of selected toxicants to six species of fish. PB-252488 or EPA-600/3-76-008. National Technical Information Service, Springfield, VA.

Carr, R.S. and J.M. Neff. 1981. Biochemical indices of stress in the sandworm Neanthes virens (Sars). I. Responses to pentachlorophenol. *Aquat. Toxicol.* 1:313-327.

Cessna, A.J. and R. Grover. 1978. Spectrophotometric determination of dissociation constants of selected acidic herbicides. *J. Agric. Food Chem.* 26:289-292.

Chapman, G.A. 1969. Toxicity of pentachlorophenol to trout alevins. Ph.D. thesis. Oregon State University, Corvallis, OR. Available from: University Microfilms, Ann Arbor, MI. Order No. 69-19,906.

Chapman, G.A. and D.L. Shumway. 1978. Effects of sodium pentachlorophenate on survival and energy metabolism of embryonic and larval steelhead trout. In: Pentachlorophenol. Rao, K.R. (Ed.). Plenum, New York, NY. pp. 285-299. Also available as PB-287600 or EPA-600/J-78-051. National Technical Information Service, Springfield, VA.

Chapman, P.M., M.A. Farrell and R.O. Brinkhurst. 1982a. Relative tolerance of selected aquatic oligochaetes to individual pollutants and environmental factors. *Aquat. Toxicol.* 2:47-67.

Chapman, P.M., M.A. Farrell and R.O. Brinkhurst. 1982b. Relative tolerances of selected aquatic oligochaetes to combinations of pollutants and environmental factors. *Aquat. Toxicol.* 2:69-78.

Chapman, P.M., M.A. Farrell and R.O. Brinkhurst. 1982c. Effects of species interactions on the survival and respiration of Limnodrilus hoffmeisteri and Tubifex tubifex (Oligochaeta, Tubificidae) exposed to various pollutants and environmental factors. *Water Res.* 16:1405-1408.

Chowdary, V.D., P.V. Rao and R. Narayanan. 1979. Effect of copper sulfate and sodium pentachlorophenate on adenine and adenosine phosphatases in Lymnaea luteola (Mollusca: Gastropoda). *Bull. Environ. Contam. Toxicol.* 23:615-619.

Cirelli, D.P. 1978. Patterns of pentachlorophenol usage in the United States of America - an overview. In: Pentachlorophenol: Chemistry, pharmacology, and environmental toxicology. Rao, K.R. (Ed.). Plenum Publishing Co., New York, NY. pp. 13-18.

Clemens, H.P. and K.E. Sneed. 1959. Lethal doses of several commercial chemicals for fingerling channel catfish. Special Scientific Report--Fisheries No. 316. U.S. Fish and Wildlife Service, Washington, DC.

Clendenning, K.A. and W.J. North. 1959. Effects of wastes on the giant kelp, Macrocystis pyrifera. In: International conference on waste disposal in the marine environment. First Berkeley California 1959 Proceedings, Berkeley, CA. pp. 82-91.

Cleveland, L., D.R. Buckler, F.L. Mayer and D.R. Branson. 1982. Toxicity of three preparations of pentachlorophenol to fathead minnows - a comparative study. Environ. Toxicol. Chem. 1:205-212.

Coglianesse, M.P. and J.M. Neff. 1982. Biochemical responses of the blue crab, Callinectes sapidus, to pentachlorophenol. In: Physiological mechanisms of marine pollutant toxicity. Vernberg, W.B., A. Calabrese, F.P. Thurberg and F.J. Vernberg (Eds.). Academic Press, New York, NY. pp. 127-143.

Conklin, P.J. and F.R. Fox. 1978. Environmental impacts of pentachlorophenol and its products - A round table discussion. In: Pentachlorophenol: Chemistry, pharmacology and environmental toxicology. Rao, K.R. (Ed.). Plenum Publishing Company, New York, NY. pp. 389-394.

Conklin, P.J. and K.R. Rao. 1978a. Toxicity of sodium pentachlorophenate (Na-PCP) to the grass shrimp, Palaemonetes pugio, at different stages of the molt cycle. Bull. Environ. Contam. Toxicol. 20:275-279.

Conklin, P.J. and K.R. Rao. 1978b. Toxicity of sodium pentachlorophenate to the grass shrimp, Palaemonetes pugio, in relation to the molt cycle. In: Pentachlorophenol: Chemistry, pharmacology and environmental toxicology. Rao, K.R. (Ed.). Plenum Publishing Company, New York, NY. pp. 181-192.

Cook, W.L., D. Fielder and A.W. Bourquin. 1980. Succession of microfungi in estuarine microcosms perturbed by carbaryl, methyl parathion and pentachlorophenol. Bot. Mar. 23:129-131.

Crandall, C.A. and C.J. Goodnight. 1959. The effect of various factors on the toxicity of sodium pentachlorophenate to fish. Limnol. Oceanogr. 4:53-56.

Crandall, C.A. and C.J. Goodnight. 1962. Effects of sublethal concentrations of several toxicants on growth of the common guppy, Lebistes reticulatus. Limnol. Oceanogr. 7:230-239.

Crandall, C.A. and C.J. Goodnight. 1963. The effects of sublethal concentrations of several toxicants to the common guppy, Lebistes reticulatus. Trans. Am. Microsc. Soc. 82:59-73.

Crossland, N.O. and C.J.M. Wolff. 1985. Fate and biological effects of pentachlorophenol in outdoor ponds. Environ. Toxicol. Chem. 4:73-86.

Curtis, C., A. Lima, S.J. Lozano and G.D. Veith. 1982. Evaluation of a bacterial bioluminescence bioassay as a method for predicting acute toxicity of organic chemicals to fish. In: Aquatic toxicology and hazard assessment: Fifth conference. Pearson, J.G., R.B. Foster and W.E. Bishop (Eds.). ASTM STP 766. American Society for Testing and Materials, Philadelphia, PA. pp. 170-178.

Dalela, R.C., S. Rani, S. Rani and S.R. Verma. 1980a. Influence of pH on the toxicity of phenol and its two derivatives pentachlorophenol and dinitrophenol to some fresh water teleosts. Acta Hydrochim. Hydrobiol. 8:623-629.

Dalela, R.C., S. Rani and S.R. Verma. 1980b. Physiological stress induced by sublethal concentrations of phenol and pentachlorophenol in Notopterus notopterus: Hepatic acid and alkaline phosphatases and succinic dehydrogenase. Environ. Pollut. (Series A) 21:3-8.

Dalela, R.C., S. Rani and S.R. Verma. 1980c. In vivo subacute physiological stress induced by phenolic compounds on acid and alkaline phosphatases in serum of a fish, Notopterus notopterus. Toxicol. Lett. 7:181-186.

D'Asaro, C.N. and F.G. Wilkes. 1982. Cycling of xenobiotics through marine and estuarine sediments. EPA-600/3-82-074 or PB82-239252. National Technical Information Service, Springfield, VA.

Dave, G. 1984. Effect of pH on pentachlorophenol toxicity to embryos and larvae of zebrafish (Brachydanio rerio). Bull. Environ. Contam. Toxicol. 33:621-630.

Davies, R.P. and A.J. Dobbs. 1984. The prediction of bioconcentration in fish. Water Res. 18:1253-1262.

Davis, H.C. and H. Hidu. 1969. Effects of pesticides on embryonic development of clams and oysters and on survival and growth of the larvae. Fish. Bull. 67:393-404.

Davis, J.C. and R.A.W. Hoos. 1975. Use of sodium pentachlorophenate and dehydroabiatic acid as reference toxicants for salmonid bioassays. J. Fish. Res. Board Can. 32:411-416.

DeLaune, R.D., R.P. Gambrell and K.S. Reddy. 1983. Fate of pentachlorophenol in estuarine sediment. Environ. Pollut. (Series B). 6:297-308.

DeVault, D.S. 1985. Contaminants in fish from Great Lakes harbors and tributary mouths. Arch. Environ. Contam. Toxicol. 14:587-594.

