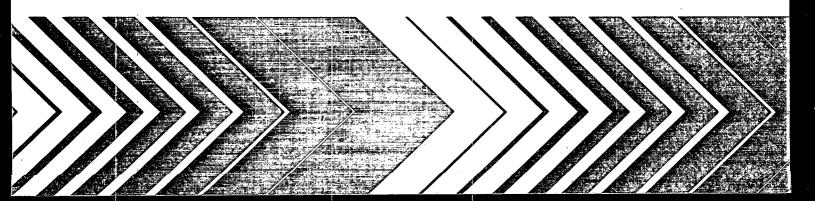
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Effects of Sludge Irrigation on Three Pacific Northwest Forest Soils



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EFFECTS OF SLUDGE IRRIGATION ON THREE PACIFIC NORTHWEST FOREST SOILS

by

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Grant No. R-802172

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FOREWORD

The Environmental Protection Agency was created because of increasing public and government concern about the dangers of pollution to the health and welfare of the American people. Noxious air, foul water, and spoiled land are tragic testimony to the deterioration of our natural environment. The complexity of that environment and the interplay between its components require a concentrated and integrated attack on the problem.

Research and development is that necessary first step in problem solution and it involves defining the problem, measuring its impact, and searching for solutions. The Municipal Environmental Research Laboratory develops new and improved technology and systems for the prevention, treatment, and management of waste water and solid and hazardous waste pollutant discharges from municipal and community sources, for the preservation and treatment of public and aesthetic effects of pollution. This publication is one of the products of that research; a most vital communications link between the researcher and the user community.

Major metropolitan areas are encountering increasingly complex problems with disposal of wastes. This report addresses one phase of the problem-disposal of municipal-industrial sewage sludge. Cities in forested regions can dispose of sludge in forests with economic benefits and no significant environmental consequences.

Francis T. Mayo, Director Municipal Environmental Research Laboratory

ABSTRACT

Forested regions of the United States offer an alternative for disposal of liquid wastes with the potential benefits to Society of economical waste disposal and accelerated forest growth. Digested liquid municipal sludges contain bio-degradable nutrients and sufficient water to enhance forest growth. Study was initiated in May 1974 of efficient methods of liquid sludge and chemical properties of forest soils and chemistry of soil water. A sprinkler irrigation system was developed for uniform applications of sewage sludge to forest plots.

The renovation capacity of the forest soil for most suspended and dissolved constituents in sewage sludge is very good (95 to 99+%). Nitrogen is the exception as nitrification rates increased with increased rates of sludge applications, resulting in significant leaching of NO_3 -N and concomitant cation losses. Phosphorus in all forms was never found in significant amounts in soil solutions of tested soils.

Sludge applications generally increased the growth rate of the forest stand. Water applied after sludge irrigation may further enhance tree growth over sludge only applications. Heavy metals were either absorbed in the aluminum oxide lysimeter plates or tied up in the soil profile. Human pathogens of the bacteria and virus types were not isolated from the limited number of soil and soil solutions analyzed.

This report was submitted in fulfillment of grant R-802172 by the University of Washington under the sponsorship of the U. S. Environmental Protection Agency and the Municipality of Metropolitan Seattle, Washington cooperatively with the Weyerhaeuser Company. This report covers a period of May 1, 1974 to April 30, 1978, and work was completed as of September 29, 1978.

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ABBREVIATIONS AND SYMBOLS

ABBREVIATIONS

```
-- acre
ac.
                -- basal area
BA
                -- centimeter
cm
                -- diameter breast height (the 4.5 ft. or 1.4 m height on
DBH
                   the bole)
                -- electron volts
ev
                -- foot/feet
ft.
ft2/ac
                -- square feet per acre
                -- cubic feet per acre
ft<sup>3</sup>/ac
                -- gram
q
                -- gallon
gal.
                -- grams per cubic centimeter
g/cm<sup>3</sup>
                -- hectare
ha
                -- horsepower
hp
                -- inch
in.
                -- kilogram
kg
                -- kilograms per squre centimeter
kg/cm<sup>2</sup>
                -- kilometers
km
                -- liter
                -- pounds per acre
lbs/ac.
                -- meter
                -- milli-equivalent
mea
                -- Municipality of Metropolitan Seattle
Metro
                -- milligram
mg
                -- mile
mi.
                -- minute
min.
                -- milliliter
ml
                -- millimeter
mm
                -- most probable number
MPN
                -- square meters per hectare
m²/ha
                -- cubic meters per hectare
m<sup>3</sup>/ha
                 -- metric tons per hectare
mt/ha
                 -- metric tons per hectare per year
 mt/ha/yr
                 -- oven dry
 0.D.
                 -- periodic annual increment
 pai
                 -- parts per million
 ppm
                 -- pounds per square inch
 psi
                 -- revolutions per minute
 rpm
                 -- square foot/feet
 sq. ft.
                 -- square meter
 sq. m
```

ABBREVIATIONS AND SYMBOLS (Continued)

```
-- volume
vol
yr.
                 -- year
SYMBOLS
                 -- calcium
Ca
                 -- cadmium
Cd
Cl
                 -- chlorine
                 -- copper
Cn
CO2
                 -- carbon dioxide
                 -- chromium
Cr
K
                 -- potassium
                 -- magnesium
Mg
Na
                 -- sodium
NH_4+
                 -- ammonium ion
NH<sub>4</sub>-N
                 -- ammonium nitrogen
Ni
                 -- nickel
N03-N
                 -- nitrate nitrogen
Pb<sup>°</sup>
                 -- lead
                 -- acidity
рΗ
P04-P
T0C
                 phosphate phosphorus
                 -- total organic carbon
                 -- total Kjeldahl nitrogen
Total N
                 -- total phosphorus
Total P
                 -- zinc
Zn
                 -- alpha
α
<
                 -- less than
μ
%
                 -- micro
                 percent
```

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Mr. Gerald Stern, project officer, U. S. Environmental Protection Agency, provided invaluable guidance throughout the three plus years required in the conduct and reporting of the study. Again, successful completion could not have been achieved without his help.

Section 1

INTRODUCTION

A national objective of improved quality of the Nation's waters has been implemented through state and federal laws and regulations. These laws and regulations have far reaching effects on disposal of wastes. Major municipalities concentrate both industrial and municipal solid and liquid wastes. Concentrations of urban population tend to concentrate problems with waste disposal. Rural and agricultural areas, both domestic and foreign, produce a host of products utilized in urban centers. Advent of the kitchen sink disposal unit transfers considerable of the solid waste disposal problems to the liquid waste system, with large quantities of organic wastes processed through liquid waste digestion rather than solid waste processes.

Burial, burning and disposal in water have become questionable solutions in waste disposal. Regions of the United States with extensive forest lands have the opportunity for beneficial use of certain of these liquid wastes in the forest. Water has an irrigation benefit during summer periods of soil moisture stress, and utilization of bio-degradable nutrients by the forest ecosystem will accelerate forest growth rates.

A research project funded in part by EPA and cooperative with the Weyerhaeuser Company, the Municipality of Seattle (Metro), and the College of Forest Resources, University of Washington studied application of liquid digested wastewater sludge in the forest environment. When applied to the forest, liquid digested municipal sludge can increase tree growth rates by adding nutrients and improving soil moisture. The renovation capacity of the forest soil and utilization of nutrients is dependent upon a series of physical, chemical and biological interactions in the forest ecosystem.

RESEARCH OBJECTIVES

To accurately interpret the impacts of sludge application on forest soils, as well as growth rate of trees on the forest plots, a very uniform liquid sludge application was required so that each area of each plot was receiving equal quantities of sludge and water.

- 1. Establish efficient methods of sludge application to forests.
- 2. Establish the short-term impacts of sludge application on the forest including effects on physical and chemical properties of the forest soil and chemistry of soil water.

- 3. Establish the rate of sludge application which has maximum benefits to forest growth with minimum impact on soil water quality and be non-polluting to surface or ground waters.
- 4. Establish the effects of application of sludge on forest growth rates.

The above objectives address applied sludge disposal problems, searching for economic solutions to liquid waste or sludge management. A variety of experimental designs would answer certain of the questions. These could include laboratory studies of leaching sludge in soil columns for soil and water chemistry; however, the complex interactions of weather, soils and related environmental factors suggested a field study, particularly for developing an effective method of sludge application in the forest environment and forest growth responses to applications of sludge.

The following report provides results of three years' study on applications of municipal-industrial sludge on a series of forest plots on soils with varying characteristics. By the very nature of the study, it was necessary to develop it in phases. The first year concentrated on development of the most feasible method of obtaining uniform application of the sludge to forested research plots.

The requirement for a uniform application of sludge over the experimental plots required a more complex sludge irrigation system than would be required for routine disposal of sludge in the forest environment. In the following sections, the results of the application phase of the study (Objective 1) will be reported prior to the other results, as this phase developed the final method of application which was used on plots established in later phases of the research.

The second year concentrated on determination of the impacts of varying rates of sludge application. The final phase investigated the impacts of both rates of sludge application to plots and impacts on three different forest soil types.

Section 2

THE STUDY AREA

The research sites are located 110 km (70 mi.) south of Seattle, Washington near the town of Eatonville and adjacent to and on Pack Demonstration Forest, College of Forest Resources, University of Washington. The initial work tested methods of liquid digested sludge application on a 1.6-ha (4-ac.) tract of Weyerhaeuser Company land at lower elevations on the Everett soil series. Later research assessed weekly rates of sludge application on the Everett, Mashel and Wilkeson soils series on Pack Forest. Forest growth, decomposition and aggregate impacts of three years of sludge application on soil and soil water chemistry were studied on the Everett soil series on Weyerhaeuser Company lands. Research on the Wilkeson and Mashel series was limited to impacts of rates of sludge application on soil and soil water chemistry.

CLIMATE

Climate in the vicinity of the study area has in general characteristics of the maritime climate of the Puget Sound Lowlands. Summers are relatively warm and dry followed by humid temperate winters with substantial winter rainfall. Annual precipitation increases markedly with increased altitude, an average of 25.4 cm (10 in.) per year per 305 m (1000 ft.) increase in elevation, with a marked decline in average annual temperature with increasing altitude. Air masses originate over the Pacific Ocean moderating both winter and summer temperatures. Precipitation is usually caused by orographic storms resulting in persistently high moisture content of winter air masses. Widespread precipitation of moderate to light intensity occurs as air masses cool and rise over land in winter. Fifty percent (60 cm or 24 in.) of the average total annual precipitation (120 cm or 48 in.) at Pack Forest falls in the four months of October through January. Total rainfall for July and August is usually less than 5% (11.9 cm or 4.7 in.) of the annual and often drought-like conditions may exist on well drained soils. At lower elevations (the Everett soil series), snow is common in the winter but seldom persists for over a few days. At higher elevations (the Wilkeson soil series), snow is more frequent and persistent.

During the warmest summer months, average maximum monthly temperatures are 21 to 24 °C (70 to 75 °F), while the coldest winter temperatures average slightly above freezing.

Annual evaporation from a Class A pan is estimated at 63.5 cm (25 in.) at nearby Puyallup, Washington 24 km (15 mi.) to the northwest. Monthly

evaporation rates vary from 7.5 to 16.5 cm (3 to 6.5 in.). Based on average normal temperatures and precipitation, estimates of actual annual evapotranspiration range from 38 to 46 cm (15 to 18 in.) in the mountains to 38 to 56 cm (15 to 22 in.) in the lower valleys (Pacific Northwest River Basin Commission, 1970).

THE EVERETT SOIL SERIES

The Everett soil series is a light colored, podzolic soil developed on loose gravelly to slightly modified or poorly sorted glacial outwash. Stratification is usually poor with a wide range in particle sizes of rocky gravelly components. The research site has a gravelly outwash overburden of 9 to 12 m (30 to 40 ft.) of gravelly outwash deposits covering a fine textured lacustrine strata. The lacustrine strata is almost impermeable to vertically percolating soil water; thus, most ground water flows laterally along the interface forming several springs in the vicinity of the research plots.

The soil classified as an Everett type is very coarse textured, immature, formed from these outwash deposits of the last glaciation. Everett soils are light colored podzolic soils developed on loose, coarse, gravelly, slightly modified or poorly sorted glacial outwash materials, stratified in places and having a wide range of rocks (Figure 1).

The Seventh Approximation (SCS, 1960) groups the Everett series in the brown podzolic group as a gravelly loamy sand, with the class Haplorthod, coarse loamy over sandy, skeletal mixed mesic (Schlichte, 1968). The soil profile description is:

<u>Horizon</u>	<u>Depth</u>	<u>Description</u>
Litter	4-0 cm (1.5-0 in.)	Very dark grayish brown (10 YR 3/2) loose or very friable matted layer of forest litter. Strongly acid (pH 5.6).
Α	0-15 cm (0-6 in.)	Pale brown or brown (10 YR 6/3, brown, 10 YR 4/3 moist) loose gravelly sandy loam. Structureless and with included hard rounded shot cemented with iron oxides.
В	15-60 cm (6 to 24 in.)	Light yellowish brown (10 YR 6/4, dark yellowish brown, 10 YR 4/4, or yellowish brown, 10 YR 5/4 moist) gravelly sandy loam with less shot.
С	60+ cm (24+ in.)	Light yellowish brown, light brownish-gray, gray with dark gray gravel, sand and cobbles, loose and porous, poorly sorted and somewhat stained with brown and reddish brown.

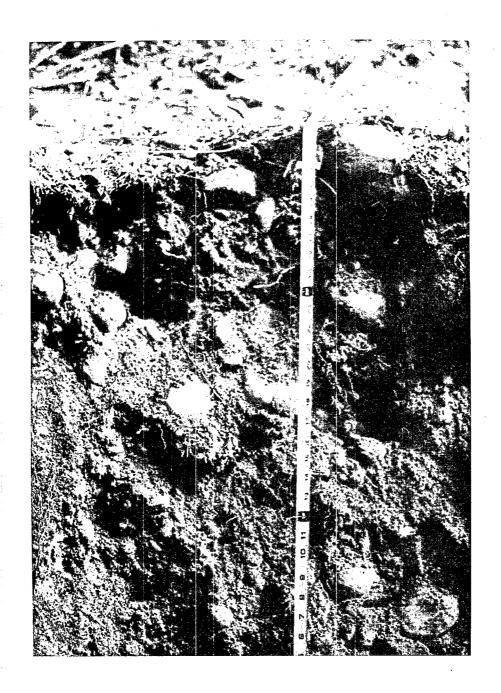


Figure 1. The Everett soil series is an immature, slightly podzolic soil developed on loose, poorly sorted glacial outwash.

The following lists the soil density, the percent larger than 2 mm (0.08in) fraction, and percentage of sand, silt and clay for the Everett soil.

Depth	Density	% Larger	Texture	of the 2mm Fra	action
(cm)	<u>(g/cm³)</u>	Than 2mm	<u>Sand</u>	<u>Silt</u>	Clay
			· -	Percent	
0-15	0.71	82	62	31	7
15-30	1.46	80	72	23	5
30-60	1.65	84	80	16	5
60 ÷	1.74	95	88	9	3

THE MASHEL SOIL SERIES

The Mashel series consists of deep, moderately well drained soils formed in glacial till. Mashel soils occur on uplands with slopes of 5 to 65 percent. Elevations range from 210 to 550 m (700 to 1800 ft.). Mean annual precipitation is about 22 cm (55 in.). Mean annual air temperature is about 9°C (49°F). The taxonomic class is a fine, halloysitic, mesic Ultic Haploxeralfs. (Colors are for moist soil unless otherwise noted.)

<u>Horizon</u>	<u>Depth</u>	Description
Litter	8-0 cm (3-0 in.)	Partially decomposed roots, leaves and twigs, 0 to 8 cm (0 to 3 in.) thick.
A	0-20 cm (0-8 in.)	Dark brown (10 YR 3/3) loam, brown (10 YR 5/3) dry; moderate fine and medium subangular blocky structure; slightly hard, friable, slightly sticky and slightly plastic; I percent hard rounded pebbles; many fine pores, very strongly acid (pH 5.0); clear smooth boundary, 13 to 23 cm (5 to 9 in.) thick.
В	20-41 cm (8-16 in.)	Dark brown (10 YR 4/3) heavy loam, pale brown (10 YR 6/3) dry; few faint dark brown mottles; moderate medium subangular blocky structure; hard, firm, slightly sticky and slightly plastic; 1 percent hard rounded pebbles; many fine and medium roots; many fine pores; thin patchy clay films on some faces of peds; very strongly acid (pH 4.7); clear wavy boundary, 15 to 23 cm (6 to 9 in.) thick.
C	41-90 cm (16-36 in.)	Yellowish brown (10 YR 5/4) clay loam, light yellowish brown (10 YR 6/4) dry; common bleaches silt and sand particles on faces of some peds and within some peds; moderate medium and coarse subangular blocky structure; hard, firm, sticky and plastic; 2 percent

strongly weathered rounded gravel, less than I percent hard pebbles; common fine and medium roots; common fine and medium pores, many thin to moderately thick dark brown clay films on faces of peds and in pores; few black stains; very strongly acid (pH 4.5); gradual wavy boundary, 25 to 56 cm (10 to 22 in.) thick.

Forest vegetation is dominantly western red cedar with some western hemlock. Subordinate vegetation is sword fern. The site is characteristically wet throughout the year.

THE WILKESON SOIL SERIES

The soils of the Wilkeson series are deep, well-drained soils formed in materials weathered from andesite and basalt. They are on slopes of the foothills. Mean annual precipitation is about 203 cm (80 in.) and mean annual air temperature is about 8 °C (49 °F) The taxonomic class is Eutric Glossoboralfs; fine loamy, mixed family. (Colors are for moist soil unless otherwise noted.)

Horizon	Depth	<u>Description</u>
Litter	5-0 cm (2-0 in.)	Partially decomposed needles, leaves, twigs and roots.
A :	0-10 cm (0-4 in.)	Very dark grayish-brown (10 YR 3/2) gravelly silt loam, grayish brown (10 YR 5/2) when dry; weak, very fine, subangular blocky structure; slightly hard, friable, slightly sticky and slightly plastic; common very fine roots; 25 percent shot and angular pebbles; medium acid (pH 5.6); abrupt, smooth boundary, 7 to 13 cm (3 to 5 in.) thick.
В	10-69 cm (4-27 in.)	Dark brown (10 YR 3/3) gravelly silt loam, brown (10 YR 5/3) when dry; weak, fine, subangular blocky structure; slightly hard, friable, slightly sticky and plastic; many fine, medium and coarse roots; 30 percent shot and angular pebbles; strongly acid (pH 5.4); diffuse wavy boundary, 13 to 18 cm (5 to 7 in.) thick.
C .	70-90 cm (27-36 in.)	Yellowish-brown (10 YR 5/4) gravelly silty clay loam, light yellowish-brown (10 YR 6/4) when dry; moderate, medium, prismatic, parting to moderate, medium, subangular blocky structure; hard, firm, sticky and plastic; few fine and coarse roots; common

very fine and fine pores; moderate thick clay film in fine pores; 10 percent angular pebbles; medium acid (pH 5.8); gradual, wavy boundary, 20 to 38 cm (8 to 15 in.) thick.

The forest stand was predominantly old growth western hemlock with a few young hemlock in the understory. Subordinate vegetation was a mixture of salal, red huckleberry and some grass species. The area of plot study was generally open as shown in Figure 2.



Figure 2. Plots on the Wilkeson soil were in an old growth western hemlock stand with larger older trees and an understory of young hemlock. The plywood box contains the sample collection bottles for the lysimeter plates.

Section 3

METHODS OF PLOT LAYOUT, SOIL SAMPLE AND ANALYSES

Study methods used for soil sampling and analyses, plot layout, etc. are conventional and accepted for similar phases of field forestry research. They involve the establishment of plots in the forest stand and random sampling of the forest soils. Tests of methods of sludge application resulted in design of an irrigation system for applications of sludge to experimental plots.

PLOTS ON THE EVERETT SOIL SERIES

Experimental plots were established on Weyerhaeuser Company land on the west edge of Pack Forest in Section 29, Township 16N, Range 4E, the Willamette Meridian (latitude 46° 50' W; longitude 122° 22' N). The elevation of plots on the Everett soil series is 230 m (750 ft.) above mean sea level. The bench (outwash on which the plots were located) breaks about 60 m (200 ft.) west of the plots, dropping steeply into the Nisqually River.

Plots on the Everett soil series were the most intensively studied. The first study phase developed a method of sludge application, and later phases studied forest growth rates and biological decomposition.

The forest stand was second growth Douglas-fir established in 1939 with an average density of 1580 trees per ha (640 stems per ac.) and average basal area of 35 sq. m per ha (150.7 sq. ft. per ac.). A few survivors of the old growth forest occurred randomly throughout the plot area. The understory consisted primarily of salal, Oregon grape and elderberry. Mosses and lichens are also numerous as soil cover and on downed organic materials (Figure 3).

Two areas were selected on the Everett soil series; one for initial testing of methods of sludge application, and the second for detailed studies of soils, soil solution chemistry, decomposition and forest growth as per the objectives. Square plots (0.04 ha or 0.1 ac.) were established using compass and tape for measurement of distances and angles. Three plots were established for testing methods of sludge application. Seventeen additional 0.04 ha (0.1 ac.) plots were established for testing rates of sludge application on soils, decomposition and impacts on forest growth rates.

Responsibility for assessment of effects of sludge applications on forest growth was assumed by the Weyerhaeuser Company as a cooperator. All trees in each plot were numbered with aluminum tags and measured for diameter breast height (DBH) to the nearest 0.25 mm (0.01 in.). Increment cores



Figure 3. The Everett soil supports a 36-year old Douglas-fir stand with an average of 1580 trees per ha. Standing dead trees and cull hardwoods were removed from the plot (pile in lower righthand corner).

were taken from dominant and co-dominant Douglas-fir to develop regression equations which estimated the previous rate of basal area growth. Stand biomass was estimated by DBH regressions (Dice, 1970) after testing the equations with additional trees.

The general layout of intensive studies on Everett plots is shown in Figure 4. Lysimeters were installed to sample volumes and chemistry of soil water under the forest floor at the boundary between the A and B horizons (15 to 25 cm or 6 to 10 in.) and the boundary of the B and C horizons (46 to 60 cm, or 18 to 24 in.). Lysimeters were also installed at a depth of 210 cm (7 ft.) below the C horizon. A continuous vacuum was maintained on each lysimeter at 0.1 atmospheres (3 in. or 76 mm of mercury) which approximates the tension of retention of gravitational water in the Everett soil matrix. A pressure sensitive switch operates on a central vacuum system activating the vacuum pump automatically to maintain constant tension on all lysimeters. Extracted soil water samples identified by plots and depths were taken to the lab at the University for analyses.

Pretreatment sampling of the forest soil by plots was made at the time of lysimeter plate installation. The forest floor and soil horizons were sampled for physical properties and chemical constituents.

SOIL SOLUTION MONITORING

A Metro-Data 616 data logger was coupled in the lysimeter soil water flow system with probes for field recording of pH, dissolved oxygen, conductivity and temperature of extracted soil water plots and at varying soil depths. These probes were scanned at frequent intervals depending on sludge applications or the dynamics of climatic events. Intermittent measurements for the above parameters of soil water extracted were taken with routine instrumentation to verify the accuracy of the recorded data.

Volumes of soil solution collected for laboratory analyses varied seasonally with rainfall; therefore, in calculating nutrient flux, irrigation and precipitation were assumed to be uniform over plots and infiltrated and percolated to sampled depth. Total flux of water was calculated for a given sample period including a correction factor for seasonal evapotranspiration. Chemical concentrations were averaged for a particular sampling period and then used in the determination of a nutrient flux. It should be noted that years 1, 2 and 3 are 10, 10 and 7 months, respectively.

Analyses for biological components (bacteria and virus) and heavy metals also were based on soil solution chemistry as monitored by lysimeter extracts for various soil horizons by plots.

SAMPLING FOR GROUND WATER NUTRIENTS

Wells were drilled as located in Figure 4 to study the chemistry of ground water adjacent to sludge treated plots. The glacial outwash overburden in the plot locations was approximately 10 m (33 ft.) thick overlying a relatively impermeable lacustrine substrata. Natural flow of water percolates through the Everett soil to the lacustrine outwash interface, then

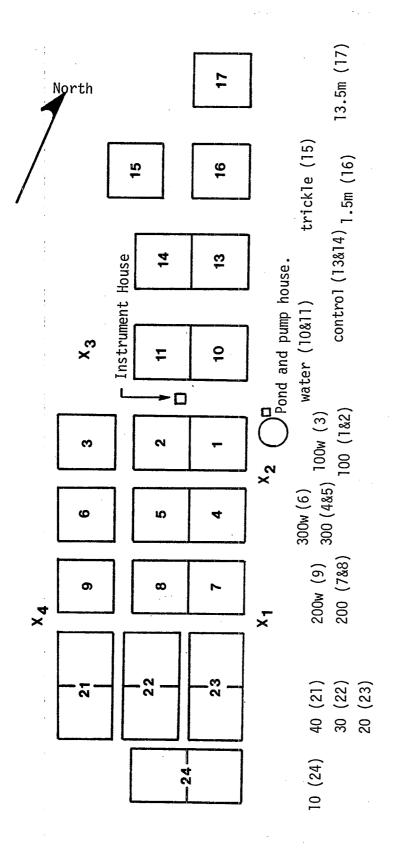


Figure 4. The arrangement of plots and treatments is shown for the study area on the Everett soils. Plot numbers are in parentheses while treatments indicate the mt/ha/yr. Locations of wells are shown with an XI, etc.

moves laterally to a number of springs on the slopes to the Nisqually River. Both the springs and water samples from the wells were monitored to assess impacts of sludge application on ground water chemistry.

Four wells were drilled to depths of 10+ m (33+ ft.). The wells were lined with 15 cm (6 in.) inside diameter PVC pipe and tops covered. Ground water samples were obtained by lowering sample containers into the free water and extracting a sufficient volume for total chemical analyses. Wells 1 and 2 provided a stable supply of water for chemical analyses. Wells 3 and 4 were frequently dry during drier periods, supplying only water during wet winter periods. The rates of flow of ground water through the well could not be measured; consequently, a measurement of nutrient flux in ground water was not possible. Only changes in concentration of ground water over time are reported.

Section 4

SLUDGE TREATMENTS TO PLOTS

The area of Weyerhaeuser forest land used for the testing of liquid digested sludge application was based on plot size spacing, proximity to an access road, and maximum replications possible within the constraints of transport of large quantities of sludge from Metro's West Point plant to Pack Forest. The plots were laid out with buffer strips separating different sludge treatment applications and the water only and control plots. Figure 4 shows the relative location of all plots on the Everett soil, the swimming pool sludge storage tank and pump house, and location of wells for sampling of ground water.

The initial study of the methods for liquid sludge application used plots 15, 16 and 17 (Figure 4) which are actually more remote from the study area than indicated. Plot 15 tested trickle sludge applications; 16 was under-the-canopy on 1.5 m (5 ft.) risers; and 17 was over-the-canopy on 9 m (30 ft.) risers.

The experimental design used duplicate plots with 100, 200 and 300 metric tons per ha per yr. of sludge application. Plots 1 and 2 are referred to as the 100 (100 mt/ha/yr or 44.5 t/ac/yr) series; 4 and 5 as the 300 (300 mt/ha/yr or 133.6 t/ac/yr) series; and 7 and 8 as the 200 (200 mt/ha/yr or 89 t/ac/yr) series. Plots 3 (100W), 6 (300W), and 9 (200W) each received 1.3 cm (0.5 in.) of irrigation water after the sludge application. Plots 10 and 11 received 1.3 cm (0.5 in.) of irrigation water only per inch, while plots 13 and 14 are untreated control plots.

Recognition of the need to reduce sludge application rates in the second year of study resulted in the establishment of plots with sludge application rates of 10, 20, 30 and 40 mt/ha/yr (4.5, 8.9, 13.3 and 17.8 t/ac.). These duplicate plots are identified as plots 24, the 10 series; plots 23, the 20 series; plots 22, the 30 series; and plots 21, the 40 series (Fig. 4).

PLOTS ON MASHEL AND WILKESON SOILS

The sludge study on the Mashel and Wilkeson soils was limited to the effects of sludge applications on soil solution chemistry. This investigation required a smaller plot area, reducing the problem of logistics for sludge transportation and applications.

Sludge was transported from the receiving area at the Everett plots to the sites of Mashel and Wilkeson soils on Pack Forest. Duplicate 0.01 ha plots (0.025 ac.) were established on a uniform slope to receive applications

of 10, 20, 30 and 40 mt/ha/yr (4.5, 8.9, 13.3 and 17.8 t/ac.). Lysimeter plates were installed in a manner similar to that of the Everett plots to a depth of 60 cm (24 in.), a zone midway in the B horizon. As with the Everett plots, lysimeters were placed under the forest litter layers and at the interface of the A and B soil horizons. Soil solution sampling and soil chemistry were similar to those methods used on the Everett plots.

Section 5

ANALYTICAL ANALYSES OF SOILS, SOIL SOLUTIONS AND GROUND WATER

Soil pits excavated to the C horizon for lysimeter installations were sampled in duplicate by horizons, establishing pretreatment exchangeable and total soil chemistry. Post sludge treatment soil samples were taken from new pits in a similar manner. Soil samples were air dried, then sieved for the less than 2 mm (.08 in.) fraction prior to analyses.

SOIL PROPERTIES

Soil pH was determined with a glass electrode on a solution of 1:1 soil to distilled water, allowing time to come to equilibrium. Litter layers were mixed with a 1:2 ratio water (Forest Soils Committee, 1953).

A modified Kjeldahl method was used to determine total soil nitrogen. The form of most nitrogen in forest soils is protein as amino acid groups attached to carbon. Organic matter is oxidized in concentrated sulfuric acid, a slow reaction catalyzed by selenium and additions of sodium sulfate. After digestion, the solution is made strongly basic and ammonia distilled into a dilute acid. Titration of the distilled solution with a standard determines the amount of ammonia absorbed (Bremner, J. M., 1965).