Dimick, R.E. and W.P. Breese. 1965. Bay mussel embryo bioassay. In: Proceedings of the 12th Pacific Northwest Industrial waste conference. University of Washington, Seattle, WA. pp. 165-175.

Dive, D., H. Leclerc and G. Persoone. 1980. Pesticide toxicity on the ciliate protozoan Colpidium campylum: Possible consequences of the effect of pesticides in the aquatic environment. Ecotoxicol. Environ. Safety 4:129-133.

Dixon, W.J. and M.B. Brown (Eds.). 1979. BMDP Biomedical Computer Programs, P-series. University of California, Berkeley, CA. p. 521.

Dominguez, S.E. and G.A. Chapman. 1984. Effect of pentachlorophenol on the growth and mortality of embryonic and juvenile steelhead trout. Arch. Environ. Contam. Toxicol. 13:739-743.

Dougherty, R.C. 1978. Human exposure to pentachlorophenol. In: Pentachlorophenol: Chemistry, pharmacology, and environmental toxicology. Rao, K.R. (Ed.). Plenum Publishing Co., New York, NY. pp. 351-361.

Dougherty, R.C., M.J. Whitaker, L.M. Smith, D.L. Stalling and D.W. Kuehl. 1980. Negative chemical ionization studies of human and food chain contamination with xenobiotic chemicals. Environ. Health Perspect. 36:103-118.

Doughtie, D.G. and K.R. Rao. 1978. Ultrastructural changes induced by sodium pentachlorophenate in the grass shrimp, Palaemonetes pugio, in relation to the molt cycle. In: Pentachlorophenol: Chemistry, pharmacology and environmental toxicology. Rao, K.R. (Ed.). Plenum Publishing Company, New York, NY. pp. 213-249.

Elnabarawy, M., A.N. Welter and R.R. Robideau. 1986. Relative sensitivity of three daphnid species to selected organic and inorganic chemicals. Environ. Toxicol. Chem. 5:393-398.

Erickson, S.J. 1981. Manuscript. Inhibition of photosynthesis in estuarine phytoplankton by mixtures of copper and pentachlorophenol. U.S. EPA, Gulf Breeze, FL.

Erickson, S.J. and A.E. Freeman. 1978. Toxicity screening of fifteen chlorinated and brominated compounds using four species of marine phytoplankton. In: Water chlorination: Environmental impact and health effects. Vol. 2. Jolley, R.L., H. Gorchev and D.H. Hamilton (Eds.). Ann Arbor Science Publishers, Ann Arbor, MI. pp. 307-310.

Ernst, W. 1979. Factors affecting the evaluation of chemicals in laboratory experiments using marine organisms. Ecotoxicol. Environ. Safety 3:90-93.

Faas, L.F. and J.C. Moore. 1979. Determination of pentachlorophenol in marine biota and sea water by gas-liquid chromatography and high-pressure liquid chromatography. J. Agric. Food Chem. 27:554-557.

Fairchild, J.F., T.P. Boyle, E. Robinson-Wilson and J.R. Jones. 1984. Effects of inorganic nutrients on microbial leaf decomposition and mitigation of chemical perturbation. *J. Freshwater Ecol.* 2:405-416.

Fisher, S.W. 1986. Effects of temperature on the acute toxicity of PCP in the midge Chironomus riparius Meigen. *Bull. Environ. Contam. Toxicol.* 36:744-748.

Fisher, S.W. and R.W. Wadleigh. 1986. Effects of pH on the acute toxicity and uptake of [¹⁴C] pentachlorophenol in the midge, Chironomus riparius. *Ecotoxicol. Environ. Safety* 11:1-8.

Fogels, A. and J.B. Sprague. 1977. Comparative short-term tolerance of zebrafish, flagfish, and rainbow trout to five poisons including potential reference toxicants. *Water Res.* 11:811-817.

Folke, J., J. Birklund, A.K. Sorensen and U. Lund. 1983. The impact on the ecology of polychlorinated phenols and other organics dumped at the bank of a small marine inlet. *Chemosphere* 12:1169-1181.

Fountaine, J.E., P.B. Joshipura and P.N. Keliher. 1976. Some observations regarding pentachlorophenol levels in Haverford township, Pennsylvania. *Water Res.* 10:185-188.

Fox, F.R. and K.R. Rao. 1978. Effects of sodium pentachlorophenolate and 2,4-dinitrophenol on hepatopancreatic enzymes in the blue crab, Callinectes sapidus. In: Pentachlorophenol: Chemistry, pharmacology and environmental toxicology. Rao, K.R. (Ed.). Plenum Publishing Company. New York, NY. pp. 265-275.

Fox, M.E. and S.R. Joshi. 1984. The fate of pentachlorophenol in the Bay of Quinte, Lake Ontario. J. Great Lakes Res. 10:190-196.

Freitag, D., L. Ballhorn, H. Geyer and F. Korte. 1985. Environmental hazard profile of organic chemicals. An experimental method for the assessment of the behavior of organic chemicals in the ecosphere by means of simple laboratory tests with ¹⁴C labelled chemicals. Chemosphere 14:1589-1616.

Geyer, H., R. Viswanathan, D. Freitag and F. Korte. 1981. Relationship between water solubility of organic chemicals and their bioaccumulation by the alga Chlorella. Chemosphere 10:1307-1313.

Geyer, H., G. Politzki and D. Freitag. 1984. Prediction of ecotoxicological behavior of chemicals: Relationship between n-octanol/water partition coefficient and bioaccumulation of organic chemicals by alga Chlorella. Chemosphere 13:269-284.

Geyer, H., I. Scheunert and F. Korte. 1985. The effects of organic environmental chemicals on the growth of the alga Scenedesmus subspicatus: A contribution to environmental biology. Chemosphere 14:1355-1369.

Glickman, A.H., C.N. Statham, A. Wu and J.J. Lech. 1977. Studies on the uptake, metabolism, and disposition of pentachlorophenol and pentachloroanisole in rainbow trout. *Toxicol. Appl. Pharmacol.* 41:649-658.

Gluth, G., D. Freitag, W. Hanke and F. Korte. 1985. Accumulation of pollutants in fish. *Comp. Biochem. Physiol.* 81C:273-277.

Goel, H.C. and R. Prasad. 1978. Action of molluscicides on freshly laid eggs of the snail Indoplanorbis exustus (Deshayes). *Indian J. Exp. Biol.* 16:620-622.

Goodnight, C.J. 1942. Toxicity of sodium pentachlorophenate and pentachlorophenol to fish. *Ind. Eng. Chem.* 34:868-872.

Gotham, I.J. and G.Y. Rhee. 1982. Effects of hexachlorobiphenyl and pentachlorophenol on growth and photosynthesis of phytoplankton. *J. Great Lakes Res.* 8:328-335.

Gupta, P.K. 1983. Acute toxicity of pentachlorophenol to a freshwater teleost, Rasbora daniconius neilgeriensis (Hamilton). *Arch. Hydrobiol.* 93:127-132.

Gupta, P.K. and V.S. Durve. 1984a. Evaluation of the toxicity of sodium pentachlorophenate, pentachlorophenol and phenol to the snail Viviparus bengalensis (L.). *Arch. Hydrobiol.* 101:469-475.

Gupta, P.K. and V.S. Durve. 1984b. A study on the effect of temperature upon the toxicity of sodium pentachlorophenate to the freshwater snail Viviparus bengalensis L. Acta Hydrochim. Hydrobiol. 12:369-375.

Gupta, P.K. and P.S. Rao. 1982. Toxicity of phenol, pentachlorophenol and sodium pentachlorophenate to a freshwater pulmonate snail Lymnaea acuminata (Lamarck). Arch. Hydrobiol. 94:210-217.

Gupta, P.K., V.S. Mujumdas, P.S. Rao and V.S. Durve. 1982a. Toxicity of phenol, pentachlorophenol, and sodium pentachlorophenolate to a freshwater teleost Lebistes reticulatus (Peters). Acta Hydrochim. Hydrobiol. 10:177-181.

Gupta, S. and R.C. Dalela. 1986. Liver damage to Notopterus notopterus following exposure to phenolic compounds. J. Environ. Biol. 7:75-80.

Gupta, S., S.R. Verma and P.K. Saxena. 1982b. Toxicity of phenolic compounds in relation to the size of a freshwater fish, Notopterus notopterus (Pallas). Ecotoxicol. Environ. Safety 6:433-438.

Gupta, S., R.C. Dalela and P.K. Saxena. 1983a. Effect of phenolic compounds on in vivo activity of transaminases in certain tissues of the fish, Notopterus notopterus. Environ. Res. 32:8-13.

Gupta, S., R.C. Dalela and P.K. Saxena. 1983b. Effects of phenolic compounds on 5-nucleotidase activity in some tissues of Notopterus notopterus (Pallas): A biochemical study. Toxicol. Lett. 17:167-173.

Gupta, S., R.C. Dalela and P.K. Saxena. 1983c. Influence of dissolved oxygen levels on acute toxicity of phenolic compounds to fresh water teleost Notopterus notopterus (Pallas). Water Air Soil Pollut. 19:223-228.

Gupta, S., R.C. Dalela and P.K. Saxena. 1983d. Influence of temperature on the toxicity of phenol and its chloro- and nitro-derivatives to the fish Notopterus notopterus (Pallas). Acta Hydrochim. Hydrobiol. 11:187-192.

Halbach, U. 1984. Population dynamics of rotifers and its consequences for ecotoxicology. Hydrobiologia 109:79-96.

Halbach, U., M. Siebert, M. Westermayer and C. Wissel. 1983. Population ecology of rotifers as a bioassay tool for ecotoxicological tests in aquatic environments. Ecotoxicol. Environ. Safety 7: 484-513. :

Hall, L.H. and L.B. Kier. 1984a. Molecular connectivity of phenols and their toxicity to fish. Bull. Environ. Contam. Toxicol. 32:354-362.

Hall, L.H. and L.B. Kier. 1984b. A molecular connectivity study of phenols and their toxicity to fish. In: QSAR in Design of Bioactive Compounds: Proceedings of the 1st Telesymposium on Medicinal Chemistry. Kuchar, M. (Ed.). J.R. Prous Publishing Co., Barcelona, Spain. pp. 53-59.

Hall, L.H., L.B. Kier and G. Phipps. 1984. Structure-activity relationship studies on the toxicities of benzene derivatives: I. An additivity model. *Environ. Toxicol. Chem.* 3:355-365.

Hall, W.S., R.L. Paulson, L.W. Hall, Jr. and D.T. Burton. 1986. Acute toxicity of cadmium and sodium pentachlorophenate to daphnids and fish. *Bull. Environ. Contam. Toxicol.* 37:308-316.

Hallas, T. 1973. On the accumulation and metabolization of pentachlorophenol in fish. *Medd. Dan. Fisk. Havunders.* 7:75-84.

Hamilton, S.J., L. Cleveland, L.M. Smith, J.A. Lebo and F.L. Mayer. 1986. Toxicity of pure pentachlorophenol and chlorinated phenoxyphenol impurities to fathead minnows. *Environ. Toxicol. Chem.* 5:543-552.

Hanke, W., G. Gluth, H. Bubel and R. Muller. 1983. Physiological changes in carps induced by pollution. *Ecotoxicol. Environ. Safety* 7:229-241.

Hansen, D.J. and M.E. Tagatz. 1980. A laboratory test for assessing impacts of substances on developing communities of benthic estuarine organisms. In: *Aquatic toxicology*. Eaton, J.G., P.R. Parrish and A.C. Hendricks (Eds.). ASTM STP 707. American Society for Testing and Materials, Philadelphia, PA. pp. 40-57.

Hanumante, M.M. and S.S. Kulkarni. 1979. Acute toxicity of two molluscicides, mercuric chloride and pentachlorophenol to a freshwater fish (Channa gachua). Bull. Environ. Contam. Toxicol. 23:725-727.

Hashimoto, Y. and Y. Nishiuchi. 1983. Effects of herbicides on aquatic animals. In: Pesticide chemistry: Human welfare and the environment. Vol. 2. Takahashi, N., H. Yoshioka, T. Misato, and S. Matsunaka (Eds.). Pergamon Press, New York, NY. pp. 355-358.

Hashimoto, Y., E. Okubo, T. Ito, M. Yamaguchi and S. Tanaka. 1982. Changes in susceptibility of carp to several pesticides with growth. J. Pestic. Sci. 7:457-461.

Hattori, M., K. Senoo, S. Harada, Y. Ishizu and M. Goto. 1984. The Daphnia reproduction test of some environmental chemicals. Seitai Kagaki 6:23-27.

Hattula, M.L., V. Wasenius, H. Reunanen and A.U. Arstila. 1981. Acute toxicity of some chlorinated phenols, catechols, and cresols to trout. Bull. Environ. Contam. Toxicol. 26:295-298.

Hauch, R.G., D.R. Norris and R.H. Pierce, Jr. 1980. Acute and chronic toxicity of sodium pentachlorophenate to the copepod, Pseudodiaptomus coronatus. Bull. Environ. Contam. Toxicol. 25:562-568.

Hedtke, S.F. and J.W. Arthur. 1985. Evaluation of a site-specific water quality criterion for pentachlorophenol using outdoor experimental streams. In: Aquatic toxicology and hazard assessment: Seventh symposium. Cardwell, R.D., R. Purdy and R.C. Bahner (Eds.). ASTM STP 854. American Society for Testing and Materials, Philadelphia, PA. pp. 551-564.

Hedtke, S.F., C.W. West, K.N. Allen, T.J. Norberg-King and D.I. Mount. 1986. Toxicity of pentachlorophenol to aquatic organisms under naturally varying and controlled environmental conditions. Environ. Toxicol. Chem. 5:531-542.

Hermens, J., H. Canton, N. Steyger and R. Wegman. 1984. Joint effects of a mixture of 14 chemicals on mortality and inhibition of reproduction of Daphnia magna. Aquat. Toxicol. 5:315-322.

Hermens, J., P. Leeuwangh and A. Musch. 1985. Joint toxicity of mixtures of groups of organic aquatic pollutants to the guppy (Poecilia reticulata). Ecotoxicol. Environ. Safety 9:321-326.

Hodson, P.V. and B.R. Blunt. 1981. Temperature-induced changes in pentachlorophenol chronic toxicity to early life stages of rainbow trout. Aquat. Toxicol. 1:113-127.

Hodson, P.V., D.G. Dixon and K.L.E. Kaiser. 1984. Measurement of median lethal dose as a rapid indication of contaminant toxicity to fish. Environ. Toxicol. Chem. 3:243-254.

Holcombe, G.W., G.L. Phipps and J.T. Fiantdt. 1982. Effects of phenol, 2,4-dimethylphenol, 2,4-dichlorophenol, and pentachlorophenol on embryo, larval, and early-juvenile fathead minnows (Pimephales promelas). Arch. Environ. Contam. Toxicol. 11:73-78.

Holmberg, B., S. Jensen, A. Larsson, K. Lewander and M. Olsson. 1972. Metabolic effects of technical pentachlorophenol (PCP) on the eel Anguilla anguilla L. Comp. Biochem. Physiol. 43B:171-183.

Hooftman, R.N. and G.J. Vink. 1980. The determination of toxic effects of pollutants with the marine polychaete worm Uphryotrocha diadema. Ecotoxicol. Environ. Safety 4:252-262.

Huang, J. and E.F. Gloyna. 1967. Effects of toxic organics on photosynthetic reoxygenation. PB 216749. National Technical Information Service, Springfield, VA.

Huber, W., V. Schubert and C. Sautter. 1982. Effects of pentachlorophenol on the metabolism of the aquatic macrophyte Lemna minor L. Environ. Pollut. (Series A) 29:215-223.

Mckins, J.N. and J.D. Petty. 1983. Dynamics of purified and industrial pentachlorophenol in fathead minnows. Arch. Environ. Contam. Toxicol. 12:667-672.

Inglis, A. and E.L. Davis. 1972. Effects of water hardness on the toxicity of several organic and inorganic herbicides to fish. Technical Paper No. 67. U.S. Bureau of Sport Fisheries and Wildlife, Washington, DC.