Soil organic material was determined by the modified wet combustion method (Walkley and Black, 1934). Soil samples are treated with an excess of oxidizing agent, sulfuric acid in chromic acid. Excess chromic acid is determined by a back titration with a standard ferrous sulfate solution. Heat of dilution from added concentrated sulfuric acid speeds the reaction. Chromic acid is not carbon specific but will react with any readily oxidizable substance. Approximately 77% of the total carbon in soil organic matter is oxidized. Approximately 58% of soil organic matter is carbon.

Total cation exchange capacity for soil samples was determined by leaching samples with normal ammonium acetate buffered to neutral pH (Jackson, 1958). Determination of cation exchange capacity involves saturating the soil colloidal exchange complex with a known cation (NH $_4$) and will exchange cations on the soil exchange complex. Displacing NH $_4$ cations by another leaching solution (Na) allows determination of total soil cation exchange capacity as milli-equivalents per 100 grams of oven dry weight of soil (meq/100g 0.D. soil).

The ammonium acetate solution containing the original displaced cations from the soil sample is analyzed for exchangeable cations (Ca, Mg, K and Na),

which are also expressed as meq/100g 0.D. soil. The sum of exchangeable cations divided by the total cation exchange capacity is termed 'base saturation' and expressed as percent.

Total chemistry of soils required digestion until dry (overnight) of a soil sample in a nitric-hydrofloric and hydrofulvic acid mixture (Jackson, 1958) in teflon beakers. The residue is then redissolved in distilled water and the solution analyzed for cations on the atomic absorption spectrophotometer. Pretreatment total soil chemistry values were used to define the statistical confidence intervals. Variability in total chemistry was high because less than 0.15 g of soil was digested. There are also large differences in actual soil particle sizes which make up the 0.15 g sample. It is assumed that chemical composition of less than 2 mm (0.08 in.) fraction is similar to the greater than 2mm (0.08 in.) fraction.

Carbon-nitrogen ratios are calculated from the average values for soil organic matter and total nitrogen. Organic matter is assumed to consist of 58% organic carbon. Dividing the percentage of soil organic carbon by the percentage of total nitrogen gives the carbon-nitrogen ratio.

SOIL SOLUTION AND GROUND WATER ANALYSES

Soil solution and ground water samples were transported to the University of Washington lab. One subsample is filtered through Whatman I filter paper and digested with sulfuric acid and hydrogen peroxide for total Kjeldahl nitrogen and total phosphorus determinations. Another subsample is filtered through Whatman glass fibre paper and used for anion and cation determinations. The remaining sample is used for pH measurement, alkalinity titration, conductance and TOC. All samples are stored at 2-40 C (36-390 F).

Total Kjeldahl nitrogen and total phosphorus are determined by a modified Kjeldahl (Standard Methods, 1971) which converts most forms of organic nitrogen and phosphorus into ammonium and phosphate. Ammonium and phosphate are then determined by Auto-analyzer methods. Total Kjeldahl nitrogen, nitrite and nitrate nitrogen are summed as total nitrogen. The difference between total Kjeldahl nitrogen and ammonium nitrogen is organic nitrogen.

Ammonium nitrogen (NH $_{4}$ -N), nitrate nitrogen (NO $_{3}$ -N), total Kjeldahl nitrogen (Total N), phosphate phosphorus (PO $_{4}$ P), total phosphorus (Total P), and sulfate sulfur are determined by use of a Technicon Autoanalyzer II. NH $_{4}$ -N is determined according to Technicon Industrial Method No. 108-71W/ Preliminary. Ammonium in the prepared sample is complexed with sodium phenoxide followed by sodium hypochlorite resulting in a blue indophenol-type compound, which is determined colorimetrically and is directly proportional to the ammonium nitrogen concentration (NH $_{4}$ -N).

NO₃N is determined according to Technicon Industrial Method No. 158-70W/Preliminary. Nitrate is reduced to nitrite in a copper-cadmium column. Reaction with sulfanilamide and coupling with N-I napthylethylene-diamine dihydiochloride produces a red compound which is measured colorimetrically. This method determines nitrite as well as nitrate nitrogen, but nitrite concentration in these water samples is assumed to be negligible.

Phosphate phosphorus is determined according to Technicon Industrial Method No. 155-71W. Phosphate reacts with ammonium molybdate in the presence of ascorbic acid and antinomy to form a blue phosphomolybdenum complex which is measured colorimetrically. This method determines only orthophosphate but contribution from other forms is negligible.

Sulfate sulfur is determined according to Technicon Industrial Method No. 226-72W. Interfering cations are first removed by a cation exchange column. Sulfate is then reacted with barium chloride. Excess barium forms a blue chelate with methylthymol blue, which is measured colorimetrically.

Calcium, sodium, potassium, and magnesium are determined by atomic absorption spectrophotometry using an Instrumentation Laboratory Model IL 353 atomic absorption spectrophotometer. Procedures described by the manufacturer are used (Procedure Manual for Atomic Absorption Spectrophotometry, Instrumentation Laboratory, Inc.).

Soil solution pH is determined using a Radiometer model 26 pH meter by methods outlined in Standard Methods (1971) and manufacturer's instructions. Total alkalinity is determined by an automated titration procedure (Standard Methods, 1971) with a Radiometer model II titrator and model II autoburette in conjunction with the pH meter, according to manufacturer's instructions. Conductivity is measured using a YSI model 31 conductivity bridge according to manufacturer's instructions and Standard Methods (1971). Chloride is determined on a Technicon Autoanalyzer II according to Technicon Industrial Method No. 99-70W Preliminary.

VIRUS DETERMINATION

Soil and soil water samples were analyzed for virus and certain bacteria by the Virology Lab. at the Children's Orthopedic Hospital, Seattle.

Soil and water samples were incubated at $35-36^{\circ}$ C ($95-97^{\circ}$ F) for 2 weeks on a variety of cell lines to test for viruses. These cell lines included African Green Monkey, Rhesus Monkey Kidney; Heteroploids Hep II and HL; and Diploid, Human Embryonic Tonsil. Presence of enteric viruses was determined by cytopathetic effects on cell deformations (Lenette and Schmidt, 1969 or minor modifications).

TOTAL AND FECAL COLIFORM DETERMINATIONS

Coliform analyses were provided by Metro. Soil and water samples were tested for presence of total and fecal coliforms. Samples were incubated 24 hours on a lactose broth as a presumptive test. Transfer to brilliant green lactose bile broth and continued incubation was the confirmed test. Gas production in fermentation tubes is the positive visual interpretation of coliform presence. Dilution tubes offer an estimate of the most probable number (MPN) per sample volume (Standard Methods, 1971).

Section 6

TOTAL ORGANIC CARBON COMPONENTS

The majority of organic carbon analyses measures humic and fulvic acid fractions (Schnitzer and Kahn, 1972; Goring, 1972; Gieseking, 1975). Organic molecules smaller than fulvic acids in forest soils have received less attention but are of concern in sludge studies, since refractory organic compounds such as pesticides are in this classification. Organic compounds are isolated by extraction with solvents such as ether, chloroform and methanol (Braids and Miller, 1975). Identification of extracted organic materials is done in several ways: chromatographic separation by gas, thin layer or gel filtration chromatography in combination with spectral identification.

Soil samples were extracted by redistilled ether saturated with water and sodium chloride in a stoppered erlynmeyer flask and gently shaken for 42 hours. The ether extract was decanted, dried with magnesium sulfate, and concentrated on a rotoevaporator. For gas chromatographic analysis, diazomethane was added to the sample to make methyl esters or free acids, according to directions from Aldrich Chemical Company.

TOTAL ORGANIC CARBON MEASUREMENTS

Total organic carbon measurements were made on a Dohrmann-Envirotech DC-50 TOC analyzer. The unit receives a 30-microliter injection of acidified (pH 2) aqueous sample in a platinum boat containing manganese dioxide as an oxidation promoter. The boat is moved into a 90° C(194° F) zone where volatile carbon and CO2 are removed. Volatile carbon components are trapped on a column of Porapak Q, while the CO2 is vented. The column is heated and backflushed. Volatile organics are burned to CO2. Residual organics in the boat are then moved to an 850° C(1560° F) zone where they are burned to CO2.

Measurement is done by passing the $\rm CO_2$ from pyrolsis over a Ni catalyst in a hydrogen-rich atmosphere. The $\rm CO_2$ is reduced to methane, then bubbled through a humidifier and burned in a flame ionization detector. Response of the flame detector is integrated so total mg/l or ppm of carbon are read directly from the instrument.

Gas Chromatography

A Perkin-Elmer model 990 gas chromatograph with flame ionization detection was used. Two columns were used with the following conditions:

	Column 1	Column 2
Column type Column length Injector temp. Manifold temp. Attenuation Column phase Program Rate Initial Split ratio Injection volume Flow	SCOT 15.2 m (50 ft.) 200° C. (392° F.) 200° C. (392° F.) x 40 DEGS 90°-180° C. (194-356° F.) 4° C./min. (7.2° F/min.) 1 min. 10/1 .2 1 5.5 ml/min.	SCOT 15.2 m (50 ft.) 275° C. (527° F.) 300° C. (572° F.) x 80 DEXSIL-300 180°-270° C. (356-518° F.) 6° C./min. (11° F/min.) 1 min. 10/1 .2 1 5.5 ml/min.

Mass Spectrometry

The gas chromatograph was connected to a Perkin-Elmer Hitachi RMS-4 mass spectrometer. The units are interfaced with a gold jet separator with other connections being gold or glass. Ionization voltages were normally 70 ev. Output was achieved through a B&F oscillograph, model number 3006.

Gel Filtration Chromatography

Sephadex G-75 gel with a solvent of 0.01 N HaOH was used for gel filtration work. A chromatronix CMP-1 pump was used for solvent delivery. Injection into the system was through a valve and sample loop. Detection was done with ultraviolet detectors at 254 mm (10 in.). The columns were 100 cm x 6 mm (39.4 in. x 0.24 in.) glass with chromatronix teflon fittings. Typical conditions were:

	System 1	System 2
Solvent Temp. Flow Sample size Chart speed Recorder Detector Wave length Attenuation	0.01N NaOH 22.5° C. (72° F.) 6 ml/hr. 1 ml 1.5 cm/hr. (0.6 in./hr.) Varian A-25 Chromatronix 254 mm (10 in.) 0.04	0.01N NaOH 22.8° C. (73° F.) 11.8 ml/hr. 1 ml 1.5 cm/hr. (0.6 in./hr.) ISCO, UA-5 ISCO, UA-5 254 mm (10 in.) 0.05

Using gel filtration chromatography, the largest molecules are eluted first, with the smaller molecules later. This assumes small interactions between the column packing and the organics being analyzed.

BIOLOGICAL DECOMPOSITION

Wood stakes (1 x 4 cm or 0.4 x 1.6 in.) prepared from sapwood of Douglas-fir (American Standard Testing Methods ASTM D1758-60T) were coded with a permanent identifying number and dried to a constant weight at 105° C (220° F) in a forced air oven (referred to as oven dried weights). Within the designated treatment plots, the stakes were placed at the following positions in the soil profile: 1) surface of the litter; 2) 1 cm (0.4 in.) below the litter; and 3) 10 cm (3.9 in.) below the litter layer. Five replicas of 10 stakes were placed at each position at the initiation of the test treatments. Representative wood stakes set aside as standard samples were stored in capped glass jars until needed for analysis.

At the prescribed sample time, one group of 10 stakes was recovered from the designated treatment site and soil profile positions. In the laboratory, extraneous debris clinging to the outside of the stake, such as soil particles, were removed. Immediately thereafter, the code number was recorded along with the wet weight of each stake. Samples were then oven dried at 105° C (220° F) for 24 hrs. and cooled in a vacuum desicator.

DECOMPOSITION ANALYSES

The percent stake moisture and weight losses were individually determined from the oven dry weight as follows:

Moisture content = $\frac{\text{wet weight - oven dry weight}}{\text{oven dry weight}} \times 100$

Weight loss = $\frac{\text{initial oven dry weight - final oven dry weight}}{\text{final oven dry weight}} \times 100$

1% Sodium Hydroxide Solubility

Solubility of wood in one percent sodium hydroxide, a measure of cellulose decomposition, was determined following the procedures of TAPPI Standard Methods (T4m-59) as modified by Cowling (1962). Stakes were ground in a Wiley mill until they passed through a 40-mesh screen (TAPPI T11m). Replicates from each treatment and profile position were combined to make a single sample.

One gram of the ground sample was added to a tared tall form 200 ml (6.8 oz.) Pyrex beaker. The sample was returned to the oven and its oven dry weight determined. To the beaker, 50 ml (1.7 oz.) of standardized 1%

sodium hydroxide were added. The beaker was covered with aluminum foil and placed in a boiling water bath for one hour. Every 15 minutes, the mixture was stirred to facilitate digestion.

Digested samplers were suction filtered using a tared cintered glass crucible of medium porosity. Quantitative transfer of the digestion mixture was accomplished by using two 50 ml (1.7 oz.) aliquots of hot distilled water to rinse the beaker. Next, the sample was washed with 50 ml (1.7 oz.) of 10% glacial acetic acid from the residue. The crucible and contents were removed to a 105° C (220° F) oven for 48 hours, cooled in a vacuum desicator and then weighed.

Solubility was expressed as a percentage of the moisture-free weight of the loss during digestion. Relative solubility was computed by comparing the solubility of untreated, unexposed stored standard samples with that of the treated sample as follows:

Above procedures are routine for analyses of biological decomposition of wood samples as per procedures of Minyard and Driver (1975).

EVALUATION OF TREE GROWTH

The Weyerhaeuser Company used the twelve 0.04 ha (0.1 ac.) plots for forest growth rate research. Growth plots coincided with sludge treatments identified in Figure 4. All dead trees and cull hardwoods were felled and piled outside of the plot area. Remaining trees were tagged by number at 1.2 m (3.9 ft.) above the ground line. Diameters of each tagged tree were measured at DBH (1.4 m or 4.5 ft.) and identified by a paint mark. Total tree heights were measured to the nearest 0.3 m (1 ft.) on 7 trees in each plot plus 5 additional trees to complete the range of diameters encountered. Increment core borings were taken from 10 trees per plot to assess the past growth rates.

Analysis of tree growth rates tested changes in diameter growth, basal area and volume of individual trees within each treatment by the following regression models.

1. (DBH75 - DBH70) / 5 = a + b (DBH70) 2. (DBH76 - DBH75) / 2 = a + b [(DBH75 - DBH70) / 5] 3. (DBH77 - DBH75) / 3 = a + b [(DBH75 - DBH70) / 5] 4. (BA75 - BA70) / 5 = a + b (BA70) 5. (BA76 - BA75) / 2 = a + b [(BA75 - BA70) / 5] 6. (BA77 - BA75) / 3 = a + b [(BA75 - BA70) / 5] 7. (Volume 76 - Volume 75) / 2 = a + b (Volume 75) 8. (Volume 77 - Volume 75) / 3 = a + b (Volume 75)

Where DBH 75 is the diameter of the tree in 1975, likewise DBH 70 would be the diameter in 1970. BA 75 is the basal area of the plot in 1975 and volumes are spelled out.

Slopes of regression lines for each treatment were compared to the slope of the unirrigated control using a "t" test.

$$t = \frac{\hat{\beta} - \beta o}{SE_{\hat{\beta}}}$$

where

 $\hat{\beta}$ = slope of treatment

 $\beta o = slope of unirrigated control$

 SE_{β}^{\wedge} = standard error of treatment slope

to test the null hypothesis that there were no differences in growth between each treatment and the control. A simple one-day analysis of variance DBH growth was conducted also with those plots that had replications to test for treatment effects.

A final analysis arrayed all trees by pretreatment diameter and periodic annual increment (pai). From this list, trees were selected and paired with each sludge treatment and with the control. Paired trees had equal pai and approximately the same initial DBH. Changes in growth rate were analyzed by DBH, basal area, and volume testing the sludge treated trees against equivalent control trees.

LIQUID WASTE PROCESSING AND SLUDGE PROPERTIES

Metro serves approximately a million residents plus industry in the Greater Seattle area for treatment of domestic and industrial waste liquids. Several smaller outlying communities provide primary and secondary waste treatment; however, residual sludge solids are pumped to and stabilized at the Metro West Point plant. Over the year, average daily flow into the treatment plant at West Point is 475 million liters (125 million gallons) of raw sewage. Portions of the Metro sewer system are still combined, mixing municipal-industrial effluents with urban runoff. Large seasonal variations occur in volumes of effluent treated and the ratio of solid-organic constituents.

Anaerobic digestion reduces solids by 40% in an average digestion time of 21 days. Residence time in the digesters ranges from 14 to 35 days, dependent upon volumes of input to the plant. These conditions cause highly variable sludge composition over time. Average daily solids production over the year is 50 metric tons (55 tons) on a dry weight basis.

Treated sludge was discharged into Puget Sound through a deep water dissipator pipeline until the 1972 amendments to the Federal Water Pollution Control Act. The Clean Water Amendments banned discharges of solids but allowed continued discharges of primary treated waste waters. Sludge dewatering facilities were installed at West Point plant to facilitate sludge disposal as a dewatered cake. Dewatering and disposal are very expensive, leading to consideration of alternatives such as sludge irrigation for disposal at lesser costs and with possible benefits as a fertilizer and water supplement.

Liquid digested sewage sludge (3-7% solids) is trucked from West Point to Pack Forest (approximately 110 km or 70 mi.) in a closed 19,000-liter (5000-gal.) tanker. Sludge may be applied directly or diluted with river water to provide a sufficient volume at the design concentration to irrigate all plots on a weekly basis (Fig. 5).

CHEMICAL COMPOSITION OF SLUDGE

Table I summarizes maximum, minimum and average concentrations of various constituents in Metro sludge applied to research plots. Total solids are the oven dry product of evaporation of a given quantity of liquid sludge. Total nitrogen, phosphorus and heavy metal constituents are determined as portions of the dried sludge. Other chemical constituents are determined on sludge and makeup water.

Table 1. MAXIMUM, MINIMUM AND AVERAGE COMPOSITION OF SEWAGE SLUDGE APPLIED TO PLOTS

Component	Minimum	Concentration Maximum	Average
All values percent (%)			
Total Solids	2.10	3.40	3.20
Calcium	2.40	3.20	2.64
Magnesium	0.50	0.70	0.55
Potassium	0.15	0.48	0.29
Sodium	0.32	0.76	0.51
All values parts per million	n (ppm)		
Total Nitrogen	945	1,860	1,600
NH ₄ -N	674	1,090	895
N0 ₃ -N	120	135	131
Total Phosphorus	625	922	900
Orthophosphate	150	900	850
Zinc	2,200	3,100	2,740
Copper	900	1,200	1,020
Chromium	500	800	620
Lead	240	400	356
Cadmi um	50	62	55
Mercury	7	12	11

Total solids varied from 2.1 to 3.4% averaging 3.2%. Total nitrogen ranged from 945 to 1860 parts per million (ppm) with an average of 1600 ppm. The NH₄-N component of total nitrogen dominated and ranged from 674 to 1090 ppm with an average of 895 ppm. Nitrate levels remained fairly constant ranging from 120 to 135 ppm with an average of 131 ppm. Of the total phosphorus (625 to 922 ppm), orthophosphate was the dominant form (150 to 900 ppm). Quantities of orthophosphate were more variable than total phosphorus with an average of 900 ppm for total and 850 ppm for orthophosphate.

Limited determinations were made for the heavy metals, zinc, copper, chromium, lead, cadmium and mercury. Large concentrations of zinc found in Seattle's municipal sludge indicate the probable contribution of industrial zinc plating and galvanizing processes. Zinc concentrations are relatively constant ranging from 2200 to 3100 ppm with an average of 2740 ppm. In a like manner, copper concentrations are also fairly constant ranging from 900 to 1200 ppm, averaging 1020 ppm. Chromium averaged 620 ppm with a range of from 500 to 800 ppm. Lead occurs in reduced amounts averaging 356 ppm with a range of 240 to 400 ppm. Concentrations of cadmium were fairly constant with an average of 55 ppm and a range of 50 to 62 ppm. Mercury occurred in relatively low concentrations, usually around 11 or 12 ppm and occasionally 7 ppm.

The concentrations of calcium, magnesium, potassium and sodium are sufficiently large to report their values as a percentage. Calcium concentrations are large ranging from 2.4 to 3.2%, averaging 2.64%. Magnesium averages about 20% of that of calcium with an average of 0.55%, ranging from 0.5 to 0.7%. Potassium occurs in minimum concentrations for this group of elements, averaging 0.29% and ranging from 0.15 to 0.48%. Sodium concentrations are quite similar to magnesium averaging 0.51% with a range of 0.32 to 0.76%.

STUDY OF SLUDGE APPLICATION METHODS

Initial calculations for proposed rates of sludge application were made for the plots on Everett soils on Weyerhaeuser Company lands. The proposed weekly application rates required transport of 38,000 l (10,000 gal) of (Fig. 5) liquid sludge per week from West Point to Pack Forest. The irrigation system used a 91,000 l (24,000 gal.) storage pond made from a backyard swimming pool 5.5 m (18 ft.) in diameter and 1.4 m (4.5 ft.) deep. A drain from a low spot in the center of the storage pond led to the pump and PVC piping system, sprinklers and trickle systems with gate valves and cold weather drains at spacings necessary to achieve uniform application of sludge over each test plot. Sludge discharged from the tanker under gravity (Fig. 6).

The system as designed was tested with water for the positioning of sprinklers as well as the adjustment of sprinkler heads and proper sizing of the nozzles. The three methods tested were over-the-canopy sprinklers on 9 m (30 ft.) risers, spray irrigation on 1.5 m (5-ft.) risers, and a uniform application through a trickle system which consisted of spaced 0.1 cm (0.25 in.) drain holes in PVC piping. Uniform coverage by sprinkler systems on plots was achieved by sprinklers placed 18.3 m (60 ft.) apart at corners of plots with 90° coverage. A series of trials with water resulted in uniform application of rates of water over areas of the plots.

This irrigation system, as designed by a consultant, had been used successfully for irrigation water as well as distribution of dairy barn manures over fields. Sludge, however, is another product and the pump with a closed impeller plugged immediately with hair and other coarse solids when attempting to distribute sludge. Few investigators contacted had experience with the particular problem of uniform sludge application over relatively small plots. Many investigators had experience with large irrigation canons used for spreading of a variety of waste material on a large area. Many systems were easily eliminated with a narrowing to a few pumping systems which would have the greatest potential for success. A final choice of a Moyno positive displacement pump (model 2SWG12H-CDQ) was powered by a 3-phase 10-hp motor, operating at 1250 revolutions per minute (rpm). Problems with the quantities of miscellaneous solids of varying sizes in the sludge required a pregrinding by a Maz-o-rator before sludge was delivered to the pump. The Maz-o-rator was also powered by a 10-hp motor operating at 1800 rpm (Fig. 7).

The combination of the Maz-o-rator and Moyno pump required a considerable modification of the previously designed electrical system. Power demands of the 10-hp motor required that each operate on an individual

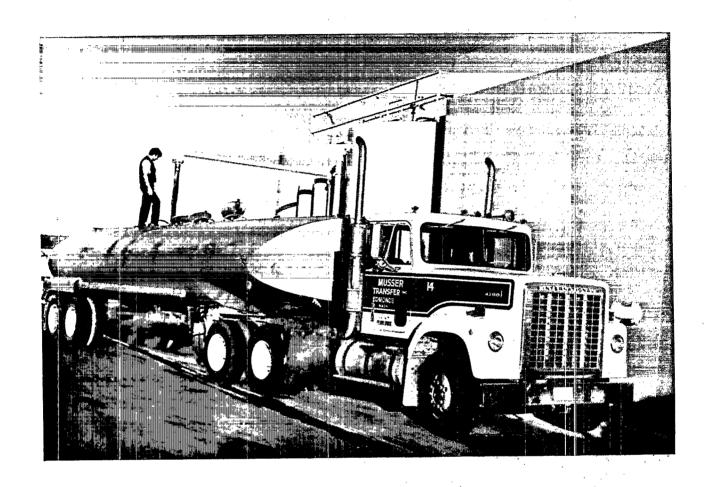


Figure 5. Sludge was transported from Metro's West Point Plant to Pack Forest in 19,000-liter quantities (5,000-gal.). Usually two trips were made each day sludge was applied.



Figure 6. A ramp was constructed so the tanker trucks could discharge sludge by gravity flow to the storage pond.

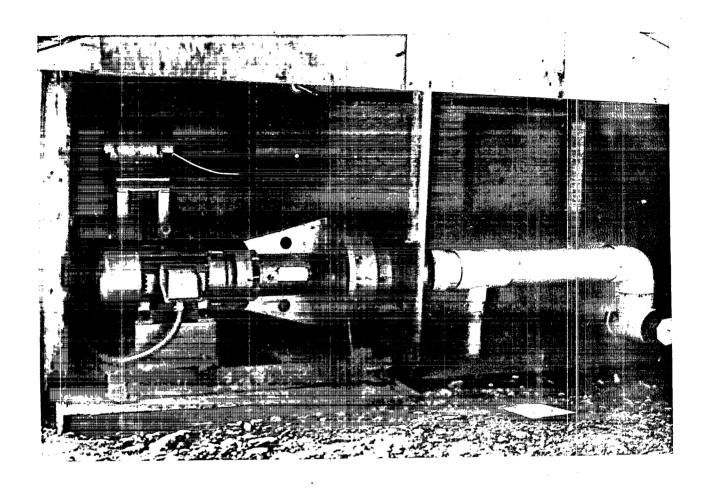


Figure 7. Adjacent to the storage pond, a small structure housed the combination Maz-o-rator and Moyno pump. Piping to the right is the suction line from the storage pond with the supply lines to plots underground.

switching and circuit system. As the Maz-o-rator initiates grinding and delivering of material to the pump, the Moyno pump is activated to distribute the ground sludge material to plots.

The above system very capably delivered sludge to the trial plots.

SPRINKLER MODIFICATIONS

The Maz-o-rator was a very effective grinder of solid materials, plastics, metals and stiffer paper products. The Maz-o-rator, however, was ineffective in grinding soft or limp materials, e. g., large quantities of hair. Care in installation of the plumbing distribution system with pipe connections and gate valves eliminated rough projections upon which hair would cling. The Rainbird 65D TNT sprinklers under trial, however, had many rough projections which retained hair, causing malfunction of the sprinklers. Malfunction included retention on spray deflectors, giving very poor distribution of sludge and obstruction of the mechanics of the sprinkler return mechanism necessary for achieving uniform distribution of sludge over plots. Occasionally, large solids would also obstruct the sprinkler nozzle orifice. The Rainbird 65D TNT sprinkler has a deflector arm with a lower bar upon which substantial amounts of hair would collect.

Modifications of the sprinkler head included removal of lower portions of the deflector arms and increasing the nozzle orifice size to 1.3 cm (0.5 in.) inside diameter. Rough projections on the brass casted sprinkler arm were ground smooth and the arm teflon coated. These modifications greatly reduced the accumulations of hair, allowing a uniform sweep and return of the sprinkler head under operation. Figure 8 shows the modified sprinkler after a year and a half of sludge applications without cleaning.

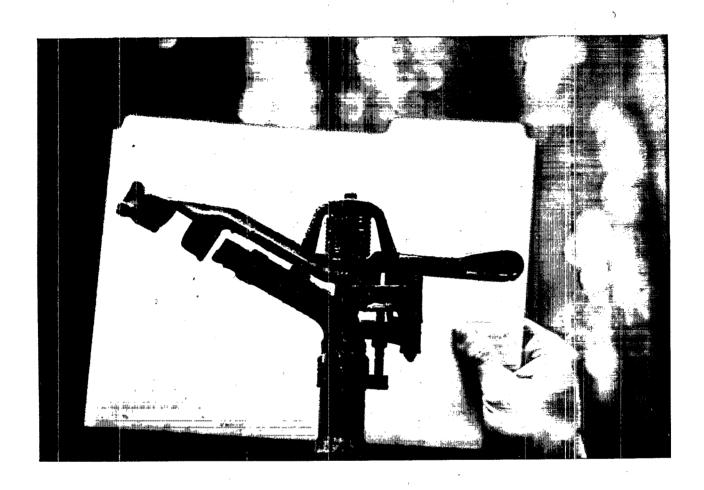


Figure 8. The Rainbird 65D TNT sprinkler had the lower deflector arm removed and was teflon coated, following removal of rough projections in the brass casting.

SUMMARY OF TESTS OF SLUDGE APPLICATION METHODS

The aforementioned modifications of the pumping and sprinkler systems allowed reasonably trouble-free sludge applications over trial plots (plots 15, 16 and 17, Fig. 4). Testing of the three methods of sludge application-over-the-canopy, spray irrigation and trickle irrigation-took place simultaneously with modification of the sprinkler heads.

OVER-THE-CANOPY APPLICATION

Over-the-canopy sludge application was suggested based on possible use of sludge irrigation through a complete forest rotation. Retention of sludge in the crowns of young forest trees and certain growth related impacts were of interest in trial of this application method. It soon became apparent that problems of cleaning plugged sprinklers on 9 m (30 ft.) risers was a major disadvantage of this application method. While risers in the forest crowns achieved over-the-canopy application, the interference of tree crowns caused poorer distribution of sludge over plots. Sludge dripping on workers from the canopy, either following sludge irrigation or during periods of rainfall at later dates, was also undesirable. Aerosol drift from over-the-canopy application remains a problem of unknown consequences.

TRICKLE IRRIGATION APPLICATION

The designed trickle irrigation system used perforated PVC pipe with sufficient 0.6 cm (0.25 in.) holes to achieve uniform distribution during water trials over the plot. Pressures of less than 1.0 kg/cm 2 (15 psi) were used for tests of sludge application by the trickle system.

Even though topography of test plots was sloped very gently, slight changes in topography (depressions) caused significant differences in pressure and quantities of sludge applied to areas of the plot. A slight sag in a pipe would induce a disproportionately large delivery of sludge to the low spot. Successive plugging of perforations also resulted in great spatial differences in rates and quantities of sludge applied. Normally occurring solid materials would be transported to the lowest ends of perforated pipes, where they became plugged preventing the drainage of sludge from the distribution system. A pond of sludge in the trickle area is shown in Figure 9.

UNDER-THE-CANOPY SPRAY IRRIGATION

Under-the-canopy spray irrigation was much easier to work with and cleaner than over-the-canopy or trickle systems. Unplugging of sprinkler

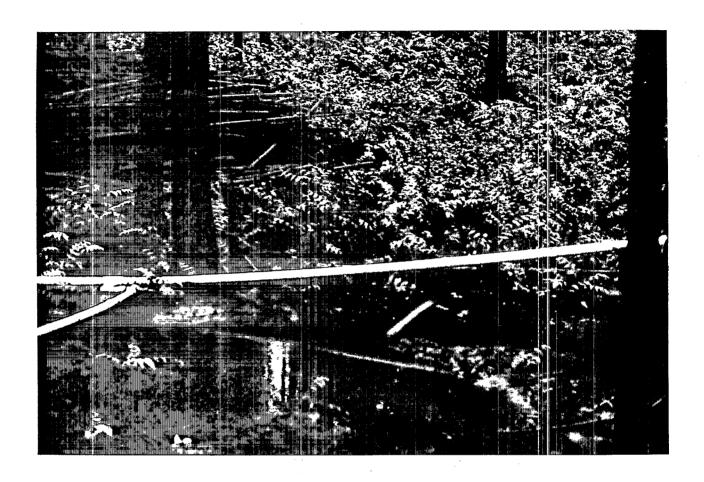


Figure 9. Heavy rates of sludge application caused impeded infiltration, resulting in ponding of sludge in depressions. This example from the trickle irrigation plot in an area of concentrated sludge applications.

heads on 1.5 m (5-ft.) risers was easily achieved. Distribution of sludge over the plots was also uniform without problems of sludge dripping from the forest canopy. Aerosol drift is also considerably less of a problem as wind movement under the canopy is significantly reduced.

The major disadvantage of under-the-canopy distribution is the increased amount of plumbing and more numerous sprinklers required for uniform unit area application. A final mode of operation evolved where grinding of sludge in all seasons of the year is definitely required. Metal bottle caps, all sorts of plastic, occasional dense hair mass, paper products and organic materials such as fruit pits are effectively ground to pass the plumbing system and sprinkler nozzles by the Maz-o-rator. The pumping system requires about 6.3 kg/cm² (90 psi) delivered by the pump on fairly level ground. Pressure drops within the pipe system result in about 2.8 to 3.2 kg/cm² (40 to 45 psi) delivered to the sprinkler orifice. This pressure is sufficient to operate the sprinkler efficiently and achieve good distribution of sludge over the plots. An operating sprinkler is shown in Figure 10.

Nozzle pressures of about 3.2 kg/cm 2 (45 psi) did not cause damage to the bole of Douglas-fir when over 3 m (10 ft.) from the sprinkler. Ideally, sprinklers should be farther from trees as poor distribution results from trees too close to sprinklers.

Initial rates of sludge application on the Everett plots were 0.6, 1.3 and 1.9 cm (0.25, 0.5 and 0.75 in.) weekly in one short application to duplicate plots. Irrigation through the large modified nozzles required 4.25 minutes to apply an area depth of 0.6 cm (0.25 in.) to a plot. The experimental design called for a third series of plots to receive an additional 1.3 cm (0.5 in.) of water. This water, in effect, washed off the sludge from subordinate and forest vegetation and provided an additional quantity of water to test the impacts of irrigation.



Figure 10. Sprinklers were located on 5-ft. risers at each corner on the square plots.

INITIAL SLUDGE APPLICATIONS

Results of certain initial trials and development methods of sludge application have been summarized in preceding sections. Results described in the following section are a combination of visual observations of the physical impacts of sludge application at design rates to particular plots, as well as analyses of soil solution chemistry.

A range of initial rates of sludge application was calculated to adequately replace soil moisture lost by summer evapotranspiration. Solid loading was projected on the basis of carbon-nitrogen ratios and total chemical constituents expected in sludge.

SLUDGE APPLICATIONS AND CONSTITUENTS

Solids in sludge received from Metro averaged 3.2% compared to expected 2.1% average. Total nitrogen averaged 1.6% compared to an expected concentration of 0.9%. Average carbon-nitrogen ratio was 11:1, rather than 18:1 as expected.

Even had constituents been as expected, impairment of surface soil infiltration capacity would have occurred in plots when total applied sludge equalled about 25.4 cm (10 in.). Reduced infiltration occurred about 12 to 14 weeks after initiation of sludge application for the heaviest 1.9 cm (0.75 in.) weekly rates (300 series, 300 mt/ha/yr or 135 t/ac/yr) of sludge application. Ponding of sludge occurred over 5 to 25% of the plot area (300 plots) by surface flow of sludge from localized spots of impeded infiltration to low spots or micro-depressions. Sludge applications (300W plots) followed by 1.3 cm (0.5 in.) of irrigation water had considerably less problems with impeded infiltration, even at greater sludge application rates. Plots 200 developed similarly impaired infiltration problems as total sludge applications approached 25 cm (10 in.). Sludge applications were terminated for the winter in December 1975. Plots 100 and 100W maintained good infiltration over 98% of the plot area with only very minor ponding occasionally occurring in low spots. Total application to 100 plots was 12 to 13 cm (4.7 to 5.1 in.). Undoubtedly, if these coarse textured, gravelly Everett soils developed impeded infiltration with the 200 and 300 mt/ha (88 and 135 t/ac.) rates of waste and sewage sludge application originally used, most other forest soils would develop similar problems at equivalent application rates.

Initial heavy rates of sludge applications (1.9 cm or 0.75 in./wk.) established the upper limit of sustained short-term (in 12 to 14 weeks)

sludge loadings. Impeded infiltration resulted from repeated heavy weekly applications of sludge which formed a nondrying film over the forest floor. Applications of sludge to the 100 series plot usually dried before a repeat application a week later. Short-term sludge applications to the 100 series plots were about one-half (12 to 13 cm or 4.7 to 5.1 in.) of the maximum short-term loadings (25 cm or 10 in.). Maximum loading rates have an economic advantage in that reduced forest areas as well as reduced amounts of plumbing would be required for maximum amounts of sludge disposal. Quantities of solid constituents in sludge and the percentage of total nitrogen content exceed estimated amounts used in designing initial sludge application rates. Nitrogen constituents were double those expected, based on Metro's chemical analyses, and the heterogeneous quantity and variety of solid constituents also was much greater than expected.

Large inherent variability in properties of forest soils, along with variability in sludge chemistry, makes interpretation of significant impacts of sludge application to soils difficult. When applications of sludge were terminated in December 1975, it appeared the 100 and 100W (100 mt/ha or 44 t/ac.) treatments were probably approaching the maximum short-term limit in rate of sludge application. Heavier sludge applications had resulted in excess nitrate concentrations in soil solutions. Modified rates of application were proposed at 10, 20, 30 and 40 mt/ha (4.5, 8.9, 13.3 and 17.8 t/ac.). On re-initiation of sludge application in 1976, these reduced rates were used on a series of new plots on the Everett soil and extended to the Mashel and Wilkeson soils. Applications of sludge continued on the original plots with the lowest rates (10 mt/ha or 4.5 t/ac.) going on plots which had received 100 mt/ha/yr (44 t/ac./yr.) rates during the first year. Heavier applications were applied to those plots which previously had received the heavier applications--20 mt/ha (8.9 t/ac.) on 200 plots and 30 mt/ha (13.3 t/ac.) on 300 plots.

THE IRRIGATION SYSTEM AND SLUDGE APPLICATION RATES

No problems were encountered with Metro's delivery of sludge to the Pack Forest plots. Tankers unloaded the sludge in the storage reservoir (Fig. 5) from which it was pumped by the combination Moyno-Maz-o-rator system to research plots.

The final distribution system for uniform application of sludge to 0.04 ha (0.1 ac.) plots used the modified Rainbird 65D TNT sprinklers with 1.3 cm (0.5 in.) I. D. nozzle, spaced 18.3 m by 18.3 m (60 x 60 ft.). The combination grinder (Maz-o-rator) and positive displacement Moyno pump delivered a line pressure of 6.3 kg/cm² (90 psi) with 3.2 kg/cm² (45 psi) to the sprinklers. Sprinklers on 1.5 m (5 ft.) risers gave uniform area-depth sludge applications over plots (Figure 10).

FIRST YEAR APPLICATIONS

As explained in a previous section, first year trial applications were 0.6, 1.3 and 1.9 cm weekly (100, 200 and 300 mt/ha/yr series plots; 0.25, 0.5 and 0.75 in.) in one short application to duplicate plots. A third plot in each series received an additional 1.3 cm (100W, 200W and 300W plots) (0.5 in.) of water to wash off sludge remaining on subordinate vegetation and to establish effects of additional water. As shown in Figure 4, there were also plots of no treatment (controls) to test departures from normal as well as the duplicate plots which received 1.3 cm (0.5 in.) of water only per week.

Plots were irrigated with sludge from July 22, 1975 to December 15, 1975 applying 1.8 to 5.5 mt/ha (0.8 to 2.5 t/ac.) of solids per week. Irrigation was terminated (Dec. 1975) due to excess rain and re-initiated May 15, 1976 at reduced rates of sludge applications.

SECOND YEAR APPLICATIONS

Commencing May 15, 1976, sludge was diluted 3-fold with water, resulting in a solution with 1-2.3% solids. Application rates of 0.25, 0.50 and 0.75 cm (0.1, 0.2 and 0.3 in.) per week of this sludge were made to the 100, 200, 300 and W series plots. If sludge was applied each week all year, applied solids would range from 190 to 580 kg (420 to 1,270 lbs.) per week or approximately a range of 10 to 30 mt/ha (4.5 to 13.5 t/ac.) per year.

The series of new plots on the Everett, Wilkeson and Mashel soils also received total sludge application in the above ranges. Four rates of sludge were applied to duplicate plots--0.25, 0.50, 0.75 and 1.0 cm (0.1, 0.2, 0.3 and 0.4 in.). The plots were identified as the 10, 20, 30 and 40 series plots. Plots established on the Wilkeson and Mashel soil series are remote from electrical power. Diluted sludge was transported to Wilkeson and Mashel plots by a 2 1/2-ton, 6-wheel drive truck carrying a 7600-liter (2000-gal.) tank. Sludge was gravity fed into a Deming open impeller pump powered by a 15-hp Wisconsin gas engine. The Deming pump delivers approximately 380 liters (100 gal.) per minute at a line pressure of 6.3 kg/cm² (90 psi). Travel time between sites was approximately 20 minutes with 7600 liters (2000 gal.) applied over 0.16 ha (0.4 ac.) in approximately 40 minutes.

RESULTS OF SLUDGE APPLICATIONS ON SOILS AND SOIL WATER

The following sections summarize the results of sludge applications on soils over time by describing changes in soil properties, soil water ions and related impacts. The forest soil provides renovation of suspended and dissolved constituents in soils through physical filtration and chemical reactions. The soil solution as sampled by the lysimeter plates estimates dissolved constituents intransient in the forest soil. The transient soil solution provides the mechanism for transfer of both suspended and dissolved materials from soil solution to ground water. The general term "flux" is used to indicate the transient condition of nutrients in soil or ground water.

Discussion of significant effects of sludge application on soil or soil water chemicals infers a statistically significant difference at the 95% confidence interval. An asterisk is used in the tables to indicate a significant departure in average values. Frequently, a trend in change is indicated. The interaction between limited numbers of samples and large variation in forest soils restricts a rigid interpretation of statistical significance. If additional degrees of freedom were available in sampling, statistically significant differences would be observed. The following sections first summarize the physical renovation of solids by the soil profile, followed by sections which summarize soil solution chemicals and soil properties. The Everett soils were studied over three years. Results are presented as a summary of the first and second years on Everett soils and the end impact of three years of sludge application. Only one year of data through one growing season is available for the Mashel and Wilkeson soils.

RENOVATION OF SOLIDS

Forest soils are a very efficient physical filter of the solids normally occurring in applied sludge. The Maz-o-rator ground most non-organic solids sufficiently fine so they were not apparent on plots. Large first year applications of sludge, as previously discussed, did impede infiltration causing ponding in low spots and depressions. Even these low spots eventually drained with a retention of the solids on the soil surface.

Slight penetration of solids into soils was indicated in places by a change in color of surface soils, particularly on plots which received water following sludge applications. Penetration was only a few millimeters; thus, the forest soil also effectively filtered for the usual size of organic solids in the sludge applied. Reduced rates (second year) of sludge

application did not cause reduced infiltration. In fact, plots with lowest quantities of sludge application do not look significantly different from control plots.

Total quantities of sludge actually applied to plots by years are summarized in Table 2. Table 2 shows total quantities of sludge applied to older plots on Everett soils. Heavy first year application ranged in depth from 12.1 to 26.7 cm (4.7 to 10.5 in.). Equivalent weights of solids applied ranged from 38.7 to 85.4 mt/ha/yr (17.2 to 38 t/ac/yr.). Applications of sludge were continued during the first year until obvious signs of impaired infiltration were observed. Discontinuing sludge applications at varying times resulted in different quantities of sludge applied to individual plots, but in no case did the quantities applied approach design rates.

The second year of sludge application proceeded at reduced rates. Sludge was actually applied for only 10 months; thus, maximum rates again were only approached. Depth ranged from 0.8 to 5.3 cm (0.3 to 2.1 in.) with equivalent weights of 2.6 to 17 mt/ha/yr (1.2 to 7.6 t/ac/yr.).

In the final year, emphasis was placed on application of sludge to new plots on the Everett, Wilkeson and Mashel soils. Again, early termination of sludge application did not allow sludge to be applied at design annual rates. However, in most cases, weekly application rates for much of the year were at design rates; only early termination of sludge applications resulted in reduced total amounts. Heaviest third year applications were made on Everett soils with reduced applications on Wilkeson soils. Maximum total amounts of sludge were applied to the older plots on Everett soils with a range of 45.1 to 108.4 mt/ha (20.1 to 48.3 t/ac.) over the life of the study. In depth, this is a range from 14 to 33 cm (5.5 to 13 in.).

Older plots on Everett soils as well as all new plots were excellent physical filters of suspended organic solids common in sludge. An absolute distinction has not been made between organic solids from sludge and larger organic molecules identified in the total organic carbon analyses. Leaching organic carbons deeper in soils is assumed to be as secondary decomposition products of bio-degradable materials, as compared with organic materials common in sludge in their original form. Thus, forest soils may be considered a 100% effective physical renovator for total suspended solids from sludge. Flux of nutrients and colloidal organic molecules through the soil profile must take place, in part, in dissolved forms and, in part, from accelerated decomposition of naturally occurring organic carbon to colloid size.

Table 2. TONS OF SLUDGE APPLIED BY PLOTS AND SOILS AND YEAR IN DEPTH AND TOTAL WEIGHTS

Plot	First	Year	Second	l Year	Third	d Year		otal
	Depth (cm)	Tons/ ha	Depth (cm)	Tons/ ha	Depth (cm)	Tons/ ha	Depth (cm)	Tons/ ha
Everett Soil								
100	13.3	42.6	0.8	2.6	1.2	3.8	15.3	49.0
100W	12.1	38.7	0.8	2.6	1.2	3.8	14.1	45.1
200	26.7	83.2	2.5	8.0	1.4	4.5	30.6	95.7
200W	23.5	75.2	2.3	7.4	1.3	4.2	27.1	86.7
300	26.0	85.4	5.3	17.0	1.9	6.1	33.2	108.4
300W	21.6	69.1	5.3	17.0	1.5	4.8	28.4	90.9
10				#	2.0	6.4	2.0	6.4
20					6.2	19.8	6.2	19.8
30	•				8.2	26.2	8.2	26.2
40			ı		9.3	29.8	9.3	29.8
Wilkeson Soil								
10					1.5	4.8	1.5	4.8
20					3.1	9.9	3.1	9.9
30			**		4.4	14.1	4.4	14.1
40					5.8	18.6	5.8	18.6
Mashel Soil			•					
10					1.5	4.8	1.5	4.8
20	y			,	3.3	10.6	3.3	10.6
30					4.8	15.4	4.8	15.4
40					6.5	20.8	6.5	20.8

EVERETT SOILS, SOIL SOLUTION NUTRIENTS

Those treatments shown in Table 2 which received 1.3 cm (0.5 in.) of water following sludge application also received quantities of dissolved nutrients in the water, as listed on the bottom of Table 6 (kg/ha). The amounts are insignificant by comparison with the quantities of nutrients applied in the sludge.

Table 3 summarizes the first year nutrient loading applied in sludge and the flux of soil solution to deeper than 2.1 m (7 ft.) in the soil profile. Nutrients applied as a loading in the sludge (Table 3) are calculated from average concentrations (Table 1). Flux of nutrients past the 2.1 m (7 ft.) depths incorporates large variability in both volumes of soil solution leachate (as collected by lysimeters) and concentrations of nutrients in soil solution samples. Large concentrations of nutrients in the soil solution coincoide with low soil moisture and very slow rates of moisture flow. Conversely, low nutrient concentrations in soil solutions occur during wet conditions with large amounts of rapid flow of soil water through the soil profile.

The renovation capacity of the soil is expressed as the percentage of an element applied and the quantity leaching (flux) deeper than 2.1 m (7 ft.) in the soil. A large percentage renovation (Total-P, 99+%, Table 3) indicates retention in the soil profile, with very little leaching or flux. Conversely, a large flux or high leaching losses indicates reduced renovation. Renovation of the Everett soil for total nitrogen was surprisingly efficient in view of the large quantities of total -N applied in sludge--1928 to 4261 kg/ha (1720 to 3801 lbs./ac.). The total -N flux or quantity leaching past the 2.1 m (7 ft.) depth in the soil profile was both largest and least of all treatments for the 300 and 300W treatments--221 to 2333 kg/ha (197 to 2081 lbs./ac.), respectively. Renovation of total -N also had a maximum range from 32% (300W) to 95% on the 300 plots. The 200 series plots reversed the impacts of water on the 300 plots with the 200W renovating 91% (340 kg/ha flux or 303 lbs./ac.) as compared to 65% (1500 kg/ha flux or 1338 lbs./ac.) for the 200 plot.

To date, there is no explanation for the reversal of renovation of total nitrogen between the 200 and 300 series plots with or without water. Great care was taken in obtaining these results with very careful checking for errors. Clues are not offered by the 100 series plots, as the renovation of the water added plot is not significantly different from the plot without water. The pattern established in the first year continued in future years,

Table 3. FIRST YEAR NUTRIENT LOADING AND FLUX (kg/ha) THROUGH THE EVERETT SOIL AND PERCENT RENOVATION

⊻,	3,867	3,500	7,734	6,813	7,550	6,261
	105	31	12	69	7	439
	97	99	99+	99	99+	93
Na	6,800	6,153	13,601	11,982	13,277	11,011
	46	98	139 *	109	44	196
	99	98	99	99	99+	98
Mg	7,366	6,600	14,721	12,969	14,371	11,918
	102	162	345	87	56	642
	99	98	98	99	99+	95
Ça	35,151	31,803	70,303	61,933	68,628	56,911
	617	610	1,108	436	231	2.594
	98	98	98	99	99+	95
P0 ₄ -P	200	181	400	352	390	323
	2	T*	T	T	T	1
	99	99+	99+	99+	99+	99+
Total -P	867	784	1,733	1,527	1,692	1,403
	2	1	T	1	T	1
	99+	99+	99+	99+	99+	99+
N03-N	175	730	350 1,493	308	341 220 35	2,211
NH ₄ -N	1,269	1,147	2,535	2,234	2,475	2,053
	290	1	2	29	1	13
	77	99+	99+	99	99+	99+
Total -N	2,130	1,928	4,261	3,754	4,160	3,450
	863	738	1,500	340	221	2,333
	59	62	65	91	95	32
	Applied	Applied	Applied	Applied	Applied	Applied
	Flux	Flux	Flux	Flux	Flux	Flux
	% Ren.	% Ren.	% Ren.	% Ren.	% Ren.	% Ren.
Plots	100	100 + water	200	200 + water	300	300 + water

*T denotes trace levels

and the reversal between the 200 and 300 series plots was consistent for other elements (Ca and Mg). It can only be assumed that the differences represent random variation in soils and their renovation abilities.

This random variation in renovation capacity also shows in the NH₄-N differences with the 100 plot without water having significantly greater losses (290 kg/ha or 259 lbs/ac.) than any other plots. Losses of NH₄-N were insignificant on all other plots (1 to 29 kg/ha or 0.9 to 25.9 lbs/ac.).

Renovation of NH₄-N ranged from 77 to 99% (290 to 1 kg/ha or 259 to 0.9 lbs/ac.) with best renovation on those plots which received maximum amounts of NH₄-N.

The obvious and expected losses of nitrogen from plots occurred in the NO₃-N form. Leaching losses ranged from 220 to 2211 kg/ha (196 to 1972 lbs/ac.). Again, both of these values occurred on the 300 series plots.

Again, the reversal of nitrate leaching with additions of water on the 200 series plots occurred. Almost 5 times as much nitrate was leached through 2.1 m (7 ft.) of the soil profile on the 200 plot without water application. Differences on the 100 plots are probably not significant. Quantities of nitrate applied to the plots are relatively insignificant in comparison to the total quantity of nitrogen applied. The dynamics of plot chemistry are exhibited by the rapid changes in form of nitrogen and the large leaching losses in the soil solution as nitrate.

The renovation capacity exhibited by the Everett soil for total P, PO₄-P, calcium, magnesium, sodium and potassium is excellent even in spite of the very large quantities applied. Renovation of phosphate was 99+% with a maximum leaching loss of 2 kg/ha (1.8 lb/ac.). Applications of calcium in sludge ranged from 31,000 to over 70,000 kg/ha (27,660 to over 62,450 lbs/ac.). Leaching losses ranged from 231 to 2594 kg/ha (207 to 2314 lbs/ac.); however, renovation was excellent in view of the very large quantities applied. Very similar trends were exhibited for the other base nutrients with almost insignificant leaching losses.

The greatly reduced rates of sludge application and quantities applied (Table 2) and dilution of sludge with water resulted in significantly reduced applications of nutrients during the second year (Table 4).

Large first year applications of sludge to the Everett soils initiated leaching losses in the soil solution through 2.1 m (7 ft.) of the soil profile which continued in ensuing years, maintaining similar patterns. Greatly reduced rates of weekly nutrient applications and reduced total loadings reduced but did not stem leaching losses. The percent renovation, in general, was less—not because of increased losses but in calculating percentages, the greatly reduced amounts of nutrients applied cause even a reduced flux to show generally a lesser percentage of renovation.

Total nitrogen losses (flux) for the 100 series plots were approximately one-half of the losses sustained during the first year (800 kg/ha, YR 1, to

Table 4. SECOND YEAR NUTRIENT LOADING AND FLUX (kg/ha) THROUGH THE EVERETT SOIL AND PERCENT CUMULATIVE RENOVATION

\forall	226	226	722	679	1,537	1,537
	17	106	15	16	16	9
	92	53	98	98	99	99
Na	398	398	1,270	1,193	2,703	2,703
	19	55	88	56	53	45
	95	86	93	95	98	98
Mg	431	431	1,374	1,292	2,926	2,926
	54	56	208	158	134	101
	87	87	85	88	95	97
Ca	2,055	2,055	6,559	6,164	13,960	13,960
	388	246	1,238	764	646	584
	81	88	81	88	95	96
i		12 1.7 86	37 T 99+	35 T 99+	80 T 99+	80 .1 99+
Total -P	51	51 T 99+	162 1.1 99+	152 T 99+	345 • 2 99+	345 T 99+
NO3-N	10 447	312	38 1,519	31 726 -	69 69 -	- 299 -
NH ₄ -N	70	70	224	211	477	477
	T*	T	T	2	T	T
	99+	99+	99+	99+	99+	99
Total -N	125 321 -	125 423 -	398 1,784	374 930	848 785 7	848 944 -
	Applied	Applied	Applied	Applied	Applied	Applied
	Flux	Flux	Flux	Flux	Flux	Flux
	% Ren.	% Ren.	% Ren.	% Ren.	% Ren.	% Ren.
Plots	100	100 + water	200	200 + water	300	300 + water

*T denotes trace levels

372 kg/ha, YR 2; or 713 lbs/ac., YR 1, to 332 lbs/ac., YR 2). Plot 200 sustained the continued high rate of total nitrogen loss initiated in YR 1 (1784 kg/ha or 1592 lbs/ac.). Plot 200W increased from 340 to 930 kg/ha (303 to 830 lbs/ac.). Plot 300 increased from 221 to 785 kg/ha (197 to 700 lbs/ac.) in the second year.

Again, NH₄-N losses were insignificant and nitrate leaching generally declined, particularly on the 300W plot, from an average of 2211 to 667 kg/ha (1973 to 595 lbs/ac.) in the second year. In the first year, nitrate leaching was the major form of loss of total nitrogen from the soils. However, in the second year substantial losses of nitrogen took place in other forms (as yet unidentified). Losses of total phosphate and PO₄-P continued to be insignigificant. Leaching of calcium generally declined on all plots except for plot 200 which showed a slight increase over first year leaching losses. The very high leaching loss on plot 300W, 2594 kg/ha (2314 lbs/ac.), reduced to 584 kg/ha (521 lbs/ac.) in the second year. Leaching losses of other nutrients were not significantly different in the first and second years. The reduced rates of application make even reduced rates of leaching a lesser value in terms of renovation.

Applications of sludge to the Everett soil plots in the third year continued at reduced rates (comparing plot averages of Tables 4 and 5). With the exception of plot 100, total nitrogen leaching generally declined significantly in the third year. The decline in total -N flux generally results from reduced nitrate leaching with the average for the 100 series reducing from 380 to 290 kg/ha (338 to 259 lbs/ac.). The 200 series reduced from an average of both plots of 1120 to 690 kg/ha (999 to 616 lbs/ac.), and the 300 series reduced from 650 to 470 kg/ha (580 to 419 lbs/ac.: Table 5).

Losses of total P and PO₄-P continued to be insignificant. Changes in losses of other base nutrients are probably not significantly different during year 3 as compared to year 2. In general, there either are no significant differences or slight declines in leaching losses.

The total of nutrients applied in three years of sludge application and total nutrient losses are summarized in Table 6. The most significant losses from the soil profile are the nitrate form of nitrogen. Of the 4743 kg/ha (4233 lbs/ac.) of total nitrogen lost from the 200 plot, 4041 kg/ha (3605 lbs/ac.)--85%--were in the nitrate -N form. Losses of NH₄-N are insignificant by comparison. Retention of phosphate in the soil profile is excellent in all of its forms with losses at background levels.

Large quantities of dissolved base nutrients in Metro sludge provide a very significant loading when compared to the total quantities of suspended organic solids applied. For example, the 200 series plots average 115,120 kg/ha (102,202 lbs/ac.) of combined calcium, magnesium, sodium and potassium in comparison with an average of 91,200 kg/ha (81,168 lbs/ac.) of suspended solids (Table 2). Slightly greater quantities, 122,910 kg/ha (109,033 lbs/ac.) were applied to the 300 series plots. Suspended solids in sludge averaged 99,650 kg/ha (88,399 lbs/ac.) for the 300 series plots. Leaching

Table 5. THIRD YEAR NUTRIENT LOADING AND FLUX (kg/ha) THROUGH THE EVERETT SOIL AND PERCENT CUMULATIVE RENOVATION

Y	348	348	415	377	551	441
	18	9	10	22	8	12
	95	9	98	94	99	98
Na	612	612	729	663	969	775
	19	33	66	37	53	71
	97	95	91	94	95	91
Mg	660	660	787	715	1,045	836
	62	34	134	44	107	118
	91	95	83	94	90	86
Ca	3,168	3,168	3,775	3,432	5,016	4,013
	426	176	904	265	681	752
	87	94	76	92	86	81
P04-P	102	102	121	111	162	129
	T	T	T	T	T	T
	99+	99+	99+	99+	99+	99+
Total -P	108	108	128	117	171	137
	T	T	T	T	T	T
	99	99+	99+	99+.	99+	99+
NO3-N	414	167 -	1,029	17 352 -	25 568 -	20 378 -
NH4-N	107	107	128	116	170	136
	T*	T	T	T	T	T
	99+	99+	99+	99+	99+	99+
Total -N	192 416 -	192 167 13	229 1,459	208	304 571 -	243 388 -
	Applied	Applied	Applied	Applied	Applied	Applied
	Flux	Flux	Flux	Flux	Flux	Flux
	% Ren.	% Ren.	% Ren.	% Ren.	% Ren.	% Ren.
Plots	100	100 + water	200	200 + water	300	300 + water

*T denotes trace levels

Table 6. TOTAL NUTRIENTS APPLIED (kg/ha), FLUX AND PERCENT TOTAL RENOVATION THROUGH EVERETT SOIL

Plots		Total -N	NH ₄ -N	NO3-N	Total -P	P04-P	Ca	Mg	Na	×
100	Applied Flux % Ren.	2,447 1,600 35	1,446 290 80	200	1,026 3 99+	314 2.1 99+	40,374 1,431 96	8,457 218 97	7,810 84 99	4,441 140 97
100 + water	Applied Flux % Ren.	2,245 1,328 41	1,324 1 99+	184 1,209	943 1 99+	295 2 99	37,026 1,032 97	7,691 252 97	7,163 186 97	4,074 146 96
200	Applied Flux % Ren.	4,888 4,743 3	2,887 2 99+	407	2,023 2 99+	558 .3 99+	80,637 3,250 96	16,882 687 96	15,600 293 98	8,871 37 99+
200 + water	Applied Flux % Ren.	4,336 1,623 63	2,561 31 99	356 1,389	1,796 1 99+	498 T* 99+	71,529 1,465 98	14,976 289 98	13,838 202 99	7,869 107 99
300	Applied Flux % Ren.	5,312 1,577 70	3,122 2 99+	435 1,427	2,208 T* 99+	632 .1 99+	87,604 1,558	18,342 297 98	16,949 150 99	9,638 31 99+
300 + water	Applied Flux % Ren.	4,541 3,665 19	2,666 13 99+	3,256		532 1,1 99+	74,884 3,930 95	15,680 861 95	14,489 312 98	8,239 460 94
Water	Applied Flux % Ren.	1.60	0.39 0.17 56	1.01	0.16	0.06	18.81 8.57 54	3.71 2.52 32	23.10 3.89 83	1.65 0.68 59
	*T denot	*T denotes trace	leve is							

losses of dissolved constituents are almost insignificant by comparison to the quantities applied. With the exception of nitrogen compounds, total renovation on the Everett soil ranges from 95 to 99%.