Ishak, M.M., A.A. Sharaf, A.M. Mohamed and A.H. Mousa. 1970. Studies on the mode of action of some molluscicides on the snail, Biomphalaria alexandrina. I. Effect of Bayluscide, sodium pentachlorophenate, and copper sulphate on succinate, glutamate, and reduced TMPD oxidation. Comp. Gen. Pharmacol. 1:201-208.

Iwama, G.K. and G.L. Greer. 1979. Toxicity of sodium pentachlorophenate to juvenile salmon under conditions of high feeding density and continuous-flow exposure. Bull. Environ. Contam. Toxicol. 23:711-716.

Iwama, G.K. and G.L. Greer. 1980. Effect of a bacterial infection on the toxicity of sodium pentachlorophenate to juvenile coho salmon. Trans. Am. Fish. Soc. 109:290-292.

Iwama, G.K. and G.L. Greer. 1982. Mortality in juvenile chinook salmon exposed to sodium pentachlorophenate and undergoing progressively symptomatic bacterial kidney disease. Canadian Technical Report of Fisheries and Aquatic Sciences No. 1100. Department of Fisheries and Oceans, West Vancouver Lab, West Vancouver, British Columbia, Canada.

Jayaweera, R., R. Peterson and P. Smejtek. 1982. Induced hydrogen ion transport in lipid membranes as origin of toxic effect of pentachlorophenol in an alga. *Pestic. Biochem. Physiol.* 18:197-204.

Johansen, P.H., R.A.S. Mathers, J.A. Brown and P.W. Colgan. 1985. Mortality of early life stages of largemouth bass, Micropterus salmoides, due to pentachlorophenol exposure. *Bull. Environ. Contam. Toxicol.* 34:377-384.

Johnson, W.W. and M.T. Finley. 1980. Handbook of acute toxicity of chemicals to fish and aquatic invertebrates. Resource Publication 137. U.S. Fish and Wildlife Service, Washington, DC. pp. 58.

Kaila, K. 1982. Increase in the resting potassium permeability of crayfish axons by pentachlorophenol and trinitrophenol in the absence of extracellular Ca^{2+} . *Comp. Biochem. Physiol.* 73C:353-356.

Kaila, K. and J. Saarikoski. 1977. Toxicity of pentachlorophenol and 2,3,6-trichlorophenol to the crayfish (Astacus fluviatilis L.). *Environ. Pollut.* 12:119-123.

Kaila, K. and J. Saarikoski. 1980. Inhibition of voltage-dependent potassium conductance by convulsant phenols in the medial giant axon of the crayfish. *Comp. Biochem. Physiol.* 65C:17-24.

Kaila, K. and J. Saarikoski. 1981. Membrane-potential changes caused by 2,4-DNP and related phenols in resting crayfish axons are not due to uncoupling of mitochondria. *Comp. Biochem. Physiol.* 69C:235-242.

Khangarot, B.S. 1983. Acute toxicity of pentachlorophenol and antimycin to common guppy (Lebistes reticulatus Peters). Indian J. Phys. Nat. Sci. 3:25-29.

Khangarot, B.S., A. Sehgal and M.K. Bhasin. 1985. "Man and biosphere" - studies on the Sikkim Himalayas. Part 6: Toxicity of selected pesticides to frog tadpole Rana hexadactyla (Lesson). Acta Hydrochim. Hydrobiol. 13:391-394.

Kishino, T. and K. Kobayashi. 1980. Studies on the metabolism of chlorophenols in fish - XIV. A study on the absorption mechanism of pentachlorophenol in goldfish relating to its distribution between solvents and water. Bull. Jpn. Soc. Sci. Fish. 46:1165-1168.

Klein, W., H. Geyer, D. Freitag and H. Rohleder. 1984. Sensitivity of schemes for ecotoxicological hazard ranking of chemicals. Chemosphere 13:203-211.

Knie, J., A. Halke, I. Juhnke and W. Schiller. 1983. Results of studies on chemical substances with four biotests. Deutsche Gewässerkd. Mitt. 27:77-79.

Knowlton, M.F. and J.N. Huckins. 1983. Fate of radiolabeled sodium pentachlorophenate in littoral microcosms. Bull. Environ. Toxicol. Chem. 30:206-213.

Kobayashi, K. 1978. Metabolism of pentachlorophenol in fishes. In: Pentachlorophenol: Chemistry, pharmacology and environmental toxicology. Rao, K.R. (Ed.). Plenum Publishing Company, New York, NY. pp. 89-105.

Kobayashi, K. 1979. Metabolism of pentachlorophenol in fish. In: Pesticide and xenobiotic metabolism in aquatic organisms. Khan, M.A.Q., J.J. Lech and J.J. Menn (Eds.). ACS Symposium Series No. 99. American Chemical Society, Washington, DC. pp. 131-143.

Kobayashi, K. and H. Akitake. 1975a. Studies on the metabolism of chlorophenols in fish - I. Absorption and excretion of PCP by goldfish. Bull. Jpn. Soc. Sci. Fish. 41:87-92.

Kobayashi, K. and H. Akitake. 1975b. Studies on the metabolism of chlorophenols in fish - II. Turnover of absorbed PCP in goldfish. Bull. Jpn. Soc. Sci. Fish. 41:93-99.

Kobayashi, K. and T. Kishino. 1980. Effect of pH on the toxicity and accumulation of pentachlorophenol in goldfish. Bull. Jpn. Soc. Sci. Fish. 46:167-170.

Kobayashi, K. and N. Nakamura. 1979a. Studies on the metabolism of chlorophenols in fish - XI. Isolation and identification of a conjugated-PCP in the urine of goldfish. Bull. Jpn. Soc. Sci. Fish. 45:1001-1003.

Kobayashi, K. and N. Nakamura. 1979b. Studies on the metabolism of chlorophenols in fish - XII. Major detoxification pathways for pentachlorophenol in goldfish. Bull. Jpn. Soc. Sci. Fish. 45:1185-1188.

Kobayashi, K., H. Akitake and T. Tomiyama. 1969. Studies on the metabolism of pentachlorophenol, a herbicide, in aquatic organisms - I. Turnover of absorbed PCP in Tapes philippinarum. Bull. Jpn. Soc. Sci. Fish. 35:1179-1183.

Kobayashi, K., S. Kimura and E. Shimizu. 1977. Studies on the metabolism of chlorophenols in fish - IX. Isolation and identification of pentachlorophenyl-glucuronide accumulated in bile of goldfish. Bull. Jpn. Soc. Sci. Fish. 45:601-607.

Kobayashi, K., H. Akitake and K. Manabe. 1979. Relation between toxicity and accumulation of various chlorophenols in goldfish. Bull. Jpn. Soc. Sci. Fish. 45:173-175.

Kobayashi, K., S. Kimura and Y. Oshima. 1984. Sulfate conjugation of various phenols by liver-soluble fraction of goldfish. Bull. Jpn. Soc. Sci. Fish. 50:833-837.

Koch, R. 1982. Molecular connectivity and acute toxicity of environmental pollutants. Chemosphere 11:925-931.

Konemann, H. and A. Musch. 1981. Quantitative structure-activity relationships in fish toxicity studies. Part 2: The influence of pH on the QSAR of chlorophenols. *Toxicology* 19:223-228.

Korte, F., D. Freitag, H. Geyer, W. Klein, A.G. Kraus and E. Lohaniatus. 1978. Ecotoxicologic profile analysis: A concept for establishing ecotoxicologic priority lists for chemicals. *Chemosphere* 1:79-102.

Kozak, V.P., G.V. Simsiman, G. Chester, D. Stensley and J. Harkin. 1979. Reviews of environmental effects of pollutants. XI. Chlorophenols. EPA 600/1-79-012. National Technical Information Service, Springfield, VA.

Kuehl, D.W. and R.C. Dougherty. 1980. Pentachlorophenol in the environment. Evidence for its origin from commercial pentachlorophenol by negative chemical ionization mass spectrometry. *Environ. Sci. Technol.* 14:447-449.