PROPERTIES OF EVERETT SOILS

Total chemical analyses of the Everett soil profile prior to sludge treatments identified the quantities of base nutrients per hectare. A balance sheet is presented in Table 7 where initial average total quantities of nutrients are compared with total quantities applied in sludge and nutrient flux through the soil profile. Post treatment soil chemistry would indicate accumulations of nutrients in the soil profile. In general, total amounts of nutrients show an increase following three years of sludge application; however, high statistical variation in Everett soils does not allow an exact balancing of the initial quantities of nutrients. This is particularly true with very high application rates of calcium added in the 200 and 300 series plots. Additional quantities of total calcium applied in sludge almost equalled the quantities determined in the forest soil (96.0 mt/ha, 43 t/ac. in pretreatment; 81.2 mt/ha, 36 t/ac. applied for the 300 plots).

The total nutrient flux through 60 cm (24 in.) of the Everett soils is shown as a percentage flux in Table 7. Highest percentage of base nutrient losses occurred with maximum calcium applications for the 300 series plots (2.2%). All other base nutrients—magnesium, sodium and potassium—had losses ranging from an insignificant amount to 0.8% (magnesium at the 200 level treatments). Renovation over three years for quantities of nutrients applied is good. Percentages are even more significant when calculated on total quantities of base nutrients in soil profiles.

Attempts to accurately account for initial quantities of base nutrients in the soil profile plus the quantities added by applied sludge achieved moderate success. For example, for the 100 series plots an average of 96 mt/ha (43 t/ac) of calcium existed in the surface 60 cm (24 in.) of the soil; 38.7 mt/ha (17 t/ac) were applied by sludge with a loss of 1.2 mt/ha (0.5 t/ac). Post treatment soil sampling calculated 121 mt/ha (54 t/ac) of total calcium. This is 12 mt/ha (5 t/ac) less than predicted by balancing input and outputs of calcium. It is, however, within the confidence limits for calcium determinations in soils which average + 25%. Calcium estimations in the 200 and 300 series plots for post treatment quantities are much lower (110.8 and 126.1 mt/ha; 49 and 56 t/ac), when a predicted quantity should be about 175 mt/ha (78 t/ac).

Results for magnesium were somewhat similar to the calcium results. The quantity of 70.8 mt/ha (32 t/ac) of magnesium is predicted for the 100 series plots. The 69.8 mt/ha (31 t/ac) calculated from post treatment soil sampling is amazingly close. In fact, less post treatment magnesium was found than

Table 7. TOTAL NUTRIENT BALANCE FOR PRE- AND POST SLUDGE APPLICATIONS AND TOTAL PERCENT FLUX ON THE EVERETT SOIL

Plots		Ca	Mg	Na	K
		·	(metric tons	per hectares)	
100	Pre	. 96.0	62.7	251.5	71.2
	Applied	38.7	8.1	7.5	4.3
	F1ux	1.2	0.2	0.1	0.1
	Post	121.0	69.8	290.3	65.1
i	% Flux	1.0	0.3	T	0.2
200	Pre	96.0	62.7	251.5	71.2
	Applied	76.1	15.9	14.7	8.4
1	Flux	2.4	0.4	0.2	- 0.1
	Post	110.8	56.9	167.9	52.8
٠	% Flux	2.1	0.8	0.1	Т
300	Pre	96.0	62.7	251.5	71.2
	Applied	81.2	17.0	15.7	8.9
•	Flux	2.7	0.6	0.2	0.2
	Post	126.1	83.4	296.0	82.3
	% Flux	2.2	0.7	0.1	0.3

predicted, except the 300 series plots balance very well (83.4 mt/ha, 37 t/ac.
found compared to 79.7 mt/ha, 36 t/ac. predicted).

Quantities of sodium found in post treatment plots exceeded the prediction for the 100 and 300 series and were significantly low for the 200 series plots. Total sodium was 290 mt/ha (129 t/ac.) following sludge application where only 259 mt/ha (115 t/ac.) are predicted. The 300 series plots had 296 mt/ha (132 t/ac.) following sludge applications, where 267 mt/ha (119 t/ac.) are predicted. The 200 series plots have almost 100 mt/ha (45 t/ac.) less than predicted. The variability in the sodium results are somewhat expected due to the very high variability in sodium concentrations in the soil plots. Quantities of sodium on the exchange capacity frequently had + 100% variation.

Potassium following sludge application was close to predicted quantities for the 300 series plots (80 mt/ha, 36 t/ac. predicted; 82.3 mt/ha, 37 t/ac. found). Both pre- and post treatment potassium levels were highly variable in the surface soils. Confidence limits ranged to 5 times the average quantities in pretreatments reducing to + 50% in post treatment sampling. Low estimation on 100 and 200 series plots may be attributed to natural variations.

Coarse alluvial soils are generally recognized as being highly variable in physical structure, texture, and chemical composition. The magnitude of the variability problem is amplified when large amounts of the soil matrix are larger than 2 mm (0.08 in.). Estimation of bulk density and accurate identification of the active soil fraction complicate the problems of estimation of a total base nutrient balance for pre- and post sludge applications.

EVERETT SOIL CHEMISTRY, FIRST AND SECOND YEARS

Detailed resampling of original plots for total chemistry for pre- and post treatment results of base nutrients is summarized in Table 7. The 2.1 m (7 ft.) depths were not resampled due to the extensive disturbance excavation to this depth would have caused to the surface of plots. Pretreatment soil analyses established average quantities of nutrient reserves in the soil. Quantities of nutrients applied in sludge should indicate the current total less plant uptake and flux.

Preliminary data compared pretreatment average analyses to post treatment averages for original Everett plots receiving sludge (Table 8). Confidence interval of the mean at a 95% confidence level is calculated on the statistics for either the pre- or post treatment, using maximum variability. The limited numbers of soil samples do not allow strict statistical analysis but are offered only as a relative comparison of variability of soil chemistry.

Applications of sludge during the first and second years have increased the variability for most soil properties analyses, as summarized in Table 8. Random sampling of plots prior to sludge treatment developed averages for pH, organic matter, total nitrogen and cation exchange capacity. These averages were determined for the forest litter layers (L), the surface soils termed the soil A horizons, and averages for soil B horizons, the 15 to 60 cm (5.9)

Table 8. SOIL CHEMICAL PROPERTIES FOLLOWING SLUDGE APPLICATIONS ON THE EVERETT SOIL SERIES AND 95% CONFIDENCE INTERVAL, SECOND YEAR

	Horizon	Pretreatment	Post Treatment
			·
Н	L	5.6 ± 0.1	4.8 ± 0.9
	Α	5.9 ± 0.6	5.0 ± 1.0
	В	5.9 ± 0.5	5.3 ± 0.8
Organic Matter	L	38.1 <u>+</u> 2.3	32.6 ± 11.6
(percent)	Α	11.3 ± 0.6	16.4 <u>+</u> 6.5
	В	3.4 <u>+</u> 0.2	4.3 <u>+</u> 2.0
Total Nitrogen	L	0.61 <u>+</u> 0.04	0.82 + 0.31
(percent)	Α	0.27 ± 0.02	0.29 ± 0.15
	В	0.11 <u>+</u> 0.01	0.09 ± 0.06
Cation Exchange Capacity	, L , '	50.9 ± 2.5	59.9 ± 11.3
(meq/100g 0.D. wt)	Α	28.0 + 1.0	43.5 + 7.2*
	В	15.2 <u>+</u> 0.6	18.4 <u>+</u> 6.6
Carbon:Nitrogen	L	36:1	23:1
(ratio)	А	24:1	33:1
	В	18:1	28:1

^{*}Significant difference

to 24 in.) depths. In general, the confidence limits are significantly less for pretreatment than post treatment data. For example, a confidence limit of + 0.1 pH unit was found in the pretreatment analyses for the forest litter layers (L). This has increased 9-fold to 0.9 pH units for the post treatment soil chemical analyses.

Organic matter in the litter layer has an increased variability from \pm 2.3% to \pm 11.6%. Increased variability induced by sludge treatments results from variations in sludge application rates to plots. Initial analysis of sampling attempted to establish soil property differences between plots with different rates of sludge application; however, inherent variability of soils was much too great. Thus, soils from all sludge treated plots are analyzed independent of rates of sludge application. The only significant change in soil chemistry is an increase in average cation exchange capacity of soil A horizons, increasing from 28.0 to 43.5 milli-equivalents per 100 grams of oven dry soil (meq/100g 0.D.), following two years of sludge application.

The following sections discuss briefly trends found in soil chemistry.

Soil pH

Soil pH is the negative logarithm of the hydrogen ion concentration. Soil pH values were converted to actual hydrogen ion concentrations for comparisons of increased acidity. Hydrogen ion concentrations increased from 26 to over 6000%. Increases in hydrogen ion concentrations are reported as a decrease in soil pH. Average soil pH decreased for the litter layers from 5.6 to 4.8. The large variation associated with post treatment analyses will not allow establishment as a significant difference (range 3.9 to 5.7). Assessment of only pretreatment results suggests there has been a real reduction in average pH of the forest litter layers due to sludge application. Additional soil sampling would be required to be sure of a statistically significant difference.

Soil Organic Matter

Average soil organic matter tends to decrease with sludge application in litter layers (38.1 to 32.6%) and increase slightly in the soil horizons (11.3 to 16.4%). Patterns of variability are similar in organic matter analyses as with pH determinations. There is a 5-fold increase in variability of organic material content of forest litter analyses and a 10-fold increase in variability of soil organic matter analyses in A and deeper horizons, following sludge treatments. Based on only the determination of pretreatment condition, we would suggest that there has been significant decreases in organic matter of the L layer and increases in organic matter in the soil. However, the high variability of the post treatment results does not allow statistically significant conclusions. The trends, however, are apparent and are supported by total organic carbon analyses and a theory of translocation of organic colloids deeper into the soil profile.

Using these average organic matter values, pretreatment soil analyses indicated an average of 123.7 mt/ha (55.2 t/ac.) of organic material (range

72.6 to 176 mt/ha or 32.4 to 78.8 t/ac.). Post treatment analyses show an average of 169 mt/ha (75.2 t/ac.) with a range of 96.6 to 318 mt/ha (43.1 to 142.0 t/ac.) for sludge treated plots.

<u>Total Nitrogen</u>

Percentages of total nitrogen tend to increase in the L layer (0.61 to 0.82%) and remain unchanged in soil horizons. Reductions in soil organic matter in the L layer suggest there should be a reduction in quantity of nitrogen also. Increases in organic matter in the surface soils (A horizon) suggest there should be an increase in the quantity of nitrogen in surface soils.

Cation Exchange Capacity

Cation exchange capacity of the A horizon of sludge treated plots is the only soil property to show a statistically significant (P<.05) change (28.0 to 43.5 meq/100 g O.D. soil). Increases are suggested for L and B horizons (L 50.9 to 59.9; B 15.2 to 18.4 meq/100g O.D. soil) with greatly increased variation since sludge application. A mineralogic analysis revealed no change in the type or quantity of clay minerals present in the surface soil layers; thus, changes in total exchangeable cations must be attributed to increases and/or the form of organic colloids.

Carbon-Nitrogen Ratio

The carbon-nitrogen ratio for pretreatment conditions on the Everett soils is about average for forest soils. Forest litter is relatively undecomposed; therefore, high in carbon. Advancing decomposition takes place with incorporation in the soil profile. Changes in the carbon-nitrogen ratio induced by sludge application reflect mainly a very significant increase in the quantities of total nitrogen in forest litter layers. Quantities of carbon in the organic matter layer have declined slightly; however, the significant reduction in the carbon-nitrogen ratio is explained by increased total nitrogen. Organic carbon is increasing in the soil profile in both A and B horizons without equivalent increases in total nitrogen content. The carbon-nitrogen ratio has increased in both the A and B soil horizons.

EVERETT SOILS, EXCHANGEABLE CHEMISTRY AND BASE SATURATION, SECOND YEAR

Applications of sludge over two years have significantly altered several exchangeable cation relationships (Table 9). With the exception of sodium, there is a general decline in the exchangeable cations in L layers. Calcium has reduced significantly from 12.8 to 9.8 meq/loog 0.D. Potassium has also reduced significantly from 0.6 to 0.2 meq/loog 0.D. while sodium has increased from 0.1 to 0.4 meq/loog 0.D. Magnesium shows a general trend of decline in soil horizons with a significant reduction in the B horizon (0.8 to 0.4 meq/loog 0.D.). Potassium indicates a significant decline in both the A and B horizons.

The sum of total exchangeable cations tends to decrease in the L layers exclusive of the hydrogen ion (14.9 to 11.7 meq/100g 0.D.). Thus, a decrease

Table 9. EXCHANGEABLE CATIONS AND PERCENT BASE SATURATION BY HORIZONS OF THE EVERETT SOIL, SECOND YEAR

Base Saturation %	Post		23+13	9 - 92	38+14
Ba Satura	Pre	r ·	31+3	31+2	56±3
Na	Post	:	0.4+0.1*	0.2+0.1	0.2+0.2
	Pre	oil	0.1±0.1	0.1+0.1	0.2±0.2 0.2±0.2
	Post	Milli-equivalents per 100 g 0.D. soil	0.2+0.1*	*1.0-1.0	0.1+0.04*
<u></u>	Pre	alents per l	0.6+0.1	0.3±0.04	0.4+0.1
Mg	Post	Milli-equiv	.4+0.3 1.3+0.2	0.8+0.2	0.4+0.1* 0.4+0.1
	Pre	,	1.4±0.3	1.1±0.2	0.8±0.1
-	Post		12.8+1.8 9.8+1.2*	7.8±1.1	3.4+0.4
Ca	Pre		12.8+1.8	7.3±1.5 7.8±1.1	2.8+1.0 3.4+0.4
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*Significant difference

in exchangeable cations accompanied by an increase in cation exchange capacity effectively reduces the base saturation. Changes in the quantities of exchangeable cations in soil horizons are relatively insignificant; however, again the large increase in cation exchange capacity, particularly of the A horizon, results in an effective decrease in percentage of base saturation.

EVERETT SOIL CHEMISTRY, THIRD YEAR

Third year resampling of the original 100, 200, 300 series plots sampled litter layers and the A and B soil horizons (15 to 60 cm or 6 to 24 in.). Post treatment soil properties following three years of sludge application are again compared with the pretreatment samples. The sampling scheme attempted to establish the depth within the soil column where significant changes in soil properties might be taking place. Had significant differences been found in soil B horizons, then deeper soil horizons would have been sampled. Few significant changes in soil chemistry deeper than the B horizon were found. Table 10 compares soil chemical properties following three years of sludge application.

Soil pH

The pH of soil litter layers continues to decline with additional sludge applications. pH is now significantly lower than the pretreatment pH (5.6 compared to 4.3). pH of the third year is not significantly lower than the second year due to the high variability in second year pH results. Soil A horizons have also declined significantly from the pretreatment (5.9 to 4.7). pH of soil B horizon averages lower but is not significantly different from either pretreatment or second year results. (Compare Table 10 with Table 8.)

Organic Matter

Forest litter layers (L) show no significant changes in organic matter. The trend of apparent decrease in the second year has been reversed in the third year. Values for the three sampling times show normal variation. A significant increase in the organic matter content of the soil A horizon has taken place, while the average percentage organic matter is less (15.2%) in the third year than in the second (16.4%). The significant increase is established by reduced variability in the random process of sampling the soils. The confidence limits for the third year were \pm 1% as compared to \pm 6% for the second year samples. These data would suggest that a significant increase actually took place in the organic matter content with two years of sludge application, but it was obscured by the inherent random variation. Organic matter content of soil B horizons exhibits normal variation with no significant differences.

Total Nitrogen

Total nitrogen concentration in forest litter layers has also increased significantly over the pretreatment condition (0.61 to .78%). The third year of treatment appears to have allowed soil nitrogen to also increase in concentration in soil A horizons (0.27 to .42%). Neither of these increases

Table 10. SOIL CHEMICAL PROPERTIES FOLLOWING SLUDGE APPLICATIONS ON THE EVERETT SOIL SERIES AND 95% CONFIDENCE INTERVAL, THIRD YEAR

	Horizon	Pretreatment	Post Treatment
pH	L	5.6 ± 0.1	4.3 ± 0.04*
	Α	5.9 ± 0.6	4.7 + 0.2*
	В	5.9 ± 0.5	5.5 ± 0.3
Organic	L	38.1 <u>+</u> 2.3	41.2 + 3.9
Matter (percent)	Α	11.3 ± 0.6	15.2 ± 1.0*
	В	3.4 <u>+</u> 0.2	3.1 ± 0.5
Total	L ·	0.61 ± 0.04	0.78 ± 0.10*
Nitrogen (percent)	A	0.27 ± 0.02	0,42 ± 0.07*
	В	0.11 <u>+</u> 0.01	0.11 ± 0.02
Cation Exchange	L	50.9 ± 2.5	61.5 ± 5.0*
Capacity (meq/100g O.D. wt)	Α	28.0 ± 1.0	43,0 + 4.0*
	В	15.2 ± 0.6	17.4 + 1.7
Carbon:Nitrogen	L	36:1	30:1
(ratio)	A	24:1	21:1
	В	18:1	16:1

^{*}Significant difference

are significantly different from the second year values. There have been no significant changes in the soil B horizons.

Cation Exchange Capacity

Soil cation exchange capacity has significantly increased in both the L layers and the soil A horizon (L, 50.9 to 61.5 meq/100 g 0.D.; and A, 28 to 43 meq/100g 0.D.). Increases in cation exchange capacity are attributed to the increased organic matter content of the L layer and the A horizon.

Carbon-Nitrogen Ratio

The carbon-nitrogen ratio is a function of both changes in organic matter content and changes in the total nitrogen percent. In the second year, highly variable litter layers were sampled with a reduced organic matter content. Total nitrogen content was even higher (0.82%) than in the third year This combination gave a greatly reduced carbon-nitrogen ratio in the litter layers in the second year (23:1; Table 8). Increased organic matter in the third year, along with a stable nitrogen content, results in an increase in the carbon-nitrogen ratio of the litter layer (30:1). The combination of a significant increase in total nitrogen in the A horizon with a slight decrease in organic matter results in a drastic reduction in the carbon-nitrogen ratio over the second year (33:1 to 21:1). In a like manner, the slight increase in nitrogen content of the B horizon accompanied by a reduction in organic matter results in a significantly lower carbon-nitrogen ratio (28:1 to 16:1) in the third year of the study. While significant changes have taken place between the second and third years, the carbonnitrogen ratios currently are not significantly different from the pretreatment values.

EVERETT SOILS, EXCHANGEABLE CHEMISTRY AND BASE SATURATION, THIRD YEAR

The reduction in soil pH accompanied by an increase in organic matter and an increase in cation exchange capacity is generally reflected in a changing composition of exchangeable cations in the third year of sludge applications (Table 11). The initial decline of exchangeable calcium in L layers has stabilized at a slightly lower but not significantly decreased level. Increased amounts of calcium are becoming associated with the exchange capacity in the A horizon. In the third year of the study, exchangeable calcium almost doubled, resulting in a significant increase over pretreatment amounts. A trend to increase in calcium also occurs in the soil B horizon.

Exchangeable magnesium in the soil B horizon continues to decline accompanied by a nonsignificant increase in the A horizon. Exchangeable potassium increased significantly between the second and third years in the L layer and soil A horizon. Current values, however, are not significantly different from pretreatment values.

Exchangeable sodium continues to be significantly greater with both the L layer and A horizon showing an increase over pretreatment values. Increases are not significantly greater from the third year over the second.

Table 11. EXCHANGEABLE CATIONS AND PERCENT BASE SATURATION BY HORIZONS OF THE EVERETT SOIL, THIRD YEAR

Base Saturation %	Pre Post		21+3	35+3	33+11
Batura	Pre	F.	31+3	31+2	56+3
Na	Post	•	0.7±0.1 0.1±0.1 0.3±0.1*	0.3+0.1*	0.2+0.1
	Pre	soil	0.1+0.1	0.1±0.1	0.2±0.2 0.2±0.1
∠	Post	00 g 0.D.		0.4+0.1	0.4+0.1
	Pre	lents per 1	0.6±0.1	0.3±0.04	0.4+0.1
Mg.	Post	Milli-equivalents per 100 g 0.D. soil	1.4±0.3 1.3±0.4	1.2+0.2	0.1+0.1* 0.4+0.1
	Pre	Æ	1.4+0.3	1.1±0.2 1.2±0.2	0.8+0.1
ğ	Post		10.9+1.9	13.5-1.8*	4.5+1.2
Ca	Pre		12.8+1.8 10.9+1.9	7.3+1.5	2.8+1.0
nozin	<u>oH</u>			А	മ

*Significant difference

The accumulation of exchangeable cations in the soil A horizon has resulted in a significant increase in the percent base saturation between second and third year results. There are no significant changes between preand post treatment data or other horizons between the second and third years.

EVERETT SOILS, REDUCED SLUDGE APPLICATIONS

Application of sludge to the new plot series at reduced rates resulted in very similar nutrient flux and renovation during the first year of reduced sludge application rates. Table 12 provides a summary by nutrients of the quantities applied, losses and percent renovation by plots. Generally, losses of nutrients (in kg/ha) were quantitatively least on the 10 plots (minimum levels of sludge application) and generally greatest on the 30 and 40 plots which had increased sludge applications. Complex interactions remain in processes affecting total nitrogen leaching. Losses of NH4-N are insignificant, again with rapid change of form of total nitrogen to NO3-N where leaching losses occur in the soil solution.

Phosphorus losses again are insignificant with base nutrients closely paralleling results for heavier applications of nutrients. Renovation ranged from 96 to 99% with losses varying up to 613 kg/ha (547 lbs/ac.)--calcium-- on the 40 plots. Maximum concentrations of nutrients in the soil solution occur during dry summer periods when volumes of soil solution are very low and losses are really insignificant. Minimum concentrations of nutrients in the soil solutions occur during the wet season and actually account for the major portion of nutrient losses.

EVERETT SOIL CHEMISTRY, REDUCED SLUDGE

Analyses of impacts of reduced rates of sludge application for the first year on the new Everett plots are summarized in Table 13. Again, average of analyses of all sludge treated plots is compared with the pretreatment soils analyses to assess significant impacts of sludge applications on soil chemistry.

Soil pH

There was a significant decrease in the average pH (5.8 to 4.4) of the forest litter layers following first year applications of sludge at reduced rates. Reductions in the soil A horizon (5.3 to 4.7) were not significant, but it may be anticipated that a significant reduction in soil A horizon pH will take place, as a marked decline has occurred. Large variability in pretreatment pH of soil B horizons precludes any significant change caused by sludge treatments.

Soil Organic Matter

The series of plots selected for application of sludge at reduced rates had very large pretreatment variations in quantities of organic matter in both the forest litter layers (45.9 \pm 17%) and the soil A horizon (17.7 \pm 4.9%). No significant changes in soil organic matter have taken place in the

Table 12. TOTAL NUTRIENT LOADING AND FLUX (kg/ha) THROUGH THE EVERETT SOIL (REDUCED RATES)

×	591 12 98	1,800 30 98	2,369 73 97	2,683 39
Na	1,040	3,167	4,166	4,718
	31	58	99	107
	97	98	98	98
Mg	1,122	3,415	4,493	5,088
	40	62	109	104
	96	98	98	98
Š	5,385	16,394	21,569	24,420
	213	429	565	613
	96	97	97	97
P04-P	173.	527	694	786.3
	T	T	T	T
	99+	99+	99+	99+
Total -P	183	558	735	833
	T	T	T	1
	99+	99+	99+	99+
N03-N	26 179 -	81 394	107	121 448
NH4-N	182	555	731	828
	T*	T	T	6
	99+	99+	99+	99+
Total -N	326 180 45	993	1,307 696 47	1,480 454 70
	Applied Flux Ren.	Applied Flux % Ren.	Applied Flux % Ren.	Applied Flux % Ren.
Plots	10	50	30	40

*T denotes trace levels

Table 13. AVERAGE SOIL CHEMICAL PROPERTIES OF EVERETT (New Plots) SOILS WITH REDUCED RATES OF SLUDGE APPLICATION

	Horizon	Pretreatment	Post Treatment
рН	L	5.8 ± 0.6	4.4 + 0.3*
	А	5.3 ± 0.2	4.7 ± 0.7
	В	5.6 ± 1.3	5.8 ± 0.2
Organic	L ,	45.9 + 17.0	42.3 + 9.0
Matter (percent)	А	17.7 + 4.9	15.9 + 5.0
(percent)	В	1.8 ± 1.3	2.9 ± 0.2
Total	L .	0.66 + 0.43	0.64 + 0.36
Nitrogen (percent)	А	0.24 + 0.04	0.29 ± 0.02
(per cerro)	В	0.05 ± 0.03	0.07 ± 0.01
Cation Exchange	L	42.4 + 6.4	73.1 + 8.6*
Capacity (meq/100g 0.D.wt)	А	29.2 + 3.0	38.7 + 3.5*
(med/100g 0.b.wt)	В	15.3 ± 1.3	15.4 ± 1.3
Carbon:Nitrogen	. L	40:1	38:1
(ratio)	A	43:1	32:1
	В	21:1	24:1

^{*}Significant difference

soil surface, although there is possibly a trend to increase in the soil B horizon.

Total Nitrogen

Total nitrogen content may be increasing in the soil horizons but not significantly so following one year of sludge applications. Large variation in total nitrogen usually accompanies large variation in organic matter. Total nitrogen variation is very similar to that of organic matter with no significant differences in forest litter layers.

Cation Exchange Capacity

Cation exchange capacity of L layers and surface soil (A) horizons continues to be the soil property most significantly influenced by sludge applications. Litter layers of sludge treated plots increased an average of 30.7 meq/100g 0.D. (from 42.4 to 73.1 meq/100g 0.D.). Cation exchange capacity of soil A horizons has increased significantly as a result of sludge treatments; 38.7 as compared to 29.2 meq/100g 0.D. soil. Soil B horizons largely are unaltered to date by sludge applications.

Carbon-Nitrogen Ratio

Carbon-nitrogen ratios have not departed significantly as a result of sludge treatments in L and B layers. The reduced organic matter with slightly increased nitrogen has lowered the carbon-nitrogen ratio of the A horizon following sludge treatment.

EVERETT SOILS, EXCHANGEABLE CHEMISTRY AND BASE SATURATION, REDUCED RATES

Few significant changes occurred as a result of sludge application at reduced rates on the Everett soil (Table 14). The only statistically significant change in exchangeable chemistry was an increase in the sodium content of the B horizon from 0.1 to 0.3 meq/100g 0.D. It is probably more interesting to note the pretreatment values for the new Everett plots as compared to the pretreatment values for the old Everett plots. (See Table 9 and 11). Exchangeable calcium is significantly higher in both the L and A horizons. In a like manner, exchangeable magnesium is very close to being significantly greater.

Larger quantities of exchangeable cations for the particular location of the new plots result in a base saturation percent significantly greater than reported for the older Everett soil plots. The significant increases in cation exchange capacity of the L and A horizons result in a significantly reduced base saturation percentage following sludge treatments.

Table 14. EXCHANGEABLE CATIONS AND PERCENT BASE SATURATION BY HORIZONS FOR PRE- AND POST SLUDGE APPLICATIONS FOR THE EVERETT SOIL (REDUCED RATES)

se tion %	Post		28+4	31+7	35+7
Base Saturation %	Pre		48+4	52+4	31+3
Na	Post		0.7±0.2	0.4+0.2	0.3±0.03*
	Pre	. soil	0.2-0.3	0.1+0.2	0.1±0.03
	Post	. 100 g O.D	1.0+0.2	0.5+0.2	0.4+0.1
	Pre	Milli-equivalents per 100 g 0.D. soil	2.1-0.6 2.5-0.6 0.4-2.6 1.0-0.2 0.2-0.3	1.1±0.4 0.3±1.6	1.4±0.1 0.8±0.3 0.2±0.7
	Post	Λilli-equi∿	2.5+0.6		0.8+0.3
Mg	Pre		2.1-0.6	1.5+0.3	0.4+0.1
į	Post		16.2+4.7	6.9+3.1	4.2+1.0
Са	Pre		17.8+1.9	13.4+3.8	4.1+0.5
nozia	OH			А	മ

*Significant difference

EVERETT SOIL SOLUTION NUTRIENTS, WATER AND CONTROL PLOTS

Quantities of nutrients applied in the water irrigation treatment over the three years of study, and the natural flux of nutrients from a control plot, indicate the natural levels of nutrient cycling (Table 15). The untreated control plot lost 1.6 kg/ha (1.41 lbs./ac.) of total -N composed predominantly of NO₃-N (0.9 kg/ha or 0.8 lb./ac.) or 56%. NH₄-N losses were about equal (0.2 kg/ha control; 0.17 kg/ha water only--0.17 and 0.15 lb./ac., respectively) on the two plots. Inputs of total N by water irrigation are relatively small; however, irrigation with water does accelerate total nitrogen losses (2.87 kg/ha water--2.56 lbs./ac.; 1.6 kg/ha control--1.4 lbs./ac.). Water applications apparently stimulate microbiological activity to the point of greatly reducing nitrate losses (0.08 kg/ha or 0.07 lb./ac.) on the water only plots.