Kuehl, D.W., E.N. Leonard, K.J. Welch and G.D. Veith. 1980. Identification of hazardous organic chemicals in fish from the Ashtabula River, Ohio, and Wabash River, Indiana. *J. Assoc. Off. Anal. Chem.* 63:1238-1244.

Kuehl, D.W., E.N. Leonard, B.C. Butterworth and K.L. Johnson. 1983. Polychlorinated chemical residues in fish from major watersheds near the Great Lakes. *Environ. Int.* 9:293-299.

Kwasniewska, K., D. Liu and W.M.J. Strachan. 1979. Bioassay of relative toxicity of some chlorophenols using sewage bacteria. Fisheries and Marine Service Technical Report No. 862. Fisheries and Marine Service, Pacific Environment Institute, West Vancouver, British Columbia, Canada. pp. 193-200.

LeBlanc, G.A. 1980. Acute toxicity of priority pollutants to water flea (Daphnia magna). Bull. Environ. Contam. Toxicol. 24:684-691.

LeBlanc, G.A. and B.J. Cochrane. 1985. Modulation of substrate-specific glutathione S-transferase activity in Daphnia magna with concomitant effects on toxicity tolerance. Comp. Biochem. Physiol. 82C:37-42.

Lech, J.J., A.H. Glickman and C.N. Slatham. 1978. Studies on the uptake, disposition and metabolism of pentachlorophenol and pentachloroanisole in rainbow trout. In: Pentachlorophenol: Chemistry, pharmacology, and environmental toxicology. Rao, K.R. (Ed.). Plenum Publishing Company, New York, NY. pp. 107-113.

Lemma, A. and P. Yau. 1974. Studies on the molluscicidal properties of endod (Phytolacca dodecandra): II. Comparative toxicity of various molluscicides to fish and snails. Ethiopian Med. J. 12:109-114.

Lewis, P.A. and C.I. Weber. 1985. A study of the reliability of Daphnia acute toxicity tests. In: Aquatic toxicology and hazard assessment: Seventh symposium. Cardwell, R.D., R. Purdy and R.C. Bahner (Eds.). ASTM STP 854. American Society for Testing and Materials, Philadelphia, PA. pp. 73-86.

Lipnick, R.L., C.K. Bickings, D.E. Johnson and D.A. Eastmond. 1986. Comparison of QSAR predictions with fish toxicity screening data for 110 phenols. In: Aquatic toxicology and hazard assessment: Eighth symposium. R.C. Bahner and D.J. Hansen (Eds.). ASTM STP 891. American Society for Testing and Materials, Philadelphia, PA. pp. 153-176.

Liu, D.H.W. 1981. A radiorespirometric study of the effects of a chemical substance on carbohydrate metabolism in the intact fish. In: Aquatic toxicology and hazard assessment: Fourth conference. Branson, D.R. and K.L. Dickson (Eds.). ASTM STP 737. American Society for Testing and Materials, Philadelphia, PA. pp. 449-458.

Lu, P.Y. and R.L. Metcalf. 1975. Environmental fate and biodegradability of benzene derivatives as studied in a model aquatic ecosystem. Environ. Health Perspect. 10:269-284.

Mathers, R.A., J.A. Brown and P.H. Johansen. 1985. The growth and feeding behaviour responses of largemouth bass (Micropterus salmoides) exposed to PCP. Aquat. Toxicol. 6:157-164.

Matida, Y., S. Kimura, H. Tanaka and M. Yokote. 1976. Effects of some herbicides applied in the forest to the freshwater fishes and other aquatic organisms. III. Experiments on the assessment of acute toxicity of herbicides to aquatic organisms. Bull. Freshwater Fish. Res. Lab. (Tokyo) [Eng. Transl. of Tansuika Suisan Kenkyusho Kenkyu Hokoku] 26:79-84.

Mattson, V.R., J.W. Arthur and C.T. Walbridge. 1976. Acute toxicity of selected organic compounds to fathead minnows. EPA-600/3-76-097. National Technical Information Service, Springfield, VA.

McCarty, L.S., P.V. Hodson, G.R. Craig and K.L.E. Kaiser. 1985. The use of quantitative structure-activity relationships to predict the acute and chronic toxicities of organic chemicals to fish. Environ. Toxicol. Chem. 4:595-606.

McCoy, L.E. and J.E. Joy. 1977. Tolerance of Sepedon fuscipennis and Dictya sp. larvae (Diptera: Sciomyzidae) to the molluscicides Bayer 73 and sodium pentachlorophenate. Environ. Entomol. 6:198-202.

McKim, J., P. Schmieder and G. Veith. 1985. Absorption dynamics of organic chemical transport across trout gills as related to octanol-water partition coefficient. Toxicol. Appl. Pharmacol. 77:1-10.

McKim, J.M., P.K. Schmieder and R.J. Erickson. Manuscript. The general pharmacokinetics of [¹⁴C]-pentachlorophenol in the rainbow trout (Salmo gairdneri). U.S. EPA, Duluth, MN.

McLeese, D.W., V. Zitko and M.R. Peterson. 1979. Structure-lethality relationships for phenols, anilines and other aromatic compounds in shrimp and clams. *Chemosphere* 8:53-57.

Metcalf, J.L., M.E. Fox and J.H. Casey. 1984. Aquatic leeches (Hirudinea) as bioindicators of organic chemical contaminants in freshwater ecosystems. *Chemosphere* 13:143-150.

Morgan, W.S. 1976. Fishing for toxicity: Biological automonitor for continuous water quality control. *Effluent Water Treat. J.* 16:471-475.

Morgan, W.S.G. 1977. Biomonitoring with fish: An aid to industrial effluent and surface water quality control. *Prog. Water Technol.* 9:703-711.

Mount, D.I. and T.J. Norberg. 1984. A seven-day life-cycle cladoceran toxicity test. *Environ. Toxicol. Chem.* 3:425-434.

Mrak, E.M. 1974. Herbicide Report - Chemistry and analysis, environmental effects, agricultural and other applied uses. EPA-SAB-74-001. National Technical Information Service, Springfield, VA.

Murray, H.E., G.S. Neff, Y. Hrunng and C.S. Giam. 1980. Determination of benzo(a)pyrene, hexachlorobenzene and pentachlorophenol in oysters from Galveston Bay, Texas. *Bull. Environ. Contam. Toxicol.* 25:663-667.

- Murray, H.E., L.E. Ray and C.S. Giam. 1981. Analysis of marine sediment, water and biota for selected organic pollutants. *Chemosphere* 10:1327-1334.
- Nagendran, R. and K. Shakuntala. 1979. Studies on toxicity of biocides to cyprinid forage fishes: Part I - Effects of sublethal concentrations of sodium pentachlorophenate on the ecophysiology of Puntius ticto (Ham). *Indian J. Exp. Biol.* 17:270-273.
- National Research Council of Canada. 1982. Chlorinated phenols: Criteria for environmental quality. NRCC No. 18578. Environmental Secretariat, Ottawa, Canada.
- Neter, J. and W. Wasserman. 1974. *Applied Linear Statistical Models*. Irwin, Inc., Homewood, IL.
- Niimi, A.J. and C.Y. Cho. 1983. Laboratory and field analysis of pentachlorophenol (PCP) accumulation by salmonids. *Water Res.* 17:1791-1795.
- Niimi, A.J. and C.A. McFadden. 1982. Uptake of sodium pentachlorophenate (NaPCP) from water by rainbow trout (Salmo gairdneri) exposed to concentrations in the ng/l range. *Bull. Environ. Contam. Toxicol.* 28:11-19.
- Niimi, A.J. and V. Palazzo. 1985. Temperature effect on the elimination of pentachlorophenol, hexachlorobenzene and mirex by rainbow trout (Salmo gairdneri). *Water Res.* 19:205-207.

Nilsson, C.A., A. Norstrom, K. Andersson and C. Rappe. 1978. Impurities in commercial products related to pentachlorophenol. In: Pentachlorophenol: Chemistry, pharmacology, and environmental toxicology. Rao, K.R. (Ed.). Plenum Publishing Company, New York, NY. pp. 313-324.