Losses of total phosphorus were accelerated through irrigation with water as compared with the control. However, fractions of kg/ha loss are probably insignificant. Phosphate losses were reduced in the water only plot.

The flux of calcium, magnesium, sodium and potassium are slightly greater on the control plots than on the water irrigated plots. These losses are probably within the range of normal point to point variation for the Everett soil. In general, there is a conservation of base nutrient cations within the soil profile for the quantities applied in the irrigation water.

Table 15. NUTRIENTS APPLIED IN WATER AND FLUX (kg/ha) FROM CONTROL AND WATER ONLY PLOT

								-			
			Total -N	NH ₄ -N	N0 ³ -N	Total -P PO_4 -P	P04-P	င္မွ	Mg	Na	~
71	Control	Flux	1.6	0.2	6.0	0.2	0.2	12.8	2.8	5.1	0.8
	Water	Applied	1.60	0.39	1.01	0.16	90.0	18.8	3.7	23.1	1.65
	on ly	Flux	2.87	.17	.08	.47	Ξ.	8.6	2.5	4.0	.68
		% Ren.	1	56	92	4	1	54	32	83	59

EVERETT SOILS--SOIL SOLUTION NUTRIENT CONCENTRATIONS

An important phase of interpretation of the impacts of sludge disposal in the forest is nutrient concentrations transmitted to ground water. A major portion of the renovation of nutrients should take place in surface soil horizons, 2-3m (6-9 ft.). Measurement of the maximum and minimum concentrations of nutrients passing the 2.1 m (7 ft.) depth in the soil should provide an index of maximum potential contributions of sludge application to ground water.

A summary of the maximum and minimum concentrations of nutrients for all sludge treatment plots on Everett soils is provided in Table 16. Interpretation of the meanings of average concentrations could be very misleading, as they would be more an indices of the frequency of sampling rather than a true average concentration. Minimum concentrations of nutrients in the soil solution occur during rainy periods when high quantities of water are passing the soil profile. In these circumstances, large volumes of water are collected by the lysimeters and may be collected at frequent intervals. Averaging of these data would provide an unrealistic estimation of average nutrient concentrations.

Conversely, during very dry periods, particularly with low rates of sludge application, the soils are very dry with low volumes of soil solution extracted by the lysimeter plates. Frequently, field collections were made at 2 to 3-week or longer intervals—the time period required for an extraction of an adequate volume of soil solution for chemical analyses. These dry soil samples yielded the highest concentrations of nutrients in the soil solution. Samples would be relatively infrequent, even though concentration numbers would be large.

Generally, higher concentrations of nutrients were found in the 100, 200 and 300 series plots without the additions of additional water. As per the previous discussion, the 100W, 200W and 300W series plots with water following sludge application provided an additional dilution factor. Potassium values apparently deviate from this pattern.

TOTAL NITROGEN

Maximum concentrations of total N in the soil solution occurred on the 200 series plots. Both total N and NO₃-N exceeded 400 mg/l. In general, sludge applications followed by water and the reduced sludge application rates resulted in much lower total N and NO₃-N concentrations in soil solution.

Table 16. MINIMUM-MAXIMUM CONCENTRATIONS OF NUTRIENTS IN THE SOIL SOLUTION AT 2.1 m (7 ft.), EVERETT SOILS

Plots	Total -N	N0 ₃ -N	Ca	Mg	Na	К
		(mi	lligrams per	liter)		•
100	2 - 299	2 - 298	10 - 270	1 - 37	1 - 11	1. + 38
200	2 - 466	*T - 439	20 - 328	3 - 66	4 - 44	1 - 5
300	3 - 223	T - 222	3 - 278	1 - 38	2 - 19	1 - 5
WOOI	2 - 156	1 - 155	3 - 148	1 - 31	4 - 16	1 - 61.
200W	. 2 - 144	2 - 156	5 - 171	2 - 43	5 - 15	1 - 12
300W	T - 250	T - 247	25 - 236	3 - 41	9 - 19	2 - 14
10	1 - 50	T - 49	8 - 45	3 - 10	4 - 7	1 - 3
20	3 - 136	3 - 135	10 - 127	1 - 17	1 - 7	4 - 12
30	2 - 103	1 - 101	9 - 81	1 - 30	3 - 13	2 - 12
40	T - 131	T - 128	6 - 178	2 - 23	2 - 8	4 - 16
Control	T - 13	T - 12	1 - 24	T - 3	T - 6	1 - 68
Water	T - 14	T - 12	1 - 19	1 - 21	1 - 9	T - 4

^{*}Denotes less than 0.5 mg/l

Reduced rates of sludge applications (10 to 40 plot series) resulted in reduced total N and NO_3 -N in soil solutions.

BASE NUTRIENTS

Patterns of concentrations of base nutrients in the soil solution are similar to that for total N and NO $_3$ -N. Maximum concentrations of Ca, Mg and Na occurred on the 100, 200, 300 series plots with the maximum values occurring on the 200 plots. Applications of water following sludge resulted in increased minimum values of Ca and Na for the 300W plot.

Reduced rates of sludge application with the 10, 20, 30, 40 series plots resulted in significantly reduced concentrations of these nutrients in the soil solution. Patterns of base nutrient occurrence in the soil solution suggest a complex interaction between soils and applied sludge. Consistent patterns of increasing quantities of nutrients in the soil solution with increasing rates of sludge application are not evident for many nutrients.

Table 16 also compares maximum and minimum concentrations of nutrients in the soil solution for the control and water only plots. Water irrigation does not seem to accelerate nutrient losses nor contribute particularly to excessive quantities of nitrogen or base nutrients in the soil solution.

EVERETT SOILS, SOIL SOLUTION CONDUCTIVITY, pH AND ALKALINITY

Ranges of conductivity, pH and alkalinity of soil solutions by plot treatments and ground water are summarized in Table 17. Conductivity (specific conductance measured in micromhos per square centimeter) of sludge treated plots has maximum values on the 100, 200, 300 series. Minimum values are also higher for these sludge treatments. The control plot, however, has the maximum recorded conductivity (3427 micromhos) while both the control and water plots have the lowest minimum values recorded. The wide range in maximum conductivities for other plots suggests that sludge treatments are not adding more dissolved ions to the soil solution, as treated plots do not equal the control. There is an indication, however, that minimum values on the average are increased with sludge treatments. Ground water tends to buffer at a higher minimum and a lower maximum (142-194 micromhos) conductivity.

Patterns of soil solution pH are similar to nutrient concentrations and conductivities, with the 100, 200, 300 series plots showing increased maximum and reduced minimum values (range 4.0 to 8.4). The 100W, 200W, 300W series generally have high maximum pH values (average 7.9), but the low pH is a whole pH unit higher than sludge only series.

The control plot has a very narrow range of pH (7.0 to 7.4) indicating that sludge treatments are both depressing low pH values and increasing maximum pH values. Plots with reduced rates of sludge application appear to have lower minimum values, but on the average, maximum values have not been increased. Ground water ranges from 7.3 to 8.1 with an average value of 7.8.

Table 17. CONDUCTIVITY, pH, AND ALKALINITY OF THE SOIL SOLUTION AND GROUND WATER FOR EVERETT PLOTS

Plots	Cond. μ mhos	рН	Alka. mg/l
100	62 - 1856	4.0 - 8.4	0.1 - 8.8
200	175 - 2293	6.1 - 7.7	0.2 - 1.0
300	138 - 1536	5.5 - 7.6	0.1 - 0.4
100W	53 - 906	5.8 - 7.8	0.2 - 1.0
200W	52 - 970	5.0 - 8.0	0.1 - 2.3
300W	85 - 1087	6.7 - 7.9	0.5 - 1.0
10	132 - 1060	6.2 - 7.1	0.1 - 0.2
20	85 - 1013	6.0 - 7.2	0.1 - 0.3
30	102 - 368	6.5 - 7.4	0.2 - 0.6
40	71 - 974	6.9 - 7.4	0.3 - 0.5
Water	41 - 101	6.3 - 7.5	0.2 - 0.5
Control	45 - 3427	7.0 - 7.4	0.2 - 0.7
Ground Wate	r 142 - 194	7.3 - 8.1	1.0 - 1.7
X	171	7.8	1.3

Patterns of soil solution alkalinity are difficult to interpret. In general, it appears that sludge applications have had little impact on either maximum or minimum values. Two plots are exceptions to this statement—the 100 plot and 200W. In both cases, the maximums are significantly higher than other values, particularly for the water or control plots (0.5 and 0.7 mg/l, respectively).

The results of use of the Metro-data loggers for rapid scan of the concentration of dissolved oxygen, solution pH, conductivity and temperature resulted in complex computer plots. Examples of these plots are shown in figures in Appendix B.

EVERETT SOILS, GROUND WATER NUTRIENTS

A well drilled to the interface of the outwash gravel and lacustrine deposits provided sampling of ground water adjacent to original Everett plot. Dissolved chemicals of ground water are summarized in Table 18. Total N averaged 0.8 mg/l with a range of 0.5 to 1.4 mg/l. Other forms of nitrogen, NH4-N and NO3-N, averaged 0.2 and 0.4 mg/l with no consistent relationship between the two nitrogen forms. The maximum NH4-N concentration was 0.7 mg/l while the NO3-N maximum concentration was 0.6 mg/l.

Forms of phosphorus are practically nonexistent in the ground water. A maximum of 0.18 mg/l occurred on one occasion as PO₄-P. Total P averaged 0.06 mg/l and PO₄-P averaged 0.03 mg/l. Concentrations of base nutrients are relatively stable in ground water showing slight seasonal variation. Anions of SO₄ and Cl exhibit considerably more variation with SO₄ ranging from 3 to 25 mg/l, averaging 11 mg/l. The range for Cl was 1 to 45 mg/l, with an average of 11 mg/l. The range of TOC was also large--4 to 40 mg/l.

Table 18. DISSOLVED CHEMISTRY OF GROUND WATER, EVERETT SOILS

Date	Date Total -N NH ₄ -N	NH4-N	N03-N	Total -P	PO4~P	Na	×	Ca	Mg	S0 ₄	T0C	5
					(mg/1)						•	
2-26-76	9.0	H	က်	H	—	7	3.4	19	വ	9	40	45
4-13-76	8:0	.7	4.	H	.01	16	2.3	17	က	13	9	15
92-3-76	1.3	4.		.05	-	13	2.4	16	2	14	i	∞
7-29-76	1.4	4.	9.	.07	-	9	1.8	Ε	2	16		9
8-17-76		-	ഹ	.03	H	1	2.6	15	ស	, I	i	ı
11-9-76	0.7	•04	9.	90.	-	6	2.1	14	4	œ	t	,
1-9-1	1.0	.2	4.	.04	.01	7	2.1	20	9	က	က	7
3-31-77	9.0		4.	90.	90.	.	2.1	19	ဖ	1	4	9
6-13-77	0.5	-	.2	.04	-	∞	1.6	17	ဖ .	4	10	9
9-12-77	9.0	 	ლ.	8	.18	9	2.2	20	7	25	1,	7
Ave.	0.8	.2	4.	90.	.03	6	2.3	15	വ	 	13	

MASHEL SOILS, SOIL SOLUTION NUTRIENTS

Experimental designs parallel to the Everett soil studies evaluated impacts of sludge applications at reduced rates to Mashel and Wilkeson soils. Design rates of application were not achieved as sludge was only applied for 7 months. Table 19 indicates the design rate (by plot number) and actual rate of sludge applications with nutrient loading, flux (kg/ha) through the Mashel soil, and percent renovation achieved. Patterns of flux and renovation are very similar to reduced rates of sludge applications on the Everett soil.

The characteristic wetness of the Mashel site was obvious immediately after irrigation periods. Ponding occurred in depressions and would slowly infiltrate over a period of days. Ponding was more noticeable after a total 15 tons of solids were applied on a plot.

The lateral movement of subsurface soil water was greater than vertical soil water flux on sloping Mashel soils. Calculation of nutrient renovation was more difficult without exact measurements of downward water flux.

NUTRIENT RENOVATION FOR THE MASHEL SOIL SERIES

Table 19 shows the nutrient loading, flux and percent renovation for the four rates of sludge applications to Mashel soils. As seen with the Everett plots, renovation of total P, PO₄-P, and cations Ca, Mg, Na and K is excellent. Total N flux increases with increasing applications of sludge and is again dominated by the NO_3 -N losses.

Renovation of total N varied from 46 to 82% with increased percent renovation on 40 treatment plots (82%). Losses of NH₄-N were greatest with the 20 and 30 series plots (39 and 59 kg/ha or 35 to 53 lbs/ac.). Losses of total N are dominated by NO₃-N leaching (65 to 80%). Other forms of nitrogen loss are operative though unexplained at this time, as up to 35% of total N losses are unaccounted for by the combination of NH₄-N and NO₃-N flux.

Excellent renovation (99+%) of phosphorus compounds continue with a maximum of 5 kg/ha (4.5 lbs/ac.) leached on the 30 series plots.

In a like manner, renovation of base nutrients continues to be very efficient with 93 to 99+% of applied base nutrients retained in the surface soil. Maximum losses occurred on the 20 and 30 series plots where total base

Table 19. NUTRIENT LOADING AND FLUX (kg/ha) THROUGH
THE MASHEL SOIL AND PERCENT RENOVATION
(APPLIED NUTRIENTS IN MT/HA AND FLUX OF
NUTRIENTS IN KG/HA)

H	1			•
~	435	942	1,394	.1,885
	8	23	30	17
	98	97	98	99+
Na	765	1,657	2,453	3,315
	31	117	59	46
	96	93	98	99+
Mg	825	1,787	2,645	3,575
	33	52	59	22
	96	97	98	99
Ca	3,960	8,580	12,698	17,160
	84	144	211	113
	98	98	98	99+
P0 ₄ -P	127 · T	276 T 99+	408 5 99	552 T 99+
Total -P	135	292	432	585
	T	1	5	T
	99+	99+	99	99+
N0 ₃ -N	19 84	42	63 116 -	85 151 -
NH ₄ ~N	134	290	430	581
	T	39	59	11
	99+	86		98
Total -N	240	520	769	1,040
	129	152	175	185
	46	71	77	82
	Applied	Applied	Applied	Applied
	Flux	Flux	Flux	Flux
	% Ren.	% Ren.	% Ren.	% Ren.
Total Plots Applied (mt/ha)	4.8	10.6	15.4	20.8
Plots	10	20	30	40
		80		

nutrient flux was 336 kg/ha (300 lbs/ac.) for 20 plots and 359 kg/ha (320 lbs/ac.) for 30 plots.

MASHEL SOIL PROPERTIES

Usual forest soil sampling problems with large inherent natural spatial variation in soil properties were even greater on plots on Mashel soils. Highly variable soil chemistry of pretreatment soils analyses, particularly soil organic matter and total nitrogen, made identification of significant changes in soil properties due to sludge application difficult. Again, all sludge treated plots were combined for test comparisons with pretreatment or control conditions (Table 20). The Mashel soil is inherently more acid, contains less organic matter, has lower average total N and cation exchange capacity than Everett soils.

SOIL pH

Sludge applications caused no significant changes in pH of the Mashel soil. A trend seems to indicate reduced pH in the L layer and increased pH (4.8 to 5.2) in the B horizon at 15 to 60 cm (6 to 24 in.) of depth; however, pre- and post treatment results broadly overlap.

SOIL ORGANIC MATTER

A nonsignificant trend to increased soil organic matter and organic matter of L layers exists. However, as previously noted, the very high natural variability of organic matter in the Mashel soils prevents establishment of significant differences from first year sludge treatments.

TOTAL NITROGEN

A trend to reducing total nitrogen in the L layers and increasing total nitrogen in soil horizons is suggested. Again, the highly variable results from pretreatment total N analyses preclude establishment of significant changes in total soil nitrogen of the Mashel soils.

CATION EXCHANGE CAPACITY

Statistically significant increases in cation exchange capacity were found in L (14.8 to 61.9 meq/100g 0.D.) and A (6.9 to 23.1 meq/100g 0.D.) soil layers following sludge treatments. Increases in soil organic matter may explain a portion of the increase in cation exchange capacity.

Table 20. SOIL CHEMICAL PROPERTIES FOLLOWING SLUDGE APPLICATIONS ON THE MASHEL SOIL SERIES

	Horizon	Pretreatment	Post Treatment		
pH	L	4.9 <u>+</u> 0.6	4.6 <u>+</u> 0.4		
	Α	4.7 ± 0.9	4.7 <u>+</u> 0.3		
	В	4.8 + 1.3	5.2 ± 0.5		
Organic	L	26.0 <u>+</u> 27.9	35.0 <u>+</u> 16.1		
Matter (percent)	А	3.4 + 8.4	7.7 ± 4.3		
	В	0.4 + 1.2	1.8 <u>+</u> 0.7		
Total Nitrogen (percent)	L	0.49 <u>+</u> 0.56	0.29 <u>+</u> 0.12		
	Α	0.06 ± 0.04	0.14 ± 0.09		
	В	0.03 ± 0.02	0.05 ± 0.01		
Cation Exchange	L	14.8 <u>+</u> 8.8	61.9 <u>+</u> 15.6*		
Capacity (meq/100g O.D. wt)	А	6.9 ± 3.5	23.1 ± 5.1*		
	В	6.2 <u>+</u> 6.3	9.3 <u>+</u> 1.4		
Carbon:Nitrogen	L	31:1	70:1		
(ratio)	А	33:1	32:1		
•	В	8:1	21:1*		

^{*}Significant difference

CARBON-NITROGEN RATIO

An increase in organic matter content of the B horizon, without equivalent increases in total N, caused a significant increase in the carbon-nitrogen ratio (8:1 to 21:1). No significant change occurred in L or A layers. The highly variable organic matter content of L layers combined with equally variable amounts of total N make even the 70:1 post treatment value an insignificant difference.

MASHEL SOILS, EXCHANGEABLE CATIONS AND BASE SATURATION

Analyses of impacts of sludge applications on exchangeable soil chemistry and base saturation are summarized in Table 21. Applications of sludge have either stabilized the large natural variability in exchangeable calcium throughout the soil profile, or an improved random sample was obtained for post treatment. Highly variable pretreatment Ca analyses span the range of post treatment analyses. However, it appears that the large quantities of dissolved Ca in Metro sludge have increased exchangeable Ca throughout the soil. These cannot be claimed as significant increases at this time but a trend is indicated.

A significant increase in exchangeable Mg occurred in the L layer and B horizon. A trend to increased Mg also is indicated in the A horizon.

There are no apparent trends in exchangeable K resulting from sludge applications. A significant increase similar to Mg occurred with exchangeable Na, as increased Na occurred in the L layer with a definite trend to increase in the A horizon. Again, a smaller but significant increase occurred deeper in the soil profile of the B horizon. The large significant increase in exchangeable Na in the L layer (0.1 to 0.7 meq/100g 0.D. soil) and trend to increased base nutrients in soil horizons suggest a substantial base nutrient leaching in the Mashel soils.

High variability in pretreatment base saturation analyses allows no conclusion as to impacts of sludge on the Mashel soil at this time.

Table 21. EXCHANGEABLE CATIONS AND PERCENT BASE SATURATION BY HORIZONS FOR PRE- AND POST SLUDGE APPLICATIONS FOR THE MASHEL SOIL

se cion %	Post	(percent)	28+ 7	33+10	50+64
Base Saturation %	Pre	(per	35±35	52+67	35+20
	Post		0.7+0.3*	0.3±0.1	0.3+0.1*
Na	Pre	soil.	0.1+0.1	0.4+0.2 0.1+0.2	0.1±0.04
×	Post	Milli-equivalents per 100 g 0.D. soil	.2+0.6 2.6+1.2* 1.0+2.1 0.6+0.2 0.1+0.1		0.2+0.4
	Pre	alents per	1.0-2.1	0.5+1.1	0.2±0.6
Mg	Post	4illi∽equiv	2.6+1.2*	1.3+0.8	1.5±0.2*
Σ	Pre		0.2±0.6	0.8+0.4	0.9±0.3
	Post		13.7±2.6	6.2+1.6	3.0±0.3
Са	Pre		4.9±22.3	2.3+12.3	1.9± 9.0 3.0±0.3
uozţ	чон			A	മ

*Significant difference

WILKESON SOILS--SOIL SOLUTION NUTRIENTS

Similar analyses were made for interpretation of the impacts of reduced rates of sludge application on Wilkeson soils. These results are summarized in Tables 22, 23 and 24 which analyze nutrient loading and flux, pre- and post treatment soil chemistry, and analysis of exchangeable base cations.

Renovation of total nitrogen ranged from 69 to 89% (on plots 30 and 10, respectively; table 22). Losses of NH₄-N from the 30 series plots also are large (43 kg/ha; 38 lbs/ac.) on the Wilkeson soil. Losses of NO₃-N follow usual patterns (41 to 260 kg/ha; 37 to 232 lbs/ac.). However, a larger percentage (80 to 99%) of total -N loss is in the NO₃-N form.

Usual patterns with phosphorus compounds retention prevail with insignificant leaching or flux. Base nutrients--calcium, magnesium, sodium and potassium--also exhibit usual patterns with leaching losses varying from 6% (sodium, 44 kg/ha; 39 lbs/ac.) to less than 1% (17 kg/ha; 15 lbs/ac.) of magnesium. In general, the patterns of renovation for the three soil series tested are very similar. Maximum total flux of base nutrients (548 kg/ha; 489 lbs/ac.) was associated with maximum sludge applications (40 plots).

WILKESON SOILS--SOIL PROPERTIES

The pre- and post treatment analyses of soil chemical properties are summarized in Table 23.

Soil pH

The Wilkeson soil is naturally significantly more acid (pH 3.6 to 4.5) than either the Everett (pH 5.3 to 5.8) or the Mashel (pH 4.7 to 4.9). Sludge treatments induced a significant increase in pH on the Wilkeson soil (pH 3.6 to 5.6) in the L layers and surface soil horizons (pH 3.5 to 5.0). A trend to increasing pH is also indicated at greater depth than the soil profile.

Soil Organic Matter

A marked trend to increased organic matter of L layers also occurred on the Wilkeson soil (23.3 to 44.8%). A trend to increase soil organic matter appears in the A horizon but is insignificant.

Table 22. NUTRIENT LOADING AND FLUX (kg/ha) THROUGH THE WILKESON SOIL AND PERCENT RENOVATION

⊻	435	15	97	904	27	26	1,290	42	97	1,667	66	94
N N	765	44	94	1,591	32	86	2,228	51	86	2,932	72	86
Mg	825	15	86	1,701	17	+66	2,889	39	66	3,162	99	
g _.	3,960	78	86	8,236	84	66	11,536	222	86	15,180	311	86
P0 ₄ -P	127	F	+66	268	 	+66	376		+66	488	F	+66
Total -P	135	-	+66	280	⊢	+66	393	}	+66	517	_	+66
N03-N	19	41	1	40	54	ı	27	173	ı	75	260	ţ
NH4-N	134	က	26	279	,	+66	391	43	68	514	52	95
Total -N	240	44	88	499	52	88	669	216	69	920	285	70
	Applied	Flux	% Ren.	Applied	Flux	% Ren.	Applied	Flux	% Ren.	Applied	Flux	% Ren.
Total Plots Tons Applied	4.8			9.9	ŧ		14.1	٠		18.6		
Plots	10			20			30			40		

Table 23. SOIL CHEMICAL PROPERTIES FOLLOWING SLUDGE APPLICATIONS ON THE WILKESON SOIL SERIES AND 95% CONFIDENCE INTERVAL

	· Horizon	Pretreatment	Post Treatment		
рН	L	3.6 <u>+</u> 0.5	5.6 ± 0.6*		
	А	3.5 ± 0.4	5.0 <u>+</u> 0.6*		
	В	4.5 <u>+</u> 0.7	5.0 ± 0.2		
Organic	L	23.3 + 16.7	44.8 + 8.2		
Matter (percent)	А	13.6 ± 4.9	15.4 <u>+</u> 6.1		
	В	1.3 ± 0.7	1.7 <u>+</u> 0.4		
Total	L	0.36 ± 0.06	0.38 ± 0.04		
Nitrogen (percent)	Α	0.12 <u>+</u> 0.08	0.20 ± 0.06		
	В	0.05 ± 0.02	0.04 ± 0.01		
Cation Exchange	L .	20.3 ± 14.6	77.4 + 10.3*		
Capacity (meq/100g O.D. wt)	Α	13.4 + 6.3	36.9 <u>+</u> 12.8*		
	В	3.8 <u>+</u> 4.1	12.6 ± 0.3*		
Çarbon:Nitrogen	L	38:1	68:1*		
(ratio)	Α	66:1	45:1*		
	В	15:1	20:1		

^{*}Significant difference

Soil Nitrogen

Trends to increases in soil organic matter on the Wilkeson soil following sludge treatment suggests a quantitative increase in total nitrogen. A significant increase in the concentration of total nitrogen was not found. A trend to increasing total soil nitrogen in the A horizon is indicated (0.12% pretreatment to 0.20% post treatment). An increase in concentration of this magnitude in the soils would yield a significant quantitative increase in total nitrogen per hectare following sludge treatments.

Cation Exchange Capacity

Very significant increases in cation exchange capacity in all horizons resulted from sludge treatments on the Wilkeson soil. L layers increased from 20.3 to 77.4 meq/100g 0.D. soil. Surface soil horizons (A) increased from 13.4 to 36.9 meq/100g 0.D. soil. At greater soil depths, exchange capacity also increased significantly averaging 12.6 meq/100g 0.D. soil.

Carbon-Nitrogen Ratio

The carbon-nitrogen ratio of L layers significantly increased with applications of sludge. Large increases in organic matter content of the L layer, without an equivalent increase in total nitrogen, resulted in a marked increase in the quantity of carbon (38:1 to 68:1). In contrast, soil A horizons significantly decreased in carbon-nitrogen ratio, caused by a marked increase in total N in the A horizons without an equivalent increase in the percentage of organic carbon (change was 66:1 to 45:1). No significant change occurred in the carbon-nitrogen ratio of the B horizon.

WILKESON SOILS, EXCHANGEABLE CATIONS AND BASE SATURATION

The large quantity of dissolved calcium in Metro sludge again causes a significant increase in exchangeable calcium in the L layers of the Wilkeson soil (7.1 to 24.2 meq 100/g 0.D. soil; Table 24). Average increases occur in both A horizons and deeper soil horizons but are not significant. General increases also occur in exchangeable magnesium throughout the soil layers but are not of sufficient magnitude to be statistically significant. Potassium results are more variable in the surface soil but also tend to be increasing.

Sufficient amounts of sodium have been applied in sludge and have leached through soil profiles to add a significant amount to the cation exchangeable complex. In the L layers, exchangeable sodium has increased from 0.01 to 0.9 meq/l00g 0.D. soil. A horizons have increased from 0.1 to 0.5 meq/l00g 0.D. soil. The quantities of sodium leaching to greater depths have also been sufficient to provide a significant increase in exchangeable sodium in the 15 to 60 cm (6 to 24 in.) soil layers (0.1 to 0.3 meq/l00g 0.D. soil). Impacts of sludge application on the percentage of base saturation are variable and inconclusive at this time.

Table 24. EXCHANGEABLE CATIONS AND PERCENT BASE SATURATION BY HORIZONS OF THE WILKESON SOIL

Base Saturation %	Pre Post		45± 3 39± 5	65±14 48±18	31-33 60- 6
	Post		0.9+0.2*	0.1 ±0.1 0.5±0.2* 6	
Na	Pre	-soil	0.01-0.1	0.1 +0.1	0.1 ±0.1 0.3±0.2*
	Post	Milli-equivalents per 100 g 0.D. soil	0.5±0.6 1.2±0.1	6.0-6.0	*1.0-9.0
	Pre	alents per	9.0+2.0	0.9+0.2	0.1±0.1
Mg	Post	Milli-equiv	1.6±0.9 4.0±1.0*	1.7±1.0 2.1±0.7	0.9±0.9 1.9±0.3
	Pre		1.6±0.9	1.7±1.0	6.0-6.0
Ga	Post		7.1±6.1 24.2±5.0*	12.9+6.6	1.7±3.2 4.7±0.9
	Pre		7.1±6.1	A 6.5±3.5 12.9±6.6	1.7±3.2
			_	A	മ

90

*Significant difference

MASHEL AND WILKESON SOILS, SOIL SOLUTION NUTRIENT CONCENTRATION

Concentrations of nutrients in the soil solution of Mashel and Wilkeson soils expressed as mg/l are summarized in Table 25. Minimum values for total N are higher than those reported for the Everett soil; however, maximum concentration of total N tends to be less. The flux of total N is again dominated by NO₃-N losses. Concentrations of total N and NO₃-N are not significantly different in the two soils. Minimum values for retention of Ca are lower in the Mashel than the Wilkeson. Maximum values are also higher for the Wilkeson soil. In general, the Wilkeson appears to have an increased capacity to retain cations (Tables 21 and 24), even though all base nutrients in the soil solution have generally increased minimum values when compared to the Mashel soils. The significant increase in a cation exchange capacity (Table 23) is not yet reflecting in increased quantities of cations retained in the Wilkeson soil. There is little difference in the two soils in maximum concentrations for Mg and Na; however, maximum and minimum concentrations of K are significantly greater for all treatments in the Wilkeson soil.

Table 25. RANGE IN NUTRIENT CONCENTRATIONS OF THE SOIL SOLUTION BY PLOTS FOR THE MASHEL AND WILKESON SOILS

Soils & Plot	Total -N	NO ₃ -N	Ca	Mg	Na	К
		(1	milligrams p	er liter)	T	
Mashe1					•	
10	7 - 55	7 - 54	10 - 52	4 - 18	5 - 19	1 - 5
20	11 - 55	11 - 54	9 - 58	3 - 19	8 - 31	1 - 8
30	5 - 124	4 - 124	10 - 146	5 - 34	9 - 27	4 - 15
40	5 - 127	3 - 126	7 - 71	3 - 10	8 16	1 - 16
					- - - -	
Wilkeson					* *	
10	6 - 73	6 - 73	23 - 93	6 - 16	17 - 29	4 - 18
20	6 - 52	6 - 52	15 - 74	4 - 13	9 - 18	7 - 23
30	31 - 82	30 - 82	34 - 105	8 - 18	14 - 23	13 - 20
40	9 - 128	8 - 128	18 - 158	6 - 31	12 - 27	16 - 39

EVERETT SOIL, TOTAL ORGANIC CARBON

Study of total organic carbon (TOC) leaching by the soil solution was limited to plots on the Everett soils. Analyses of TOC are presented in three phases: First, the migration of TOC in soil solution through the soil profile, where organic constituents are not identified—only quantitatively related to depths in the soil profile.