Norup, B. 1972. Toxicity of chemicals in paper factory effluents. Water Res. 6:1585-1588.

Oikari, A.O.J. 1986. Metabolites of xenobiotics in the bile of fish in waterways polluted by pulp mill effluents. Bull. Environ. Contam. Toxicol. 36:429-436.

Oikari, A. and E. Anas. 1985. Chlorinated phenols and their conjugates in the bile of trout (Salmo gairdneri) exposed to contaminated waters. Bull. Environ. Contam. Toxicol. 35:802-809.

Oikari, A.O.J. and J. Niittyala. 1985. Subacute physiological effects of bleached kraft mill effluent (BKME) on the liver of trout, Salmo gairdneri. Ecotoxicol. Environ. Safety 10:159-172.

Oikari, A., B. Holmbom, E. Anas, M. Miilunpalo, G. Kruzynski and M. Castren. 1985. Ecotoxicological aspects of pulp and paper mill effluents discharged to inland water systems: Distribution in water and toxicant residues and physiological effects in caged fish (Salmo gairdneri). Aquat. Toxicol. 6:219-239.

Owen, J.W. and S.W. Rosso. 1981. Effects of sublethal concentrations of pentachlorophenol on the liver of bluegill sunfish, Lepomis macrochirus. Bull. Environ. Contam. Toxicol. 26:594-600.

Paasivirta, J., J. Sarkka, T. Leskijarvi and A. Roos. 1980. Transportation and enrichment of chlorinated phenolic compounds in different aquatic food chains. Chemosphere 9:441-456.

Paasivirta, J., J. Sarkka, M. Aho, K. Surma-Aho, J. Tarkanen and A. Roos. 1981. Recent trends of biocides in pikes of the Lake Paiganne. Chemosphere 10:405-414.

Paasivirta, J., J. Sarkka, K. Surma-Aho, T. Humpi, I. Kuokkanen and M. Marttinen. 1983. Food chain enrichment of organochlorine compounds and mercury in clean and polluted lakes of Finland. Chemosphere 12:239-252.

Palmer, C.M. and T.E. Maloney. 1955. Preliminary screening for potential algicides. Ohio J. Sci. 55:1-8.

Parrish, P.R., E.E. Dyar, J.M. Enos and W.G. Wilson. 1978. Chronic toxicity of chlordane, trifluralin, and pentachlorophenol to sheepshead minnows (Cyprinodon variegatus). EPA-600/3-78-010. National Technical Information Service, Springfield, VA.

Peer, M.M., J. Nirmala and M.N. Kutty. 1983. Effects of pentachlorophenol (NaPCP) on survival, activity and metabolism in Rhinomugil corsula (Hamilton), Cyprinus carpio (Linnaeus) and Tilapia mossambica (Peters). *Hydrobiologia* 107:19-24.

Peterson, R.H. 1976. Temperature selection of juvenile Atlantic salmon (Salmo salar) as influenced by various toxic substances. *J. Fish. Res. Board Can.* 33:1722-1730.

Phipps, G.L. and G.W. Holcombe. 1985. A method for aquatic multiple species toxicant testing: Acute toxicity of 10 chemicals to 5 vertebrates and 2 invertebrates. *Environ. Pollut. (Series A)* 38:141-157.

Phipps, G.L., G.W. Holcombe and J.T. Fiandt. 1981. Acute toxicity of phenol and substituted phenols to the fathead minnow. *Bull. Environ. Contam. Toxicol.* 26:585-593.

Pierce, R.H., Jr. 1978. Fate and impact of pentachlorophenol in a freshwater ecosystem. EPA-600/3-78-063. National Technical Information Service, Springfield, VA.

Pierce, R.H., Jr., C.R. Brent, H.P. Williams and S.G. Reeves. 1977. Pentachlorophenol distribution in a fresh water ecosystem. *Bull. Environ. Contam. Toxicol.* 18:251-258.

Pruitt, G.W., B.J. Grantham and R.H. Pierce, Jr. 1977. Accumulation and elimination of pentachlorophenol by the bluegill, Lepomis macrochirus. Trans. Am. Fish. Soc. 106:462-465.

Rao, K.R. and D.G. Doughtie. 1984. Histopathological changes in grass shrimp exposed to chromium, pentachlorophenol and dithiocarbamates. Mar. Environ. Res. 14:371-395.

Rao, K.R., P.J. Conklin and A.C. Brannon. 1978. Inhibition of limb regeneration in the grass shrimp, Palaemonetes pugio, by sodium pentachlorophenol. In: Pentachlorophenol: Chemistry, pharmacology and environmental toxicology. Rao, K.R. (Ed.). Plenum Publishing Company, New York, NY. pp. 193-203.

Rao, K.R., F.R. Fox, P.J. Conklin, A.C. Cantelmo and A.C. Brannon. 1979. Physiological and biochemical investigations of the toxicity of pentachlorophenol to crustaceans. In: Marine pollution: Functional responses. Vernberg, W.B., F.P. Thurberg, A. Calabrese and F.J. Vernberg (Eds.). Academic Press, New York, NY.

Rao, K.R., F.R. Fox, P.J. Conklin and A.C. Cantelmo. 1981. Comparative toxicology and pharmacology of chlorophenols: Studies on the grass shrimp, Palaemonetes pugio. In: Biological monitoring of marine pollutants. Vernberg, J., A. Calabrese, F.P. Thurberg and W.B. Vernberg (Eds.). Academic Press, New York, NY. pp. 37-72.

Rao, P.S., V.S. Durve, B.S. Khangarot and S.S. Shekhawat. 1983. Acute toxicity of phenol, pentachlorophenol and sodium pentachlorophenate to a freshwater ostracod Cypris subglobosa (Sowerby). Acta Hydrochim. Hydrobiol. 11:457-465.

Ray, L.E., H.E. Murray and C.S. Giam. 1983. Organic pollutants in marine samples from Portland, Maine. Chemosphere 12:1031-1038.

Renberg, L., E. Masell, G. Sundstrom and M. Adolfsson-Erici. 1983. Levels of chlorophenols in natural waters and fish after an accidental discharge of a wood-impregnating solution. Ambio 12:121-123.

Ribo, J.M. and K.L.E. Kaiser. 1983. Effects of selected chemicals to photoluminescent bacteria and their correlation with acute and sublethal effects on other organisms. Chemosphere 12:1421-1442.

Richter, J.E. 1982. Results of algal toxicity tests with priority pollutants. University of Wisconsin-Superior, Superior, WI. (Memorandum to Charles Stephan, U.S. EPA, Duluth, MN. June 30.)

Robinson-Wilson, E.F., T.P. Boyle and J.D. Petty. 1983. Effects of increasing levels of primary production on pentachlorophenol residues in experimental pond ecosystems. In: Aquatic toxicology and hazard assessment: Sixth symposium. Bishop, W.E., R.D. Cardwell, and B.B. Heidolph (Eds.). ASTM STP 802. American Society for Testing and Materials, Philadelphia, PA. pp. 239-251.

Rubinstein, N.I. 1978. Effect of sodium pentachlorophenol on the feeding activity of the lugworm, Arenicola cristata Stimpson. In: Pentachlorophenol: Chemistry, pharmacology and environmental toxicology. Rao, K.R. (Ed.). Plenum Publishing Company, New York, NY. pp. 175-179.

Rubinstein, N. 1981. U.S. EPA, Gulf Breeze, FL. (Memorandum to S. Tagatz, U.S. EPA, Gulf Breeze, FL.)

Ruesink, R.G. and L.L. Smith, Jr. 1975. The relationship of the 96-hour LC50 to the lethal threshold concentration of hexavalent chromium, phenol, and sodium pentachlorophenate for fathead minnows (Pimephales promelas Rafinesque). Trans. Am. Fish. Soc. 104:567-570.

Saarikoski, J. and K. Kaila. 1977. Effects of two chlorinated phenols on the spontaneous impulse activity of the abdominal tonic motor system in the crayfish (Astacus fluviatilis L.). Bull. Environ. Contam. Toxicol. 17:40-48.

Saarikoski, J. and M. Viluksola. 1981. Influence of pH on the toxicity of substituted phenols to fish. Arch. Environ. Contam. Toxicol. 10:747-753.

Saarikoski, J. and M. Viluksela. 1982. Relation between physicochemical properties of phenols and their toxicity and accumulation in fish. Ecotoxicol. Environ. Safety 6:501-512.

Salkinoja-Salonen, M., M. Saxelin and J. Pere. 1981. Analysis of toxicity and biodegradability of organochlorine compounds released into the environment in bleaching effluents of kraft pulping. In: Advances in the identification and analysis of organic pollutants in water. Vol. 2. Keith, L.H. (Ed.). Butterworth, Stoneham, MA. pp. 1131-1164.

Sarojini, R., A.K. Khan, M.S. Mirajkar and R. Nagabhushanam. 1983. Effect of sodium pentachlorophenate on the calcium content of the freshwater prawn (Caridina rajadhari). J. Environ. Biol. 4:77-80.

Schauerte, W., J.P. Lay, W. Klein and F. Korte. 1982. Influence of 2,4,6-trichlorophenol and pentachlorophenol on the biota of aquatic systems. Chemosphere 11:71-79.

Schimmel, S.C. and R.L. Garnas. 1985. Interlaboratory comparison of the ASTM bioconcentration test method using the eastern oyster. In: Aquatic toxicology and hazard assessment. Bahner, R.C. and D.J. Hansen (Eds.). ASTM STP 891. American Society for Testing and Materials, Philadelphia, PA. pp. 277-287.

Schimmel, S.C., J.M. Patrick, Jr., and L.F. Faas. 1978. Effects of sodium pentachlorophenate on several estuarine animals: Toxicity, uptake, and depuration. In: Pentachlorophenol: Chemistry, pharmacology and environmental toxicology. Rao, K.R. (Ed.). Plenum Publishing Company, New York, NY. pp. 147-155.

Shim, J.C. and L.S. Self. 1973. Toxicity of agricultural chemicals to larvivorous fish in Korean rice fields. Trop. Med. 15:123-130.

Sloley, B.D., B.E. Hickie, D.G. Dixon, R.G.H. Downer and R.J. Martin. 1986. The effects of sodium pentachlorophenate, diet, and sampling procedure on amine and tryptophan concentrations in the brain of rainbow trout, Salmo gairdneri Richardson. J. Fish. Biol. 28:267-277.

Slooff, W. 1978. Biological monitoring based on fish respiration for continuous water quality control. In: Aquatic pollutants: Transformation and biological effects. Hutzinger, O., I.H. VanLelyveld, and B.C.J. Zoeteman (Eds.). Pergamon Press, Oxford, NY. pp. 501-505.

Slooff, W. 1979. Detection limits of a biological monitoring system based on fish respiration. Bull. Environ. Contam. Toxicol. 23:517-523.

Slooff, W. 1982. A comparative study of the short-term effects of 15 chemicals on fresh water organisms of different trophic levels. PB83-200386. National Technical Information Service, Springfield, VA.

Slooff, W. 1983. Benthic macroinvertebrates and water quality assessment: Some toxicological considerations. Aquat. Toxicol. 4:73-82.

Slooff, W. and R. Baerzelman. 1980. Comparison of the usefulness of the Mexican axolotl (Ambystoma mexicanum) and the clawed toad (Xenopus laevis) in toxicological bioassays. Bull. Environ. Contam. Toxicol. 24:439-443.

Slooff, W. and J.H. Canton. 1983. Comparison of the susceptibility of 11 freshwater species to 8 chemical compounds. II. (Semi)chronic toxicity tests. *Aquat. Toxicol.* 4:277-282.

Slooff, W., J.H. Canton and J.L.M. Hermens. 1983a. Comparison of the susceptibility of 22 freshwater species to 115 chemical compounds. I. (Sub)acute toxicity tests. *Aquat. Toxicol.* 4:113-128.

Slooff, W., D. DeZwart and J.M. Marquenie. 1983b. Detection limits of a biological monitoring system for chemical water pollution based on mussel activity. *Bull. Environ. Contam. Toxicol.* 30:400-405.

Smith, R.A. and M.J. Ord. 1979. Morphological alterations in the mitochondria of Amoeba proteus induced by uncoupling agents. *J. Cell Sci.* 37:217-229.

Spehar, R.L., H.P. Nelson, M.J. Swanson and J.W. Renoos. 1985. Pentachlorophenol toxicity to amphipods and fathead minnows at different test pH values. *Environ. Toxicol. Chem.* 4:389-397.

Statham, C.N. and J.J. Lech. 1975. Potentiation of the acute toxicity of several pesticides and herbicides in trout by carbaryl. *Toxicol. Appl. Pharmacol.* 34:83-87.

Stephan, C.E., D.I. Mount, D.J. Hansen, J.H. Gentile, G.A. Chapman and W.A. Brungs. 1985. Guidelines for deriving numerical national water quality criteria for the protection of aquatic organisms and their uses. PB85-227049. National Technical Information Service, Springfield, VA.

Strufe, R. 1968. Problems and results of residue studies after application of molluscicides. Residue Rev. 24:80-168.

Stuart, R.J. and J.B. Robertson. 1985. Acute toxicity of pentachlorophenol to the freshwater snail, Gillia altilis. Bull. Environ. Contam. Toxicol. 35:633-640.

Sugiura, K., M. Aoki, S. Kaneko, I. Daisaku, Y. Komatsu, H. Shibuya, H. Suzuki and M. Goto. 1984. Fate of 2,4,6-trichlorophenol, pentachlorophenol, p-chlorobiphenyl, and hexachlorobenzene in an outdoor experimental pond: Comparison between observations and predictions based on laboratory data. Arch. Environ. Contam. Toxicol. 13:745-758.

Tagatz, M.E., J.M. Ivey, J.C. Moore and M. Tobia. 1977. Effects of pentachlorophenol on the development of estuarine communities. J. Toxicol. Environ. Health 3:501-506.

Tagatz, M.E., J.M. Ivey and M. Tobia. 1978. Effects of Dowicide G-SI on development of experimental estuarine macrobenthic communities. In: Pentachlorophenol: Chemistry, pharmacology and environmental toxicology. Rao, K.R. (Ed.). Plenum Publishing Company, New York, NY. pp. 157-163.

Tagatz, M.E., J.M. Ivey, H.K. Lehman, M. Tobia and J.L. Oglesby. 1980. Effects of drilling mud on development of experimental estuarine macrobenthic communities. In: Symposium on research on environmental fate and effects of drilling fluids and cuttings. Vol. 2. American Petroleum Institute, Washington, DC. pp. 847-865.

Tagatz, M.E., J.M. Ivey, N.R. Gregory and J.L. Oglesby. 1981. Effects of pentachlorophenol on field and laboratory-developed estuarine benthic communities. Bull. Environ. Contam. Toxicol. 26:137-143.

Tagatz, M.E., C.H. Deans, G.R. Plaia and J.D. Pool. 1983. Impact on and recovery of experimental macrobenthic communities exposed to pentachlorophenol. Northeast Gulf Sci. 6:131-136.

Thomas, P. and H.W. Wofford. 1984. Elevated acid-soluble thiol content in fish hepatic tissue: A response to pollutants. Mar. Environ. Res. 14:486-488.

Thomas, P., R.S. Carr and J.M. Neff. 1981. Biochemical stress responses of mullet Mugil cephalus and polychaete worms Neanthes virens to pentachlorophenol. In: Biological monitoring of marine pollutants. Vernberg, J., A. Calabrese, F.P. Thurberg and W.B. Vernberg (Eds.). Academic Press, New York, NY. pp. 73-103.

Thurston, R.V., T.A. Gilfoil, E.L. Meyn, R.K. Zajdel, T.I. Aoki and G.D. Veith. 1985. Comparative toxicity of ten organic chemicals to common aquatic species. Water Res. 19:1145-1155.†

Tiedge, H., R. Nagel and K. Urich. 1986. Effect of substituted phenols on transaminase activity in the fish, Leuciscus idus melanotus. L. Bull. Environ. Contam. Toxicol. 36:176-180.