A second phase of analysis identifies the molecular size fractions leached by the use of gel filtration chromatography. Identification of sizes allows evaluation of physical filtering by the soil matrix and preliminary indications of rate of leaching by molecular size.

The final analysis identifies specific organic compounds identifying their source. TOC analyses are confined to organic matter of colloidal size or smaller. Soil solution samples collected from L, A, B and C horizons of control, water only, and various sludge plots are analyzed for TOC to provide an overview of effects of treatment on amounts of organics leaching through forest soils.

TOTAL ORGANIC CARBON LEACHING

Leaching of TOC through soil profiles may be best summarized in a series of figures showing quantitative amounts of TOC with depth (Fig. 11). The water only and control plots are compared with the 100, 200 and 300 plot sludge treatments. A very significant increase in the quantity and depth of leaching of TOC occurs with sludge applications. Maximum transport of TOC through the soil profile occurred with the 200 series plots. A size fraction is apparently transported to the C horizon in the 200 series plots, as both the 100 and 300 series plots show a continuous decline in the quantity of TOC through the C horizon.

The control and water only plots have much lower quantities of TOC throughout the soil profile. Water applications have increased TOC in both the A and B horizons.

Leaching of TOC in the 100W, 200W and 300W plots significantly departed the patterns established with sludge only plot (Fig. 12). Generally, reduced amounts of TOC occur in the litter layers, with even greater reduction in the soil A and B horizons. Large increases are generally identified in the C horizons with a maximum in the 300W (70 mg/l) which is nearly equivalent to the accumulation noted on the 200 plot (Fig. 11).

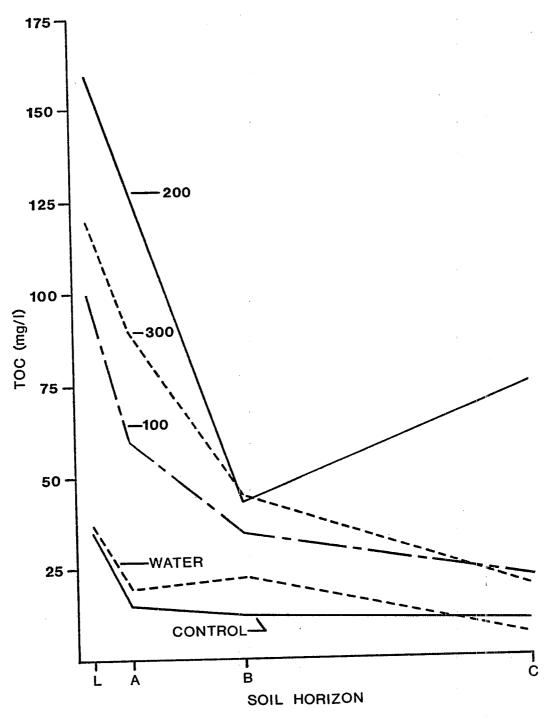


Figure 11. Average TOC values versus depth for water, control, and sludge treatments.

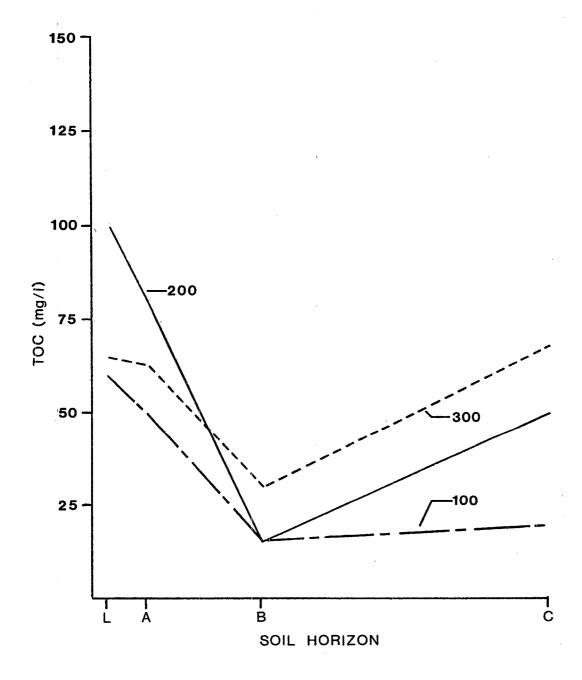


Figure 12. Average TOC versus soil depth for the sludge followed by water sites.

An example of changes in TOC over time with maximum sludge applications (300 series plot) is shown in Figure 13. Sampling is identified for the L, A, B and C horizons. Very significant reductions in TOC in the L layers occur with continued applications of sludge (250 mg/l pretreatment to 20 to 40 mg/l after 20 months of sludge applications).

Soil A and B horizons exhibit a similar early pattern of reduction in TOC declining from 165 mg/l in the A and 135 mg/l in the B to less than 30 mg/l in 20 months. Soil B horizons exhibit a somewhat inconsistent pattern of accumulation and depletion seasonally, while C horizons show irregular variations over time.

In general, the summer season has increased TOC with rapid decline in the winter and increases in the following summers. Patterns of movement were not consistent within the soil profile for equivalent times.

Gel Filtration Chromatography

Preliminary examination of the chemical nature of organic materials responsible for TOC values in the soil solution was done by fractionating the organic material according to molecular size using gel filtration chromatography. Detection by UV spectrometry at 254 mm identifies aromatic and other unsaturated structures.

Gel filtration chromatograms of the soil solution from several soil horizons in the control plot and sludge plots are compared in Figure 14. Litter samples from the control plot show a large quantity of higher molecular weight organic molecules (early peak on left), which is missing from the sludge plots (increased late peaks). The C horizon sample from the sludge plot contains a larger amount of low molecular weight organic materials.

Chromatograms for 1977 soil solutions are shown in Figure 15. The C horizon of the 300 series plot shows two large peaks which are not very evident in the C layer of the 10 series plots. These peaks tend to coincide with natural organic material in the L layer, indicating that sludge applications over time leach organic material in the soil solution to greater depths in the soil profile. The slight peaks in the well water coincide. These peaks were also identified in soil solution samples from the control plots (Figure 14, control C), indicating they are not organic contaminants from sludge application—only natural organic compounds leaching to ground water.

IDENTIFICATION OF SPECIFIC ORGANICS

Soil solutions from B and C horizons of 300 series sludge plots were composited, freeze dried, extracted with diethylether, methylated with diazomethane and analyzed by gas chromatography on a diethylene glycol succinate (DEGS) column. Organics could not be identified, other than phthalate esters—a common contaminant in this system and in trace work in general. An alternative method examined organics concentrated on soil particles. Soil samples were collected by horizons, extracted with moist ether and processed

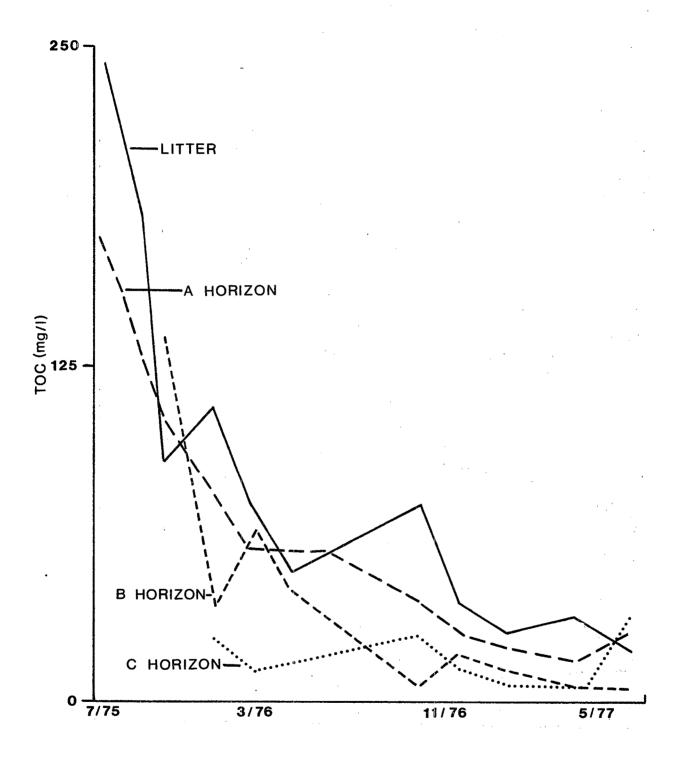


Figure 13. TOC versus date of sample collection for the $300\ \mathrm{plot}$ from the four soil depths.

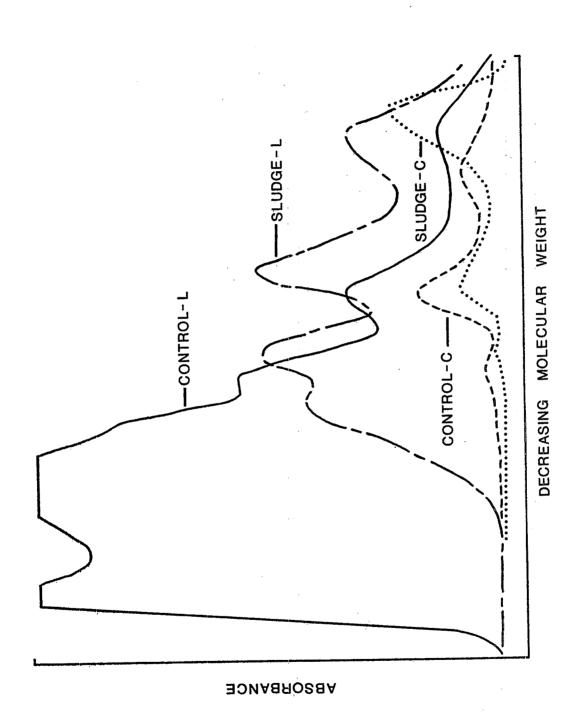
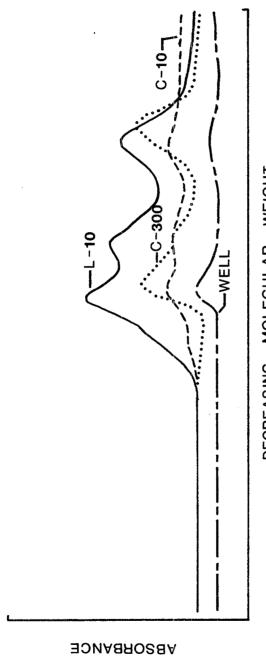


Figure 14. L layer and C horizon soil solutions from sludge control plots compared by gel filtration chromatography.



DECREASING MOLECULAR WEIGHT

Figure 15. Gel filtration chromatograms of soil solution samples from L and C layers of 10 series plots, C horizon of 300 plot and well water.

by gas chromatography. The A horizon chromatograms from the 300 series sludge plot and control show very similar patterns (Fig. 16). After three years of sludge application, there are no major organic inputs from sludge detectable by gas chromatography.

In a similar manner, B horizon chromatograms are nearly identical except for an increase in compounds labeled H, I, and J in the 300 series sludge plots. A compound labeled X (Fig. 17) is present in small quantities on the sludge plot and not shown in the control plot. A further check on the organic constituents in control and sludge treated plots was tested by gas chromatography using ether extraction and a DEXSIL 300 column. This sampling compared gas chromatograms of ether extracts of the B horizon of the 100 series sludge plots with a similar extract from the B horizon from the control plot (Figs. 18 and 19).

Chromatograms show sludge and control plots qualitatively contain the same organic compounds, again indicating little or no movement of ether extractable organics from sludge through the soil profile.

Mass spectral analysis of the gas chromatograph effluents allowed the following identification to be made:

<u>Peak</u>	Molecular <u>Weight</u>	<u>Identify</u>
Α		Unknown
В	256	Pentadecanoic acid methyl ester
С	270	Palmitic acid methyl ester
D		Unknown
E		Unknown
F	316	Isopimaric acid methyl ester
G	314	Dehydroabietic acid methyl ester
Н	354	Diacosanoic acid methyl ester
I	368	Triacosanoic acid methyl ester
Ĵ	382	Lignoceric acid methyl ester

Fatty acids B, C, H, I and J have been reported previously as organic materials found in peat and various soils (Braids and Miller, 1975).

This is the first identification of isopimaric and dehydroabietic acids in forest soils; however, these organics are common oleoresins and undoubtedly form from foliage and litter decomposition.

The identified compounds are all either natural products (isopimaric, etc.) or have been previously reported in soil; therefore, the sludge applications introduce no major new organic compounds to the system.

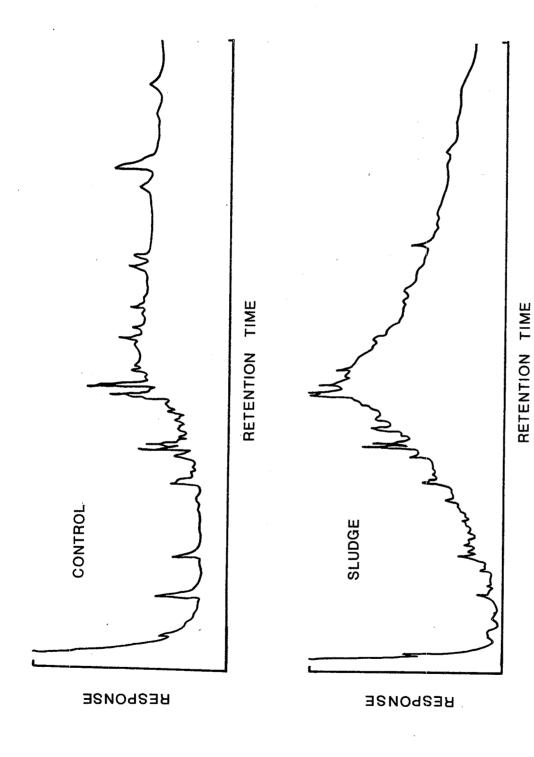


Figure 16. Gas chromatogram of ether extract of soil from 300 and control plot A horizon using a DEGS column.

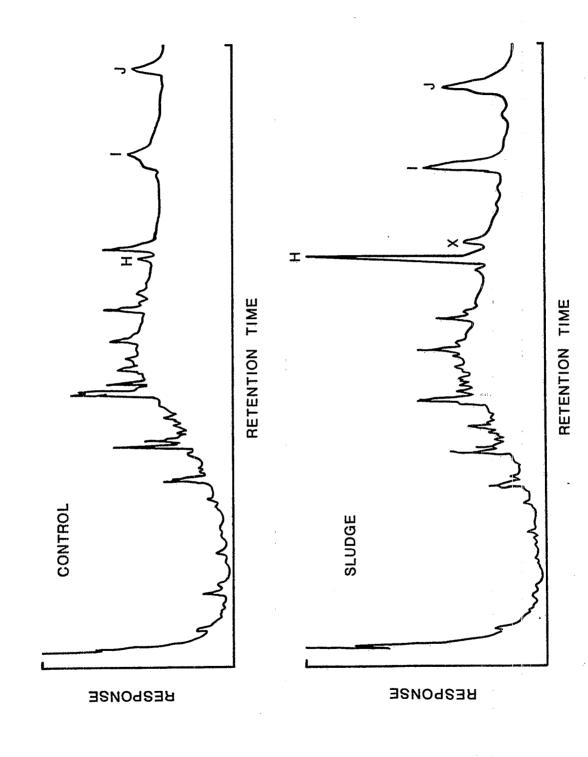


Figure 17. Gas chromatogram of ether extract of soil from 300 and control plot B horizon using a DEGS column.

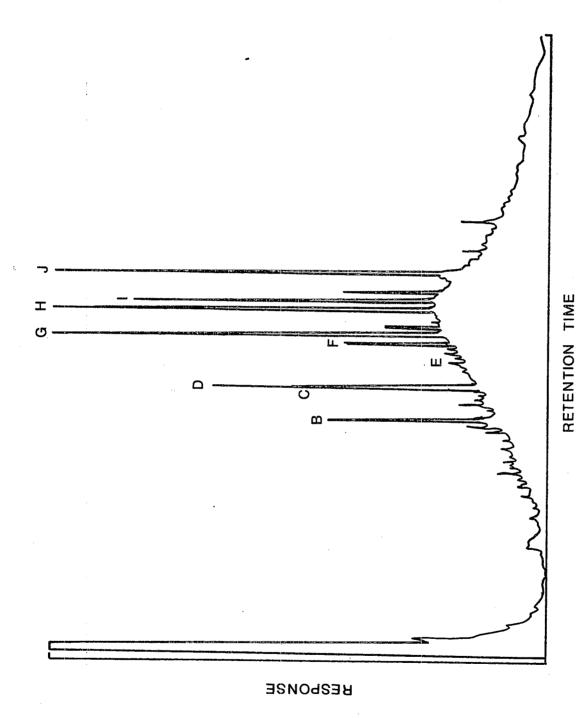


Figure 18. Gas chromatogram of ether extract of 100 B horizon using a DEXSIL 300 column.

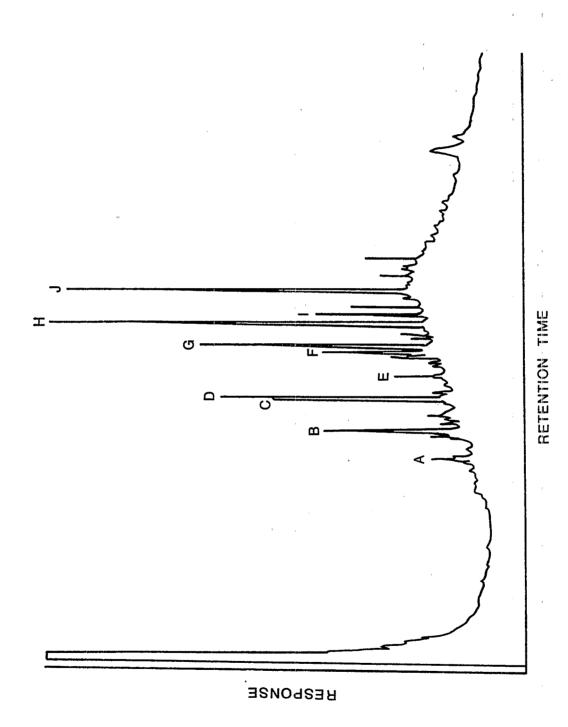


Figure 19. Gas chromatogram of ether extract of the control plot B horizon using a DEXSIL 300 column.

EVERETT SOIL, BIOLOGICAL DECOMPOSITION

Sludge applications in the forest environment potentially may alter rates of biological decomposition of naturally occurring plant residues, lignin and cellulose materials. Study of impacts of sludge applications on biological decomposition rates investigated soda solubility and weight loss of wood blocks. Results of decomposition studies are summarized by type of analyses with trends over time in the following sections.

SODA SOLUBILITY

Soda solubility is defined by the weight loss of decaying cellulose as extracted by a 1% sodium hydroxide solution (1% NaOH). Changes in soda solubility as examined by this extraction and tested against standard samples must be interpreted. As standard wood block samples are used as a base and extracted weight subtracted from the test samples, a gain in weight shows as a negative or decrease in relative soda solubility. A loss in weight shows up as a plus or increase in soda solubility. Increases in soda solubility represent increasing decomposition, as summarized in Figure 20 and Table 26.

Decomposition on Litter Surface

All sludge treated plots decreased (13 to 24%) rapidly in relative soda solubility in the first 30 days (Sept. to Oct., 1975) of exposure (Fig. 20, Table 26). Declines (1.9 to 2.5%) in soda solubility were generally maintained in all treated plots until May 1976. An exception to this general trend is the 100 series plots where there was an increase (12.6 to 13.7%) in soda solubility which occurred between October and February. This increase was followed by a slight decline (1.2%) between February and May.

The control sample did not differ significantly in soda solubility from the standard between October and May.

Decomposition in all plots increased significantly between May and July. Soda solubility increased 3% for the control plots; 2.8%, 1.6% and 3.8% for the 100, 300 and 300W plots, respectively.

Resampling a year later in July 1977 again showed significant increases in decomposition as indexed by increased soda solubility. Different rates of sludge application did not significantly alter decomposition rates on the surface of the litter layers for the wood blocks. Soda soluble weight increases were 5.3%, 7.6%, 7.9% and 6.9% for the control, 100, 300 and 300W plots, respectively.

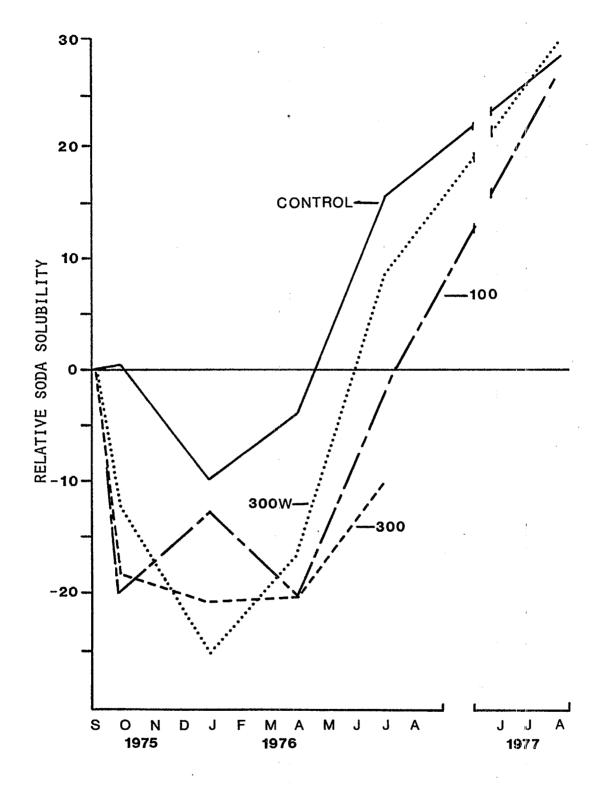


Figure 20. Changes in 1% NaOH solubility of wood blocks on top of the litter over time by sludge treatments.

Table 26. DECOMPOSITION OVER TIME ASSESSED BY SODA SOLUBILITY EXTRACTIONS (PERCENTAGE OF DRY WEIGHT) OF WOOD BLOCKS BY VARIOUS DEPTHS AND TREATMENTS₁,

:		. Samp	oling Dates		
Plots and Horizons	<u>Oct. '75</u>	Feb.	May	July '76	July '77
Control Plot					
L	15.7 <u>+</u> 0.9	14.1 <u>+</u> 0.5	15.0 <u>+</u> 0.2	18.0 <u>+</u> 0.5	23.3 <u>+</u> 1.7
Α	14.6 <u>+</u> 1.4	15.5 <u>+</u> 0.2	13.2 <u>+</u> 0.3	14.9 <u>+</u> 0.2	22.3 <u>+</u> 0.4
В	15.3 <u>+</u> 0.8	16.1 <u>+</u> 0.5	15.4 <u>+</u> 0.1	14.3 <u>+</u> 0.2	<u> </u>
100 Plots					
L	12.6+0.7	13.7+0.2	12.5 <u>+</u> 0.5	15.3 <u>+</u> 0.0	22.9+0.3
Α	12.1 <u>+</u> 0.7	13.1 <u>+</u> 0.5	12.1 <u>+</u> 0.2	15.5 <u>+</u> 0.1	22.7 <u>+</u> 0.0
В	12.6+0.3	14.4+0.3	14.0±0.1	16.9 <u>+</u> 0.2	
300 Plots					
L	12.8+0.6	12.4+0.4	12.5+0.3	14.1+0.2	22.02/
A	12.3+0.3	11.1+0.3	12.5 <u>+</u> 0.0	14.9 <u>+</u> 0.2	24.0 <u>+</u> 0.2
В	13.5 <u>+</u> 0.3	11.8 <u>+</u> 0.1	12.5 <u>+</u> 0.7	15.5 <u>+</u> 0.0	
, :					
300W Plots					
L .	13.7+0.0	11.7+0.2	13.1 <u>+</u> 0.3	16.9 <u>+</u> 0.4	23.8+1.3
A	13.6+0.2	11.7+0.2	13.1±0.5	16.6±0.0	23.0 ₂ / °
В :	15.3 <u>+</u> 0.5	14.4 <u>+</u> 0.0	13.5 <u>+</u> 0.2	17.0 <u>+</u> 0.0	

^{1/} Standard sample solubility 15.6±0.3% (September 1975)

 $[\]underline{2}$ / Estimated from A or B horizon samples

^{3/} Samples lost from B horizons for last sampling date.

Decomposition in A Horizons

Patterns of soda solubilities in A horizons were similar for all sludge treated plots. Decline in soda solubility (interpreted as a resistance to decomposition) occurred for all sludge treatments between September 1975, when the experiment was initiated, and the following May (1976).

Wood blocks in the control plot showed increased (0.9%) soda solubility in soil A horizons between October and February, with a decline (2.3%) between February and May. The control and all sludge treated plots showed significant increases in decomposition between May and July. Increases were: control, 1.7%; 100 plots, 3.4%; 300 plots, 2.4%; and 300W plots, 3.5%.

Sampling in July 1977 again revealed marked increases in decomposition. Increased decomposition did not vary significantly between plots and was equivalent to decomposition at the surface of the litter layer. Values were: control, 7.4%; 100 plots, 7.2%; and 300 plots, 9.1%.

Decomposition in B Horizons

In general, patterns of decomposition in the soil B horizons were very similar to the litter surface and A horizons. Only the 300W plot showed no significant decline in soda solubility in the first month of exposure. Between October and May, both the 300 and 300W plots showed a decline (13.5 to 12.5% and 15.3 to 13.5%) in soda solubility.

The 100 plots declined 3% in the first month of exposure (Sept. through Oct.). A significant increase in soda solubility occurred between October and February; however, no significant change took place between February and May. Patterns of increased soda solubility for treatment plots were very similar in A and B soil horizons. Soda solubility increased 2.9%, 100 plots; 3.0%, 300 plots; and 3.5%, 300W plots from May to July 1976.

The control plot showed a decline (15.4 to 14.3%) in soda solubility between May and July.

WEIGHT LOSS DECOMPOSITION

Assessment of decomposition by comparison of net weight loss with standard wood block samples provides another index of impacts of sludge applications on forest decomposition rates. In general, the control and 100 plots responded in a very similar fashion if weight loss is a guide to rates of decomposition. The 300 and 300W plots also responded in a very similar fashion; however, the response of the two sets of plots was significantly different. Results of weight losses over time are summarized in Table 27 and shown graphically for the litter surface in Figure 21.

Decomposition on Litter Surface

The first 30 days of exposure resulted in a weight loss of 4.6% and 3.4% for the control and 100 plots (Table 27). In the September through October

Table 27. DECOMPOSITION OVER TIME ASSESSED BY WEIGHT LOSS (PERCENT OF THE DRY WEIGHT) OF WOOD BLOCKS BY DEPTHS AND TREATMENTS

		S	ampling Date	٠,	
Plots and Horizons	Oct. '75	Feb.	May	July '76	July '77
Control Plot		(percer	nt net weight	loss)	
L	4.6	7.5	2.7	15.3	29.7
Α	3.8	7.9	2.7	17.9	25.2
В	5.5	9.6	3.8	16.3	3/
100 Plots					
L	3.4	6.7	1.8	14.3	22.7
Α	4.7	7.2	3.0	16.1	22.9
В	3.5	10.6	5.1	17.5	34 C
300 Plots		÷			
L	+0.11/	2.3	0.02	9.3	16.02/
Α	0.4	2.0	+0.51/	11.6	23.0
В	1.3	3.1	0.8	11.3	dis day
300W Plots					
L	0.9	2.6	0.3	11.4	18.9
Α	1.4	3.0	0.7	12.6	22.02/
В	3.3	6.5	3.8	13.8	quat dess

^{1/}Weight gain noticed

 $[\]underline{2}$ /Estimated from A or B horizon samples

^{3/}Samples lost from B horizons for last sampling date.

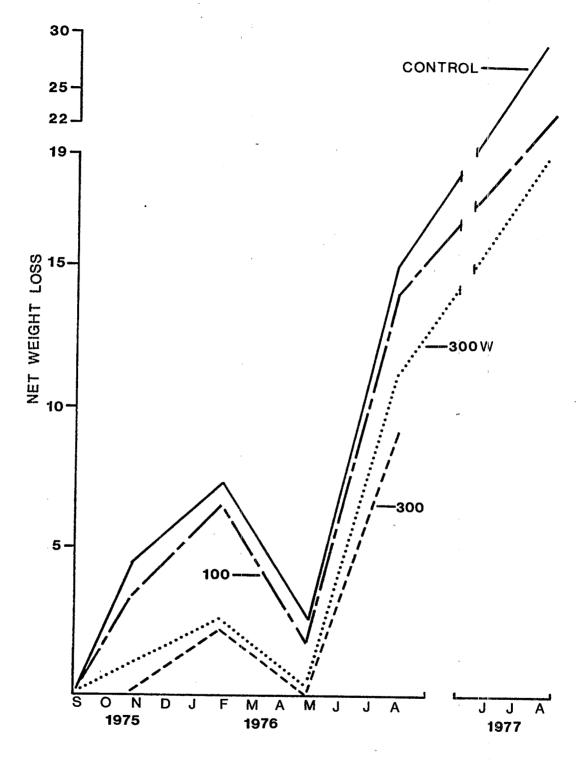


Figure 21. Changes in net weight of wood blocks on top of the litter over time by sludge treatments.

exposure, wood blocks in the 300 plot gained 0.1% in weight, while 300W lost 0.9% in weight. During the winter from October to February, decomposition continued in both the control and 100 plots (2.9% and 3.3%) and was initiated in the 300 and 300W plots (2.4% and 1.7%).