Tomiyama, T., K.Kobayashi, N. Uyeda and K. Kawabe. 1962. The toxic effect of pentachlorophenate, a herbicide, on fishery organisms in coastal waters - IV. The effect on Venerupis phillipinarum of PCP which is constantly supplied or adsorbed to estuary mud. Bull. Jpn. Soc. Sci. Fish. 28:422-425.

Tomizawa, C. and H. Kazano. 1979. Environmental fate of rice paddy pesticides in a model ecosystem. J. Environ. Sci. Health B14:121-152.

Tripp, M.R., C.R. Fries, M.A. Craven and C.E. Grier. 1984. Histopathology of Mercenaria mercenaria as an indicator of pollutant stress. Mar. Environ. Res. 14:521-524.

Trujillo, D.A., L.E. Ray, H.E. Murray and C.S. Giam. 1982. Bioaccumulation of pentachlorophenol by killifish (Fundulus similis). Chemosphere 11:25-31.

U.S. EPA. 1980. Ambient water quality criteria for pentachloropnenol. EPA-440/5-80-065. National Technical Information Service, Springfield, VA.

U.S. EPA. 1983a. Water quality standards regulation. Fed. Regist. 48:51400-51413. November 8.

U.S. EPA. 1983b. Water quality standards handbook. Office of Water Regulations and Standards, Washington, DC.

U.S. EPA. 1985a. Appendix B - Response to public comments on "Guidelines for deriving numerical national water quality criteria for the protection of aquatic organisms and their uses." Fed. Regist. 50:30793-30796. July 29.

U.S. EPA. 1985b. Technical support document for water-quality based toxics control. Office of Water, Washington, DC. September.

U.S. EPA. 1986. Chapter 1 - Stream design flow for steady-state modeling. In: Book VI - Design conditions. In: Technical guidance manual for performing waste load allocations. Office of Water, Washington, DC.

Van Dijk, J.J., C. Van der Meer and M. Wijnana. 1977. The toxicity of sodium pentachlorophenolate for three species of decapod crustaceans and their larvae. Bull. Environ. Contam. Toxicol. 1:622-630.

Van Leeuwen, C.J., P.S. Griffioen, W.H.A. Vergouw and J.L. Maas-Diepeveen. 1985. Differences in susceptibility of early life stages of rainbow trout (Salmo gairdneri) to environmental pollutants. Aquat. Toxicol. 7:59-78.

Veith, G.D., D.L. DeFoe and B.V. Bergstedt. 1979. Measuring and estimating the bioconcentration factor of chemicals in fish. J. Fish. Res. Board Can. 36:1040-1048.

- Veith, G.D., D.W. Kuehl, E.N. Leonard, K. Welch and G. Pratt. 1981. Polychlorinated biphenyls and other organic chemical residues in fish from major United States watersheds near the Great Lakes, 1978. *Pestic. Monit. J.* 15:1-8.
- Verma, S.R., S. Rani, A.K. Tyagi and R.C. Dalela. 1980. Evaluation of acute toxicity of phenol and its chloro- and nitro-derivatives to certain teleosts. *Water Air Soil Pollut.* 14:95-102.
- Verma, S.R., S. Rani and R.C. Dalela. 1981a. Effects of phenolic compounds on in vivo blood parameters of a fish Notopterus notopterus. *J. Environ. Sci. Health* 16B:273-282.
- Verma, S.R., S. Rani and R.C. Dalela. 1981b. Synergism, antagonism, and additivity of phenol, pentachlorophenol, and dinitrophenol to a fish (Notopterus notopterus). *Arch. Environ. Contam. Toxicol.* 10:365-370.
- Verma, S.R., I.P. Tonk and R.C. Dalela. 1981c. Determination of the maximum acceptable toxicant concentration (MATC) and the safe concentration for certain aquatic pollutants. *Acta Hydrochim. Hydrobiol.* 9:247-254.
- Verma, S.R., S. Rani and R.C. Dalela. 1982. Alterations in certain organic components and serum electrolytes in Notopterus notopterus chronically exposed to phenolic compounds. *Acta Hydrochim. Hydrobiol.* 10:31-40.

Verma, S.R., I.P. Tonk, A.K. Gupta and M. Saxena. 1984. Evaluation of an application factor for determining the safe concentration of agricultural and industrial chemicals. *Water Res.* 18:111-115.

Vigers, G.A. and A.W. Maynard. 1977. The residual oxygen bioassay: A rapid procedure to predict effluent toxicity to rainbow trout. *Water Res.* 11:343-346.

Vigers, G.A., J.B. Marleave, R.G. Janssen and P. Borgmann. 1978. Use of larval herring in bioassays. In: *Proceedings of the fourth annual aquatic toxicology workshop*. Davis, J.C., G.L. Greer and I.K. Birtwell (Eds.). Canadian Fisheries and Marine Service Technical Report 818. Environmental Protection Service, Ottawa, Canada. pp. 31-52.

von Rumker, R., E.W. Lawless, A.F. Meiners, K.A. Lawrence, G.L. Kelso and F. Horay. 1974. Production, distribution, use and environmental impact potential of selected pesticides. PB-238795 or EPA-540/1-74-001. National Technical Information Service, Springfield, VA. pp. 308-319.

Walsh, G.E., D.L. Hansen and D.A. Lawrence. 1982. A flow-through system for exposure of seagrass to pollutants. *Mar. Environ. Res.* 7:1-11.

Webb, P.W. and J.R. Brett. 1973. Effects of sublethal concentrations of sodium pentachlorophenate on growth rate, food conversion efficiency, and swimming performance in underyearling sockeye salmon (*Oncorhynchus nerka*). *J. Fish. Res. Board Can.* 30:499-507.

- Weinbach, E.E. 1954. The effect of pentachlorophenol on oxidative phosphorylation. J. Biol. Chem. 210:545-550.
- Weinbach, E.E. 1956. The influence of pentachlorophenol on oxidative and glycolytic phosphorylation in snail tissue. Arch. Biochem. Biophys. 64:129-143.
- Weiss, U.M., P. Moza, I. Scheunert, A. Haque and F. Korte. 1982. Fate of pentachlorophenol-¹⁴C in rice plants under controlled conditions. J. Agric. Food Chem. 30:1186-1190.
- Whitley, L.S. 1968. The resistance of tubificid worms to three common pollutants. Hydrobiologia 32:193-205.
- Woelke, C.E. 1965. Development of a bioassay method using the marine algae - Macrocystis lutheri. Progress Report - Shellfish research. Washington Department of Fisheries and Shellfish, Seattle, WA.
- Woelke, C.E. 1972. Development of a receiving water quality bioassay criterion based on the 48-hour Pacific oyster (Crassostrea gigas) embryo. Technical Report No. 9. Washington Department of Fisheries, Seattle, WA.
- Wong, S.L. 1984. Toxicity and water quality: A bioassay interpretation. J. Environ. Sci. Health 19A:377-386.

Yount, J.D. and J.E. Richter. 1986. Effects of pentachlorophenol on periphyton communities in outdoor experimental streams. Arch. Environ. Contam. Toxicol. 15:51-60.

Yousri, R. and W. Hanke. 1985. The effects of pentachlorophenol, phenol and other pollutants on the liver of carp (Cyprinus carpio L.). Comp. Biochem. Biophys. 82C:283-290.

Zarogian, G.E. 1981. Results: Interlaboratory comparison-acute toxicity tests using the 48-hr oyster embryo-larval assay. Contribution No. 223. U.S. EPA, Narragansett, RI.

Zischke, J.A., J.W. Arthur, R.O. Hermanutz, S.F. Hedtke and J.C. Helgen. 1985. Effects of pentachlorophenol on invertebrates and fish in outdoor experimental channels. Aquat. Toxicol. 7:37-58.

Zitko, V., O. Hutzinger and P.M.K. Choi. 1974. Determination of pentachlorophenol and chlorobiphenyls in biological samples. Bull. Environ. Contam. Toxicol. 12:649-653.

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