Decomposition was evidently arrested between February and May as all plots showed an increase in net dry weight of exposed wood blocks. The consistency of this phenomena suggests rapid growth of fungal hyphae with translocation of metabolic products into the wood blocks. Large net weight losses between May and July indicate accelerated rates of decomposition. Net weight loss was 12.6%, 12.5%, 9.3% and 11.1% for L layer blocks in the control, 100, 300 and 300W plots, respectively.

Decomposition on the control plot was significantly greater (14.4%) the following year than in sludge treated plots (100 plots, 8.4%; 300, 6.7%; and 300W, 7.5%).

Decomposition in A Horizons

Patterns of decomposition from September 1975 through July 1976 in A horizons were very similar for the control and 100 sludge plots (17.9 and 16.1%). Both had more rapid decomposition than the 300 and 300W plots (11.6 and 12.6%). From September through May, heavy sludge applications (300 and 300W plots) impeded decomposition in soil A horizons, while the trend of initial accelerated decomposition (Sept. through Feb.) was again reversed in the February through May period with a weight gain (+0.5%) on the 300 plot.

All plots showed significant decomposition between May and July; control plot, 15.2%; 100, 13.1%; 300, 12.1%; 300W, 11.9%.

Decomposition continued at a reduced rate for the next year in control and 100 plots (control, 7.3; 100, 6.8%). Accelerated decomposition occurred on the 300 series plots with a weight loss of 11.4% in the A horizon of the 300 plot and estimated at 9.4% for the 300W plot.

Decomposition in B Horizons

Again, samples were lost for the final sampling period of the soil B horizons; however, trends in the data suggest the results should be very similar to decomposition in soil A horizons. Decomposition trends of control and 100 plots are very similar, as are plots 300 and 300W, and significantly different from A horizons.

EVERETT SOILS, VIRUS AND BACTERIA

Sampling either the soil or soil solutions for virus and bacteria is complexed by the interaction of these organisms with the soil or soil solution sampling system. Swartz , in a study of retention of coliform, strain, E.coli by lysimeter plates concluded that bacteriological data derived from sampling soil solutions with lysimeter plates is of questionable value. He identified very marked retention of bacteria in the plates. For example, a control solution containing 1.2 x 10^6 organisms were passed through 6 different lysimeter plates. Colonies in the extracted solution varied from 0 to 1.2×10^3 , a reduction of 1000-fold greater in the numbers of colonies.

With these limitations in mind, the following information was developed in this study. Soil samples in the upper 5 cm from control plots developed 430 colonies of total coliform per gram of wet weight. Composite soil solutions typically average less than 20 organisms of total or fecal coliform per 100 mls. No significant differences were found between control and sludge plots.

Litter surfaces have 4.3 x 10^4 colonies per gram of wet weight immediately after spray irrigation. Coliform are evidently endemic as resampling of early sludge application on plots which studied methods of sludge application retained 3.9 x 10^4 colonies per gram of wet weight.

Soils and soil solutions analyzed for other bacteria and virus by the Virology Lab of Children's Orthopedic Hospital have not identified virus in any samples. Unidentified bacteria have been found throughout the soil profile. Results to date suggest that experimental designs which utilize aluminum oxide lysimeters may produce questionable results for study of bacteria and virus in forest soils.

¹R. Swartz, Bacteriological Evaluation of Lysimeter Plates, Metro memo (April 18, 1975) to G. Farris and R. Domenowske.

EVERETT SOILS, HEAVY METALS

The fate of heavy metals in the Everett soils is somewhat complex. It may be assumed that acid Northwest forest soils have a very strong affinity for heavy metals and provide excellent renovation. This conclusion is supported by certain analytical data; however, the exact value of these field data is unknown.

Earlier laboratory experiments with tension lysimeter plates indicate aluminum oxide, like soil, has an affinity for many heavy metals and will remove certain heavy metals from the soil solution as it passes the plate. In the laboratory, known concentrations of heavy metals were passed through lysimeter plates, and the recovery in the extracted solution determined. The following briefly summarizes the results (all values in mg/l).

	Cu	Cr	Cd	Pb	Ni	Zn
Original Solution	0.18	0.18	0.036	0.19	0.20	0.048
Average of Extracted Solution	<0.01	<0.01	<0.004	0.06	<0.02	0.039

There was variation in retention by individual plates, suggested to be a function of the aluminum oxide matrix making up the plate. The strong interaction of solution concentration and affinity of heavy metals for aluminum oxide would make interpretation of field soil solution data for heavy metals impossible. Trace concentrations of the above heavy metals were found in soil solutions extracted in the field. Interpretation of these data is impossible at this time.

SLUDGE APPLICATIONS AND FOREST GROWTH

Even the relatively uniform second growth Douglas-fir stand selected for sludge application treatments had high variability between average diameter, numbers of trees per plot, and average growth rates by plots (Table 28).

Applications of sludge did not significantly alter tree growth as studied by growth on a plot basis. Analysis of variance showed no treatment effects when tested as a difference between treatment means. The analysis of variance tested plot means for effects of sludge treatments with two replications using DBH or periodic annual increment (pai) after irrigation minus DBH or pai before irrigation as the test parameter.

Source of Variation	df	SS	MS	F	
Between treatments	3	0.48375	0.16125	0.3368	NS (0. 70)
Error (within treatments	4	1.91500	0.47875		$(\alpha = 0.10)$
TOTAL	7	2.39875			

where

standard error of difference = 0.49 mm coefficient of variation = 73.8%

The large error (within treatment) indicates the variability within treatments is greater than the variability between treatments (Table 28). Numbers of trees per hectare varied from 963 (390/ac) on the control plot to over 2000 (810/ac) on a 200 series sludge plot. The usual inverse relationship between numbers of trees and DBH is apparent. The control plot averaged 21.5 cm (8.5 in.) in diameter while the 200 series sludge plot averaged 14 cm (5.5 in.). Basal area and volume are frequently less variable than the numbers of trees or their average diameters. Basal areas ranged from 35.0 to 43.8 m²/ha (151 to 191 ft²/ac.). Volumes ranged from 305.3 to 455.9 m³/ha (4363 to 6515 ft³/ac.).

The post treatment results (Table 28) exhibit similar patterns of large variation between plots; however, mortality on plots with numerous trees (200 plot) reduced the ratio considerably.

Gross pai also varied considerably for all plots before and after treatments (Table 29). Average annual increments for all treatments generally

Table 28. AVERAGE STAND CONDITIONS BEFORE AND AFTER SEWAGE SLUDGE IRRIGATION BY PLOT AND TREATMENT

1	Volume (m³/ha)	371.7	433.8	382.9	420.6	467.4	442.0	488.9	432.7	456.3	426.8	549.7	454.9
Post Treatment	Basal Area (m²/ha)	37.5	39.3	35.6	41.9	44.2	42.6	46.6	41.0	43.2	40.1	49.1	42.2
Post Th	DBH (mm)	176	207	177	156	170	163	187	182	182	191	216	229
	Trees/ ha	1358	1037	1259	1852	1704	1679	1506	1333	1457	1185	1185	938
	Volume (m³/ha)	305.3	360.3	347.8	353.5	387.9	353.5	401.1	353,5	380.6	353.0	455.9	386.3
Pretreatment	Basal Area (m²/ha)	36.2	35.0	34,6	38.5	40.0	36.5	40.9	36.5	38.8	36.3	43.8	37.9
Pretr	DBH (mm)	165	187	167	140	155	142	174	171	168	178	198	215
	Trees/ ha_	1383	1111	1407	2123	1852	1852	1531	1358	1556	1235	1259	896
	Plots	100	100	100W	200	200	200W	300	300	300M	Water	Water	Control

Table 29. AVERAGE GROSS PERIODIC ANNUAL INCREMENT FOR PRETREATMENT, 2 AND 3 YEARS AFTER IRRIGATION BY PLOT AND TREATMENT

ല	<u>75-77</u> r)	23.1	25.8	24.6	26.0	29.8	30.3	30.4	27.1	29.8	25.3	33.1	24.2
Volume	$\frac{75-76}{(m^3/yr)}$	27.2	28.8	28.0	32.5	33.8	32.6	32.8	32.8	32.4	30.2	38.2	26.2
,	75-77	1.57	1.60	1.53	1.67	1.83	2.20	2.03	1.57	1.93	1,37	2.00	1.57
Basal Area	$\frac{75-76}{(m^2/yr)}$	1.70	1.60	1.65	1.90	1.85	2.20	2.00	1.75	1.85	1.45	2.10	1.55
	70-75	1.06	0.88	96.0	86.0	1.36	1.04	1.02	1.10	1.02	0.96	1.06	0.94
	75-77	3.4	3.9	3.2	2.5	3.0	3.6	3.9	3.2	3.6	2.8	4.0	3.9
Diameter	75-76 (mm/yr)	3.8	4.0	3.4	3.0	3.0	3.6	3.8	3.6		3.1	4.1	3.8
	70-75	2.6	2.4	2.4	1.8	2.6	8.	2.0	2.8	2.2	2.4	5.6	2.8
	Plots	100	100	J 00W	200	200	Z00W	300	300	300M	Water	Water	Control

increased between 1975 and 1977 as compared to the pretreatment period, 1970 to 1975, the control plot. Increased growth rates are probably due to the aerial fertilizer application in March 1974 and/or summer rainfall in 1975.

SLUDGE APPLICATION AND FOREST GROWTH, PAIRED TREES

The average pai for all plots is summarized in Table 30. Analysis of variance showed significant difference in growth rates when tested using paired trees. For most parameters tested, growth rates nearly doubled when comparing 1970-75 rates with 1975-77 rates.

Table 31 uses regression analyses to compare paired tree growth for sludge treated, water only, and control plots for diameter, basal area, and volume growth. The 200W treatment significantly increased diameter, basal area, and volume growth for both the 2 and 3-yr. growing periods.

The 300W treatment increased diameter growth in the 1975-76 year and increased basal area and volume growth for all years. Volume growth for all analyses was significantly increased for all sludge treatments except the 300 for 1975-77.

Irrigation with water only did not increase growth over the control (Table 31).

DIAMETER GROWTH RATE, PAIRED TREES

An analysis of variance for paired trees within treated plots and control used DBH pai after treatment minus DBH pai before treatment as of change in growth rate. Sludge applications significantly increased diameter growth rate at the 95% confidence level (Table 30).

Source of Variation	df	SS	MS	F
Between treatments	4	33.359	8.3398	$2.62*(\alpha = 0.05)$
Error (within treatment)	168	534.948	3.1842	
TOTAL	172	568.307		

Simple linear regression equations were developed from these paired individual trees (Table 2, Appendix C). Equations for pretreatment growth over initial size were compared with the control and were found to be homogenous. Therefore, the use of pretreatment pai as an independent variable in the other equations tested is not confounded by pretreatment growth differences, and treatment effects can be assessed (Table 31).

Comparisons of diameter growth showed 2 years of sludge applications (200W and 300W) significantly increased diameter growth rates over control. After 3 years of irrigation only, the 200W treatment had significantly greater diameter growth rates. Growth rates after 3 years of irrigation as a function of before irrigation growth are shown in Figure 22 for 200 W and control

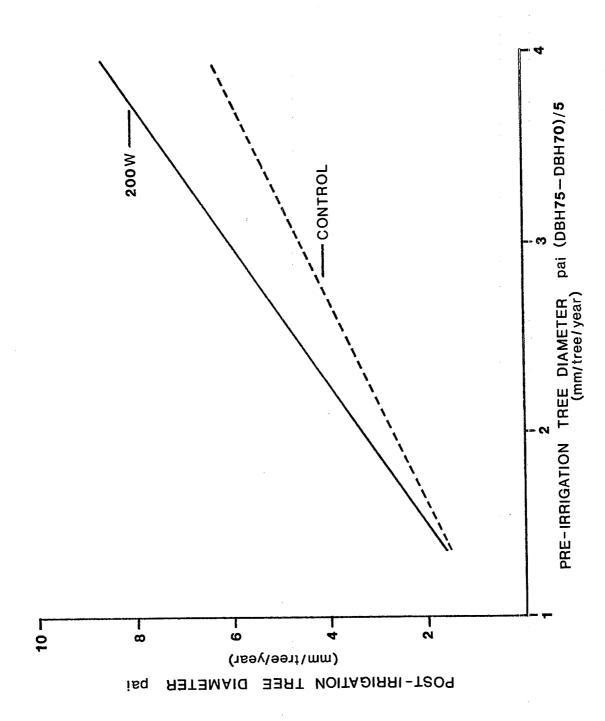


Figure 22. Individual tree diameter pai following 3 years of irrigation as a function of pretreatment diameter pai per tree (paired trees).

Table 30. A SUMMARY OF AVERAGE PERIODIC ANNUAL INCREMENT FOR PAIRED TREES, BEFORE, 2 AND 3 YEARS AFTER SLUDGE APPLICATIONS BY TREATMENT

		Diameter			Basal Area		Volume	dille
Plots	70-75	75-76	75-77	70-75	75-76	75-77	75-76	75-77
		(mm/yr)		-	$(cm^2/yr/tree)$		(m³/yr/tr	$(m^3/yr/tree \times 10)$
100	2.8	5.2	4.9	.10	.19	81.	.314	.282
100M	2.4	4.8	4.4	80.	.18	.16	.298	.255
200	2.7	5.4	4.7	60.	.18	.16	. 299	.248
Z00W	2.4	4.5	4.4	.07	.16	. 15	.225	.215
300	2.7	.7 5.1	4.9	.10	.20	.19	.340	305
300M	2.6	4.7	4.8	60.	.17	.18	. 303	.280
Water	2.8	4.4	4.1	.10	.17	91.	.341	.279
Control	2.7	4.0	4.1	.10	91.	.17	.285	.263

Table 31. COMPARISONS OF SLUDGE TREATMENTS, WATER ONLY AND THE CONTROL USING INDIVIDUAL PAIRED TREES

	Water Only	NS	NS	NS	NS	NS	NS
	300 300W Water Only	*	NS	*	*	*	*
S	300	NS	NS '	*	NS	*	NS
Treatments	Z00W	* :	*	*	* .	*	*
I	200	NS	NS	NS	NS	*	*
	100M	NS	NS	*	NS	*	*
	001	NS	NS	NS	NS	*	*
	Equation	(DBH76 - DBH75)/2 = a + b [(DBH75 - DBH70)/5]	(DBH77 - DBH75)/3 = a + b [(DBH75 - DBH70)/5]	(BA76 - BA75)/2 = a + b [(BA75 - BA70)/5]	(BA77 - BA75)/3 = a + b [(BA75 - BA70)/5]	(Vol. 76 - Vol. 75)/2 = a + b (Vol. 75)	(Vol. 77 - Vol. 75)/3 = a + b (Vol. 75)

NS = nonsignificant difference in slopes

* = significant difference in slopes α = 0.05.

trees. Slopes show a significant treatment effect for only the 200W sludge treatment.

BASAL AREA GROWTH RATE, PAIRED TREES

Basal area growth for individual paired trees showed trends similar to DBH growth. All sludge plus water treatments (100W, 200W and 300W) had significantly greater basal area growth rates than control trees after 2 years. After 3 years, only the 200W and 300W treatments maintained significantly greater growth rates. Growth during the 3 years of treatment, as a function of pretreatment basal area pai, is shown in Figure 23. Slopes of basal area growth rate were significantly different with the 200W, 300W treatments having the greatest growth rates. Again, the sludge plus water treatments all had increased growth rates over sludge only treatments (Table 31).

VOLUME ANALYSIS

Volume growth of all sludge treated trees was significantly greater than the control trees after 2 years. However, the 300 treated trees did not continue to show greater growth rates at the end of the third year (Table 31).

WATER ONLY IRRIGATION, GROWTH, PAIRED TREES

Water only irrigation did not significantly alter growth rates, whereas water plus sludge did (Table 31). Actual growth in DBH, basal area, and volume are almost identical for the paired tree analyses.

SLUDGE VS. SLUDGE PLUS WATER, GROWTH, PAIRED TREES

Growth comparisons of sludge treated trees versus sludge plus water treated trees (Table 32) indicate sludge treatments followed by water significantly increased tree growth rates. Growth rates of trees receiving water plus sludge treatments seem to increase with increased sludge applications.

Adjusted diameter pai (mm per year) as a function of weekly irrigation rates (cm) over 3 years (Figure 24) shows sludge increased diameter growth. When sludge is followed by water, diameter pai growth increased up to 25% over control diameter pai at a weekly irrigation level of 0.5 cm (0.2 in.) and then growth decreased slightly with increased irrigation.

Results from this limited study indicate that irrigation with stabilized municipal-industrial sewage sludge has increased the growth rate of individual paired trees. The 200W had the greatest effect on growth. Generally, the addition of water applied after sludge irrigation significantly increased growth over sludge-only irrigation. Therefore, to obtain the maximum growth effect of sludge application, a water irrigation treatment would be recommended to wash the sludge into the rooting zone.

Disposal of sewage sludge on this forested site has apparently had a positive effect on increasing tree growth. However, longer term studies are needed to assess impacts on soils, chemical breakdown of grease and hair

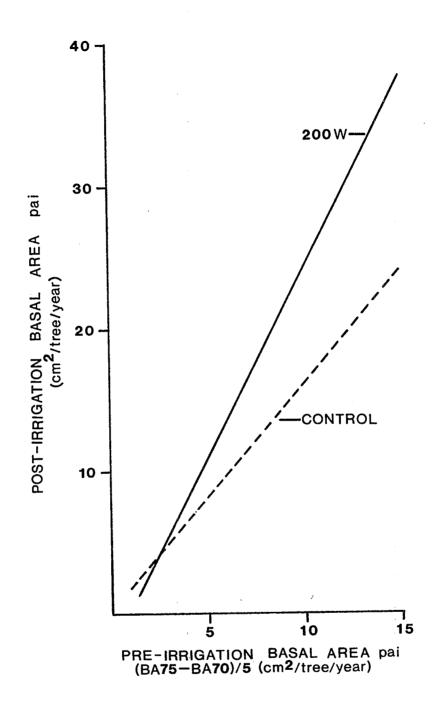


Figure 23. Individual tree basal area pai following 3 years of irrigation as a function of pretreatment basal area pai per tree (paired trees).

Table 32, COMPARISON OF TREE GROWTH SLOPES BETWEEN SLUDGE AND SLUDGE PLUS WATER TREATMENTS

300 vs.	*	÷	*	*	* *
Treatments 200 vs. 200W	*	*	*	*	SN *
700 vs. 100M	NS	NS	*	N	NS N
Equation	2-year diameter pai	3-year diameter pai	2-year BA pai	3-year BA pai	2-year Vol. pai 3-year Vol. pai

NS = nonsignificant difference in slopes

 * = significant difference in slopes α = 0.05

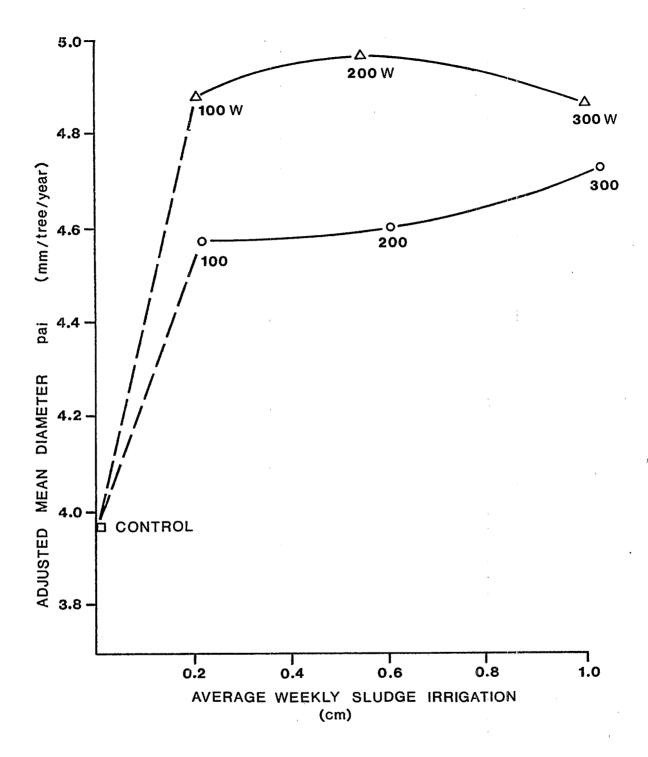


Figure 24. Adjusted mean diameter pai following 3 years of irrigation as a function of weekly irrigation rates.

applied, and impacts on microbial populations, before this technology becomes a standard silvicultural tool for intensive forest management.

DISCUSSION OF SLUDGE APPLICATIONS TO FOREST SOILS

A discussion evaluating the results of three years of testing of sludge applications in the forest must be concerned with: First, is it economically feasible as a potentially useful method; and second, what are the environmental consequences, particularly potential eutrophication of ground water? The following sections contain both a discussion and conclusions relative to the impacts of sludge applications. Few equivalent studies are available for comparison with these results.

SLUDGE APPLICATION METHODS

Disposal of sludge in Pacific Northwest forests could have either of two major objectives. A community or outlying urban area might have an objective of the disposal of maximum amounts of sludge on a minimum forest area. Investments in forest land and irrigation system would be minimum. In this case, the community would like to apply the maximum quantity of sludge per acre per year without degradation of environmental quality or eutrophication of ground water.

On the other hand, the forest industries might be interested in intensive management of a substantial forest acreage with the prime objective of accelerated forest growth. In this case, the enhancement of growth from both applied water and sludge would be spread over a many-fold greater area and use a greatly diluted sludge and a substantial quantity of effluent treatment water.

In either case, an effort would be made to keep capital investments to a minimum in the sludge application systems. This research concludes that spray irrigation is one of the most effective and possibly efficient means of distribution of sludge in the forest environment. Spray irrigation of sludge may be achieved from mobile tankers, such as truck-mounted units, under-the-canopy spray irrigation as used in this study, or over-the-canopy with the large high-powered water canons.

Costs of development of irrigation systems for sludge application to forests may be estimated to a degree by the costs of similar developments for agricultural land. Troemper (1974) developed a spray irrigation system for application of 12.1 cm (4.75 in.) of anaerobically digested sludge at a cost of \$17,900/ha(\$7245/ac.). The system installed included pipe, sprinkler, a pumping system, a drain system to intercept soil solution percolate for

recycling. Average costs for sludge disposal by the system were \$56.02 per thousand gallons of sludge (about \$260/dry ton at 5% solids in sludge).

Costs of installing a forest irrigation system would vary depending on soil types, slopes and the forest stand. Estimated costs for a system as used at Pack Forest, including pipe, sprinklers, drain valves and certain soil water monitoring equipment, are \$5000/ha (\$2000/ac.) for application of 13.3 cm (5.25 in.) of 3% solids sludge or an average cost of \$14 per thousand gallons. These costs do not include the pumping system or delivery of sludge to the disposal site. Obviously, the costs of pumps and/or transportation of sludge to the disposal site could be highly variable but probably could be capitalized over a long-term application. A minimum cost of \$110/dry ton is projected based on above cost estimates, not including a pump or hauling.

Installation of a sludge application system before establishment of the young forest would probably be the most economical. Heavy equipment could be used for burying the pipe without damage to the forest stand. As the young forest grows, a supply of water may be necessary to wash sludge from the foliage after each sludge application.

Piping at Pack Forest rested on the soil surface. Burying would be an improved alternative for a permanent system but would also add to the cost. It would have the advantages, however, in preventing freezing as well as gradual deterioration of the surface pipe and restrictions imposed on moving equipment through the forest because the pipe is on the soil surface.

In a system where uniform application of sludge is not necessary, more efficient distribution could be achieved with the use of large water canons. Sludge would be applied followed by water to wash sludge from tree crowns.

RATES OF SLUDGE APPLICATION

First year sludge applications in this study were obviously too large. Impeded infiltration occurred after sludge applications exceeded 25 cm (10 in.). Excessive amounts of nitrogen applied resulted in a very large total loading with rapid conversion to nitrate forms and leaching in the soil solution. The reduced rates which applied from 10 to 40 mt/ha (4.5 to 18 t/ac.) appear to be a reasonable range of sludge loadings. For the three soils, Everett, Mashel and Wilkeson, loadings varied from 4.8 to 29.8 mt/ha/yr (2.1 to 13.3 t/ac/yr). Physical renovation by each soil for the loadings through the given range was excellent. The total organic carbon analyses indicated no contamination within the soil profile of materials in sludge.

The original Everett soils plots had sludge applications ranging from 45 to 108 mt/ha (20 to 48 t/ac) during the study. The physical renovation capacity for selectively filtering solids was still excellent on completion of the study. Careful analyses of sludge composition are required to establish average solid content of sludge in designing a system and rates of application. This study anticipated and designed for an average solids content of 2.1%, assuming random variation during the life of the study. Solids actually averaged 3.2% over the life of the study; thus, design applications for a depth of sludge had significantly greater loadings of solids than

anticipated. In most cases, rates of sludge application to a given area will not be limited by quantities of solids but rather by the chemical constituents.

RENOVATION OF NUTRIENTS

Intensive management of second growth forest stands often requires application of numerous chemicals for control of insects, competing brush species, and in recent years, fertilizers to accelerate forest growth. Foresters are concerned with the fate of applied chemicals and potential environmental impacts, particularly in the hydrologic system from the soil solution to ground water or streamflow. Detailed forest ecosystem studies have identified nutrient leaching (Cole, et al. 1976) and impacts of forest fertilization (Gessel and Cole, 1965).

Sludge applications of this study provided nutrients in quantities greatly in excess of most other cycling studies of either fertilization or waste water management.

RENOVATION OF BASE NUTRIENTS

The renovation capacity of the three soils studied has been expressed as the percentage of retention of the total applied nutrients. In general, the percent renovation capacity for base nutrients is excellent, expressed as a percentage. Renovation by the soil retains large quantities of nutrients in the soil profile, but when very large quantities are applied, the flux or loss from the soil profile could still be significant. For example, the original 200 series plots on the Everett soil received 80.7 mt/ha (36 t/ac) of base nutrients. Renovation was 96% of that applied; however, 3.2 mt/ha (1.4 t/ac) did leach below the 2 mm (7 ft.) depth in the soil profile.

Analyses of the dissolved chemicals in ground water in wells adjacent to the plots did not reveal a trend of increasing dissolved calcium or other base nutrients in ground water. Maximum concentrations of calcium found in the soil solution of sludge treated plots ranged to 14 times that of the soil solution of control plots.

Maximum base nutrient concentrations in the soil solution were observed during driest portions of the summer. Fall rains increased the supply of soil water, thus, diluting the nutrient concentrations. It is expected that flux of the soil solution to ground water transporting the dissolved nutrients would then be greatly diluted, thus, resulting in insignificant changes in the composition of dissolved ions in the ground water. Calcium was applied in largest quantities and had the largest flux through the soil profile of any of the base nutrients. Concentrations of magnesium, potassium and sodium were greatly reduced in the ground water, and frequently were only a few-fold greater than concentrations in the soil solution of any control or water plots. In fact, the highest concentration of potassium found in any plot was in the soil solution of the control plot.

RENOVATION OF PHOSPHORUS

Renovation on all soils and plots of total P and PO₄-P was consistently excellent. The maximum loss was on the Mashel soil (30 series application) where a flux of 5 kg/ha/yr (4.5 lbs./ac./yr.) was identified when 432 kg/ha (385 lbs./ac.) of total P had been applied. On the Everett series, after three years of sludge application with quantities of total P applied in excess of 2 mt/ha (1 t/ac.), losses were a maximum of 2.0 kg/ha (1.8 lbs./ac.).

Large quantities of total P and PO_4 -P can be applied to forest soils without impacts on ground water or the soil solution.

RENOVATION OF NITROGEN

Nitrogen may well be the most critical element in land disposal of sludge, either forests or agricultural. Concentrations of nitrogen are frequently high in sludge, ranging from 0.9 to 1.8% in this study. The soil solution and organisms in the forest soils provide environmental conditions for a very rapid conversion of total N or NH₄-N forms to NO₃-N. Frequently, the renovation of NH₄-N appears excellent (ranging from 80 to 99+%). However, soil chemical reactions convert NH₄-N to NO₃-N with subsequent leaching in the soil solution, making contributions of NO₃ to ground water one of the most important implications in land disposal of sludge.

The underestimation of the average total N composition of Metro sludge resulted in large total nitrogen applications. The original Everett plots received 5.3 mt/ha (2.4 t/ac.) of total N in the 300 series with renovation of 70% (1.6 mt/ha; 0.7 t/ac. lost). Renovation by other plots was considerably less. The 200 series plots received 4.9 mt/ha (2.2 t/ac.) over the three years and had a flux of 4.7 mt/ha (2.6 t/ac.), achieving 3% renovation (Table 6).

The flux of nitrogen from the soil profile is dominated by water soluble NO_3-N . In the previous two examples, the total N flux of 1.6 mt/ha (0.7 t/ac.) on the 300 series plots was accounted for by 1.4 mt/ha (0.6 t/ac.) of NO_3-N . The 200 series plots had a very similar percentage NO_3-N flux. Of the 4.7 mt/ha (2.1 t/ac.) flux, loss of 4.0 mt/ha (1.8 t/ac.) was NO_3-N (Table 6).

Renovation of NH₄-N is excellent in these forest soils. A maximum loss of 290 kg/ha (259 lbs./ac.) was measured on the 100 series Everett plots which had received 1.4 mt/ha (0.6 t/ac.) of NH₄-N. The range for the rest of the original plots was from 1 to 31 kg/ha (0.9 to 28 lbs./ac.) over the three years of study of the original series of plots. The new series of plots established on the Everett, Mashel and Wilkeson soils had maximum NH₄-N flux of 59 kg/ha (53 lbs./ac.) on the 30 series plots on the Mashel soils (Table 19). A range of near zero to 43 kg/ha (38 lbs./ac.) occurred with the other soils and plots. The Everett soils had the best overall renovation with 6 kg/ha (5 lbs./ac.) occurring on the 40 series plots where the applied quantity was 828 kg/ha (739 lbs./ac.)(Table 12).

A volatilization loss of NH₄-N, identified by odor, was often noticed as spray irrigation of sludge commenced. Forms other than NH₄-N and NO₃-N of nitrogen in sludge were not identified. Nitrification processes with exposure of the anaerobically digested sludge to oxygen suggest a dynamic mechanism of conversion in form. Alexander (1965) studied nitrification reactions identifying exothermic and chemoautotrophic along with the groups of micro-organisms involved in biosynthetic reactions. The oxidation of NH₄-N is:

then
$$2NH_4 + 3O_2 = 2NO_2^- + 4H^+ + 2H_2O$$
$$2NO_2^- + O_2 = 2NO_3^-$$

Micro-organisms required for the above process are Nitrosomnoas and Nitrobacter. Processes of oxidation of NH₄-N are very rapid as conversion of total N and NH₄-N proceeded rapidly enough that significant increases in nitrogen could not be identified in the soil profile in the second year (Table 8). Third year analyses of Everett soils, however, did indicate increased concentrations of total nitrogen in the forest litter layer and the soil A horizon (Table 10). All sludge treated plots were averaged and tested against the pretreatment analyses to identify this significant increase in total N.

Reduced rates of sludge application (10-40 plot series) had best total nitrogen retention on the Wilkeson soils (average 79%) with the Mashel and Everett averaging 69 and 55%, respectively. Trends to increasing quantities of soil nitrogen are indicated in all three soils; however, differences are not quite statistically significant (Tables 13, 20 and 23). Retention of nitrogen applied to the soil ranged from 111 kg/ha (99 lbs/ac) on the Mashel soil (10 series plots) to 1026 kg/ha (914 lbs/ac) on Everett soils (40 series plots). In the forest ecosystem, conservation of nitrogen is related to accumulation of carbon. Micro-organisms decomposing the carbon supply require nitrogen for vital life processes. A large supply of available soil carbon frequently indicates a limited supply of available nitrogen to micro-organisms.

Applications of sludge provide a readily available source of nitrogen accelerating plant growth. As additional amounts of organic debris become available for decomposition from stimulated growth, the soil nitrogen will accumulate increasing amounts of total N.

The carbon-nitrogen ratio is an interaction of both the quantity of organic carbon, and the amount of total nitrogen shows wide variation on sludge plots at this time. The fine textured Mashel and Wilkeson soils had marked increases in the carbon-nitrogen ratio of the forest litter layer (from 31:1 to 70:1 for the Mashel and 38:1 to 68:1 for the Wilkeson). Increased organic matter occurred in L layers on these two soils without significant change in total N percentage. These data suggest that both soils will accumulate nitrogen until the carbon-nitrogen ratio is reduced.

The very coarse textured Everett soil had slight reductions in the organic matter content of the forest litter layer and A horizon. Nitrogen

contents did not change significantly, thus, a slight reduction in the carbonnitrogen ratios following sludge applications. Applied sludge and water should accelerate forest ecosystem productivity, thus, increasing the supply of carbon. With time, increased growth will increase carbon supplies and provide the mechanism for increasing the quantities of nitrogen stored in the forest floor and soil.

Quantities of NO₃-N have increased significantly in the soil solution as compared with the control and water plots. As with base nutrients, maximum concentrations occur during times of minimum soil water, thus, greatly concentrating both NO₃-N and total N concentrations. Rainfall or irrigation increases the supply of soil water, diluting the concentration of dissolved materials. The dissolved chemistry of ground water (well water) does not reveal any significant trends or a departure from normal concentration over time.

EVALUATION OF TOTAL ORGANIC CARBON

The trend to increasing quantities of the soil organic matter is verified by the analyses of total organic carbon (TOC). Increases in TOC in the C horizon with sludge applications were identified; however, analyses detected only natural decomposition organic compounds. These compounds were also identified in the ground water well samples. Large applications of sludge (45 to 108 mt/ha; 20 to 48 t/ac.) have not increased organic matter contents of the forest litter layers (Table 10). This suggests a marked increase in biological decomposition of fine or colloidal organic materials. TOC analyses also identify increased metabolism when sludge is applied. Marked increases in amounts of aliphatic fatty acids found in the soils are related to acids of normal metabolic products of micro-organisms. Increases in TOC appear to be the result of increased metabolic activities and do not represent, for an example, the appearance of refractory compounds in the ground water resulting from sludge applications.

EVALUATION OF BIOLOGICAL DECOMPOSITION

Rates of biological decomposition depend on inherent organic matter structure and chemical composition, decay organisms, and environmental conditions, interacting with action of extra cellular enzymes and physical weathering. Processes which degrade wood include immobilization, incorporation, leaching, physical damage and enzymatic dissolving. Tests used in this study include relative weight changes and the solubility of woody material in 1% NaOH. Relative changes in weight loss do not accurately measure the decomposition process, as identified in this study. Frequently, there is a slight gain in block weight attributed to micro-organisms invading the wood sample. This occurred on the 300 series plots (Table 27) and is suggested by the decrease in solubility of the wood blocks over time when extracted with 1% NaOH. With initiation of the growing season in May 1976, rates of decomposition as measured by both weight loss and soda solubility increased rapidly. There are apparently no significant differences between treatments for the duration of the study (July 1977). However, some differences in response by horizons were noted.

SLUDGE APPLICATIONS AND FOREST GROWTH

The study of forest growth by plots and treatments yielded negative responses to sludge applications. Statistically significant increases in growth could not be identified because of the very high variability in numbers of trees, sizes and basal area from plot to plot.

Re-analysis of the data on a paired tree basis using trees of approximately equal growth rate and initial diameter developed significant growth responses from sludge applications. The addition of water after sludge irrigation provided the most significant growth increases. The applications of sludge alone significantly increased growth over the control or water irrigated treatments.

Though the short-term results of sludge treatments to forest trees show promising growth acceleration, several important questions remain to be answered. Chemical breakdown of suspended materials in sludge should occur in a short time. If greases, hair, etc. are not attacked by microbial populations, then accumulations could be detrimental. To date, heavy metals seem to be no problem. In a highly organic system, they should be effectively chelated or fixed in the soil. Applied water and nutrients should stimulate biomass productivity, increasing the production of carbon which would allow greater retention of nitrogen. Increased quantities of nitrogen would continue to provide accelerated growth, thus, increasing ecosystem productivity over the long term.

Section 30

SUMMARY OF RESULTS OF SLUDGE APPLICATIONS

The initial very large amounts of sludge application to the 200 and 300 series Everett soils resulted in ponding of sludge on 5 to 25% of the surface area of certain plots. Even these very large applications did not result in noxious odors nor a generally unsightly condition in the forest. Reduced rates of sludge application (10 to 40 mt/ha/yr or 4.5 to 18 t/ac.) are almost unobservable. Excellent renovation of suspended and dissolved materials in sludge has been achieved with the exception of nitrate nitrogen. All rates of sludge application have resulted in significant nitrification and leaching of NO3-N in the soil solution. Ground water samples taken from wells at over 10 m (30 ft.) deep show no enrichment in dissolved ions from sludge applications. Significant renovation takes place in the surface 2 m (7 ft.) of the forest soil. It is quite possible that the additional 8 m (26 ft.) between the deepest sampling and the ground water table continue to renovate the soil solution.

Impacts of nutrient flux through the soil profile to ground water and/or the stream will require studies on a watershed basis.

A few significant changes in soil chemical properties were identified. While statistically significant, the ecological significance is as yet unknown. Cation exchange capacity was frequently the most significantly influenced soil property. Sludge applications generally increase cation exchange capacity of the forest litter layer and the A horizon. A general trend to increasing soil organic matter, decreasing pH, and increasing total cation exchange capacity has been identified. Large applications of calcium have generally resulted in increases in exchangeable calcium in litter and A horizons. Other changes in exchangeable cations are sporadic--statistically significant, but again ecologically inconsequential.

Modified rates of sludge application in the range of 10 to 30 mt/ha/yr (4.5 to 13 t/ac./yr.) could be received by many forest soils. Additional quantities of waste water applied following sludge applications would be beneficial. Further research should isolate impacts on a watershed basis to evaluate leaching to ground water and potential eutrophication of streams.

This short-term study did not identify significant increases in forest growth on a plot basis. Natural variability in stand densities and growth rates obscured short-term growth responses. A more sensitive analysis of paired trees of equal size resulted in significant growth increases of the paired tree receiving sludge applications. Sludge applications followed by

water seemed to be the most beneficial with the 200W treatments probably enhancing forest growth to the greatest extent. Longer term studies are necessary to adequately answer questions relative to long-term impacts of sludge application in the forest environment. These studies, if conducted on a watershed basis, would have the opportunity for a greatly expanded plot network for sludge versus no sludge treatments.

This short-term study of sludge applications in the forest provides encouraging results for the ability of forest soils to renovate constituents in sludge, along with initial indications of enhanced forest growth based on paired tree comparisons. It is also encouraging that human pathogens of the bacteria and virus natures were not isolated, and heavy metals were either absorbed in the aluminum oxide lysimeter plates or tied up in the soil profile.

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APPENDIX A

Computer printout of all soil solution analyses by treatments, sample date, cm of sludge, water and precipitation for sample period, sample pH, total nitrogen, ammonium nitrogen, nitrate nitrogen, total phosphorus, orthophosphate phosphorus, sodium, potassium, calcium and magnesium as mg/l.

S25 = 100

Wonly = water

\$25W5 = 100W

Contl = control

S50 = 200

\$50W5 = 200W

S75 = 300

S75W5 = 300W

P21 = 40 on Everett

P22 = 30

P23 = 20

P24 = 10

P31 = 40 on Mashel

P32 = 30

P33 = 20

P34 = 10

P41 = 40 on Wilkeson

P42 = 30

P43 = 20

P44 = 10

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		923EU	S25W5B	S25M5C	825	SZSMSA	SCHOOL STATE		200	SZSWSB	825W5C	822K	SZSWSA	825W5B	SZOMOC		825438	825W5C	823K	825W5A	S25W5B	SZSMSC	000				000	830	830	200	0 00	830	830	830	000		830	830	920	830	830	200	830	200	9

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Mg	24, 950	23, 750	85.850	200		900	4. 230	9.000	13.900	16. 600	19. 700	4. 120	8. 240	30.00	4.120	16. 280	10. 540	40.510	11.380	10. 180	7. 570	17. 990	11.780	9	20.000	11. 780	8.770	11. 390	22. 630	26. 310	25.030	42. 970	8. 680	B. 360	1. 450	1. 450	7.070	000	10.4.00	1.300	909	ZNZNZ	2, 350	1.300	3. 550	NNNN (2.320
පු	90.650	111.050	480.000	2000	24 000	180.000	20, 700	44. 600	64.900	143.400	102.800	25. 600	38. 6 00	139 000	25. 600	69. 160	77. 610	190.810	62.020	55. 510	45.760	40.40	42.470		132,810	83, 220	63. 730	83, 220	108, 740	203. 900	47, 210	170, 900	57, 430	64. 120	113.390	10.010	52 770	71 300	116.080	13. 300	3, 530	NNNN	11, 100	15, 500	18. 800	ZZZZZ.	11.100
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, Na	35, 9000	37. 0000	38. 7500 17500	3.4000	3. 6000	3. 2000	14, 4500	. 7700	3. 3000	15. 9000	4. 5000	. 1800 1800 1800	10.500	9.8100	2. 1800	6. 2400	11.0800	13. 4200	4. 3466	4. 2300	4. 0800 4.0800	1,00	1.00	4600	10. 5000	8. 7100	7, 3800	9.0700	10. 5700	6. 4000 9.000 9.000	3. 9400	3,9400	8.3500	9. 2900	10.8000	7 4100	6.6400	7, 6100	9, 4500	8. 9500	7. 4000	ZZZZ	4. 6000	8. 9500	18. 4500	NNNNN	4. OUUC
PO4-F	26. 4330																																														
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NO3-N	105, 630	153. 390	100	64.760	67. 600	205. 810	. 147	35. 340	58. 220	150.380	110.500	26. 750	77, 160	155, 360	17. 730	57. 400	74. 400	174, 250	7.5. 040	77, 420	142.280	47. 660	16. 520	77. 510	87. 600	73. 640	55. 460	81. 600	104. 310	113.080	77. 580	87.160	47. 570	54. 790	45 200	52, 760	44.080	65.880	115. 760	. 561	. 197	. 023	3000	. 516	05%	282.	
NH4-N	100.890	115.010	. 001	56. 710	58. 810	5.880	5. 190	345	1. 420 024 021	78/ ·	1000	2 10	272	. 001	. 328	1.240	080		1 5	36	070	968	. 421	. 001	. 151	1. 595	. 313	.001	. 02.0	9. 433	. 049	. 051	1. 783	315	3 6	3.740	1.040	. 001	. 001	. 377	345	. 042	. 151	. 377	66. CCC		•
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ర్	21. 4400	21. 4400	21. 4400	35.0800	35, 0800	35, 0800	35.0800	17. 4800	10.40	19 9600	11, 9700	11.9700	11.9700	11. 9700	19. 2800	14. 2800	17. K800	21 4700	21. 4700	21, 4700	21. 4700									B. 1300							28. 0700	2B. 0700	28. 0700	4. B000	4. B000	4. 8000 4. 8000	4. 8000 4. 8000	7. 6000	7. 6000	7. 6000	
Date	751113.		•				760106.	•	•	•		•	•			7,0003.	740403.	760729	-	•	760729.	760921.					741107.					770106.		_			770613.	770613.	770613.	750703.	750703.	750703.	780004	750806.	750306.	750806.	
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ğ	s. 900	3. 450		200		747747	2, 330	31. 600	21. 150	NANA NA	NANA Z	22, 100	26. 600	15, 150	330	9 . 700	40.800	30.00	. , 60	100	36		900	6.410	31, 200	18,800	19, 160	6.410	32, 710	31. 480	20. 990	10. 950	34. /10	12, 610	21. 380	32. 710	21. 510	14.00	30.730	26. 7.0 26. ABD) C	31.130	19 930	42.970	16.900	12, 930	28, 020	35. 960	
පු	36. 700	22. 500	Newsystem .	47 800		NAMAN	11, 100	122, 600	64. 600	ZZZZZ	NNNN	86. 950	125, 850	85, 100	11. 100	6. 700	38. 400	150.000	200	49. 700	000	14.000	3 2	43.70	146.000	79, 200	82, 680	43, 400	146.880	114. 780	134. 350	62.400	146.880	78, 150	129, 160	146.880	143.740	83. 250 80. 800 80. 800	7. 380	116.380	475 555	220.400	110 300	214 300	99 310	99, 310	166. 170	197.830	
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Na	30, 3500	27, 4500		7000	22.000	NAMAN	4000	17,0000	16.0500	NNNN	NNNN	37, 5000	65, 0000	52, 5000	4. 6000	4.0000	7. 0000	52, 0000	2. 3000	4. 5000	15.0000	52,0000	3000	4. ago	31.000	4.	3 4300	7, 3100	21.4000	14, 5800	7. 5300	6. 6900	21. 4000	7 3000	9.9800	21.4000	14. 1400	7. 6500	B. 6200	13, 1400	14. 7300	4. 4300	0000	. 6300	4. 0700	11, 9300	13.8200	19, 1000	
PO4-P	6.8750		٠		9000																												. 0460										•	•	•	•	•	. 0010	•
Tot P	となるとこと	NAMES OF	2222			MINIMINE	NIMININ	NNNN	ZZZZZ	NNNNN	NNNNN	12 6800	0400	ZZZZ	NNNN	NNNN	ZZZZZ	ヹヹヹヹヹ	ZZZZ	. 6490	. 1140	0500	0800	7. 5400	0040		000	9000	ZZZZZ	0280	NNNNN .	. 1140	NNNN	0280	1240	ZZZZZ		. 9600	0080	ZZZZ	0300	. 2650	0.00	0820		NAMANA	NNNNN	ZZZZ	: : :
NO3-N	57.000	45. 500	023	9013		110.000	200	244 740	202 240	A 23.	CEO.	124 330	244. 460	155. 790	•	ů.	24.25	SSS.	×	5.	S	23	6.	ਨੂੰ i	3			6	9	ij	187	53	187.380	125	. 0	187	15%	78	2	5	5	9	22 1		9	20, 230	2000	172 640	
NH4-N		96. 700	. 042		164, 167			101.010	201.010	EV1. EEC	400		161, 900	111,480	. 151	100, 860	39.460	95,000	. 058	13, 210	5. 980	77, 250	. 065	6. 190	4. 330 0.00	4B. 740	200	9 6	39.00	033	B. 860	2. 660	39.010	.033	9 0	39, 010	. 075	. 127	980 .	909 .	. 036	. 367	. 025	460	. 094	085	3 6	1 440	
Tot-N	2222	スマスマン	アズズス	ZZZZ	ZZZZ	NNNN	NNNN		MINIMINI		MINIMIN	070 076	426 360	NNNNN	ZZZZ	ZZZZ	ZZZZZ	NNNN	ZZZZZ	54, 990	67.130	342, 190	70. 422	37. 720	41.760	ZZZZ	104. 942	74. /40	NAME OF THE	125, 316	NNNN	59, 210	NNNN	125, 316	53. 762	NAMAN	ZZZZZ	78. 780	92.800	ZZZZ	155, 520	162, 355	159, 330	123.840	ZZZZ	89, 722	71. 388	B/O ZII	
Hd	040	7. 780	NNNNN NNNNNNNNNNNNNNNNNNNNNNNNNNNNNNNN	スペスペス	7.920	7. 800			7. /BO	7. 840			7.950	. 4	ZZZZZ	8 080	6. 660	4. 730	7. 590	7. 100	5, 720	4.750	7. 100	6. 750	5. 630	4. 710	5.880	6.940		5.00	6.370	5, 720	4. 530	6. 340	7. 130	. 300 . 4	6. 970	7.090	6. 640	5.170	6. 320	5.360	5.060	4. 600	6. 470	5. 290		4. 510	
ర్	0080	13.0800	13,0800	13.0800	5.8600	5.8600	5.8600	5.8600	7. 1700	7. 1700	7. 1700	200	14. 4600	44.400	14. 4600	3000	30 0000	30.000	30.0000	19,9600	19,9600	19,9600	19.9600	11.9700	11.9700	11. 9700	11. 9700	11. 6800	11. 6800	11.6800	8 7900	B. 7900	B. 7900	8. 7900	13. 9300	13. 4300	13, 9300	10, 2600	10.2600	10. 2600	10, 2600	11, 1200	11. 1200	11. 1200	11. 1200	20.8700	20.8700	20.8700	KO. 0/20
Date	750037	750826.	750326.	750826.	750918.	750918.				751014.			751113.									760226.		•			760413.		760603.					-	760921.	760721.	760921	761107.	761109.	, 761109.									770331.
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		Date	ర్	Hd	Tot-N	NH4-N	NO3-N	Tot P	PO_4-P	Na	M	బ్ర	Mg
875HSLTR	477	•	26. 7900	4. 920	33.598			2. 3140	4. 2200	5.9900	5.240	37, 620	2 700
B75W5A	478		٠o٠	5. 730	50,440			NAME.	. 3590	6. 4200	8. 600	56. 790	7.380
875W5B	479	770613.	٠ò٠	5. 280	130, 654			. 0340	. 0010	9. 8800	16. 440	133,020	14 750
MONE YE TR	2	750703.	4. 8000	6. 170	3. 276			. 2930	. 0550	9, 2500	9 900	12,000	-
MONTAN	7		4. 8000	6. 770	1. 002			. 1040	0320	6. 8000	3 800	4.0	
HONCAS	6		₽ . 8000	7.000	ZZZZ			- 0010	.0010	9 3000	200	6	
HONITAC	2	•	4 . 8000	6. 830				77772	0010	8.8500	006	900	
FONLY TR	5	750006.	9. 7800	7.020	1.940			. 3330	0640	6. 4000	200	900	
MONIL VA	8	750806.	9. 7800	6.760	1.832			1200	0000	0000		8 € 6 €	3 8
FOR VB	\$	750806.	9. 7800	7.000				- 0010	0100			9 6	3
MONT VC	8	750306.	9. 7800	6. 670	ZZZZZ			NANNAN		7500		3 6	3
JONLYLTR	109	750826.	ö	7. 250	ZZZZ			NAMAN	000			15.500	1.430
MONLYA	110	750876.	o	7.020	1. 477			0890		9.00		0 0	OCB.
MONL YB	7	750826.	10, 3900	7.000	NAMA					4.6300	100	000	06/
HONITAC	4	750826.	o	6.800	NWWW			OTO :		2000		900	1. 400
JONLYLTR	126	750718.	K	7, 130	MANNA			ANNAMA	PICE COLUMN	9.000	1.100	12. 700	1.870
MONLYA	127	750918.	7,6400	6. 620	14 545				0000	9.000	9.40	300	200
JONE		750918	7.6400	4 920				966	0100	7. 2500	3. 650	000 000	00 83 90 90
MON! VC	159	750918	7.6400	190	NAMAN A			0000	0010	6. 1500	1. 950	B. 700	300
WONLYLTR		751014	7 1700	2.5	202				0100	9,000	1. 400	12. 200	1.800
HOWLYA		751014	7 1700	7.040	3 c			. 675	0/80.	e. 5500	3. 700	11.500	2.030
EA INDA		751016	4400) () () () (A. 74.1			3080	. 0170	5, 7500	3.600	3. 830	. 400
AONIL VC		751014	7 1700	7. 1.				NANA.	. 0010	5. 3000	1.750	3. 800 3.	1. 200
JON LYLTR		751113	3400	. 4				0010 0100	. 0010	3. 3000	800	11.000	1. 700
MONT AN	17	751113	16.3600	9 0				2580	. 0380	2. 9000	9 9	9.300 300	800
NON! VB		731113	٠.	7 0 0				1440	. 0160	3. 2500	9. 30.	a. 500	. 650
ONL VC		751113	16.3600	000	1. C/4			0940	. 0010	2. 6000	1.300	4.050	. 750
ADNI VI TR	6	760106		2.5				0010	0100	4. 5000	. 000	9. 200	1. 500
Y.		740104	•	7. 199				3380	. 2370	3000	1. 280	ი 000 00	. 160
MONI VB	-	760106	•	27.6	9 6			0840	. 0460	8000	1. 480	1. 700	001
	-	746104	•	200	014.44			. 2790	. 2430	1.3000	. 570	2. 300	. 130
ADM. VLTR	-	760226	10.00	. K				- 0010	. 0010	5.5000	1.200	8. 200	1.300
MONEYA		760254	٠.	7.750	1. OK3			. 2060	4490	1. 1000	2 . 800	3. 700	. 890
. YB	_	740224	40.400	1 . 6	7 6			0080	0690	1.0000	. 800 100	9. 900 900	200
FONT AC		760224	٠.	7 340				. 0220	0100	1. 1000	1.100	3. 400	290
ADM: YL TR	•	760413	11 9700		1. 33			1080	0010	2. 6000	. 620	7.000	 8
ADNEXA	•	760413		7.180				. 2420	. 2040	1. 7500	1. 940	4. 650	7. 670
DNLYB	-	760413	220	7.170				0830	. 0290	1. 6800	00 00 01	2. 490	. 496
JONIL YC	•	760413		7. E.O	10 400			. 0440	0100	2.0100	1.100	3. 140	909
JONIL YL TR	•	760603	: _	. 4 000					0100	4000 1000	946	4. 410	. 736
WILYA	•	760603	14 2000	7 6	. 740			NNNK	1760	2. 0200	2.840	6. 240	800
WAL YE		740403	: _					1000	0140	1. 9300	2. 930	3.880	350
2× 1×C		740403		200	MINIMA				. 0010	2. 6300	1.070	4. 210	. 260
		160000	•	000	NIMMIN				. 0010	3. 5500	330	5. 730	. 830
	•	720767.		7. 190				3800	. 3200	1. 1600	2.360	3.600	. 320
	9 6	740770	13. 8300	7. 230	. 548			. 0840	. 0710	1. 0800	1.440	1.890	300
		100727.		7. 180	. 327			. 0340	0100	1. 2700	. 600	1. 980	300
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MONILYL IK	•	760921.	٠,	7. 190	ZZZZ			. 3800	3200	1.1600	2.360	3.600	. 520
		760721.		7. 230	. 548			. 0840	. 0710	1. 0800	1.440	1.890	300
	755	760921.	13. 9300	7. 310	ZZZZZ			. 0420	4710	4. 9300	1.760	5.450	. 670
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3 5	C B		21.00	ě	114, 314	00100	114, 140	. 0360		8	00	113, 150	13.2100
3 5	B C	70 770513	21.4800	'n	62, 788	0010	62,650	.0580		8	200	65, 260	10.0700
3	֝֝֝֟֝֝֝֝֝֜֝֟֝֝֝֡֓֓֓֓֓֓֜֝֡֜֜֜֟֜֓֓֓֓֓֓֜֜֜֜֜֜֡֓֓֡֡֜֜֜֜֡֡֡֜֜֡֡֡	• •	NAMANA		1,955	0010	. 033	. 0200		8	8	12, 360	3.5800
	֭֡֝֞֝֝֞֜֝֝֓֓֓֓֞֝֜֜֜֝֞֜֓֓֓֓֓֡֜֜֜֜֡֓֓֓֓֜֜֜֡֡֡֓֓֡֡֡֡֡֓֜֡֡֡֡֡֡֡֡	22 74110%	NAME OF THE PERSON OF THE PERS	j	1.001	0230	.00	9009		8	300	5, 980	1. 5500
,	CB	-	ZINAN	3	. 561	. 0010	. 601	1200		윦	8	8.170	2. 6100
,	9	-	7 8500	4	2, 679	0180	. 30	20B0		8	28	25, 340	4.0400
	<u>.</u>	37 770106	7, 8500	j	1.036	0310	300	0090		윦	8	11.730	2. 6300
	c á	•	7, 8500	Š	4.074	0010	3, 330	. 0320		8	000	12. 560	2.8900
100	9	-	21,0400	ø	NAME OF THE PERSON OF THE PERS	7610	スシンシン	. 5240		8	800	25. 200	3, 4700
100	<u> </u>	•	21.0400	Ġ	ZZZZZ	0880	るること	. 0640		S S	904	18. 510	9. 2000 1.
10	C ss		21.0400	i	41.084	0100	40. 730	. 0280		န္တ	400	42, 520	9.2400
100	1		18, 6200	ó	96.800	1. 1100	73, 420	. 5600		8	8	113, 150	13. 6700
40		R1 770413	18, 6200	4	43, 574	. 0830	42, 630	. 0320		ş	00	47.870	9. 6100
	C a	-•	18, 6200	Ġ	49. 472	. 0660	47, 120	. 0300		8	200	44, 750	9.7600
			25. 3400		2, 503	1990	. 723	. 6000		ş	92,00	9. 930	3. 6000
	١,	•	25.00	4	5.098	0010	4.720	. 0560		800	8	7. 120	2. 7200
7		• •	0100	i 4	NNNN	50.2700	71.810			80	900	29. 150	10.7600
5	۲. ۱	770413	0700	•	6.740	8310	5, 510	. 0700		စ္တ	000	6. B60	3.0900
		·r	22 7000	4	11.850	10 8010	3. 290	2522		700	2700	12, 750	3. 33 00
3 5	7 Y	•	23 7900		10. 474	5, 7680	4.790	. 2360		ş	200	10.990	4. 6300
N 9	c .	. 1	23. 7700	: 4	34 770	13 0700	44.250	7, 1300		8	200	61.240	16. 3900
N C	¥	٠,	K3. 4700	jr	24 730	16, 3300	12, 610	7880		800	9003	19,340	9. 6100
3 5			20.00	: 4	11, 584	5.0470	5, 080	3000		700	90	12, 390	4 . 8700
3 5] 4	• •	22 0400	ir	6. 372	1.1690	4. 830	. 0440		8	300	9, 230	9000 1000
35	٤ -		20 9700	: <	111, 330	24, 9300	75, 770	1. 3240		8	2400	76. 400	21. 1100
2 6	¥.	BB 770613	20 4700	i	13.820	8.2100	8.040	. 0980		8	500	15. 330	7. 4600
3 5		·r	20, 2200	ė	10, 328	2. 9110	7, 350	. 2040		8	200	10. 630	2. 8800
5 2	.	• • •	20, 2200	9	6.668	. 0010	5. 670	. 1660		Ş	80	9. 580	4. 1500
7	41		18, 3700	ø	33, 390	3700	31, 550	. 6980		8	200	35, 830	11. 1400
46		90 770613.	18, 3700	Ġ	29.490	. 2980	26. 730	0660		8	2500	25.380	2000
7	LTR	91 770613.	26. 9800	ø	49.110	33. 4500	15.610	5. 4940		3	2 6	9 6	30.00
17	<	92 770613.	26. 9800	۲.	28, 280	5. 4900	23. 280	. 1080		3 5	3	10. 400	1 1 2 2
17	æ	93 770613.	26. 9800	۲.	13.810	4. 0800	B. 370	0240				7.000	3 5
27	LTR	•	24, 2300	Κ.		65, 3200	B?. 040			3	3	96.	1100
670	•		24, 2300	۲.	38.390	9. 2900	28. 730	. 0880		8	009	34.400	36
5	· #	,	24, 2300	۲.	54, 220	23, 1500	30, 340	0880 .		8	0000	33.610	36.5
E			21, 9900	ø	11.900	8. 3200	3.340	1. 3340		8	3	23, 130	976
2	•		21, 9900	Ġ	16. 296	1310	15, 200	. 0740		000	200	28. C30	7. 6300
. E		•	21, 9900	4	6. 140	. 0010	5, 190	. 0340		28	9	15. 330	4. CIC
	Ē		1200	4	11.000	3.0600	5. 100	. 3380		28	9400	12. 210	, 4000 1000 1000 1000 1000 1000 1000 1000
1	: •		19, 1200	,	3. 182	. 0010	1. 680	. 1020		80	0800	13.540	3. 4700
} 4	S di	100 770412	0.00	Þ	ZZZZZ	2540	£. 630	NANA NANA		16. 6900		23. 320	5. 7000
}	3		:	i						•			

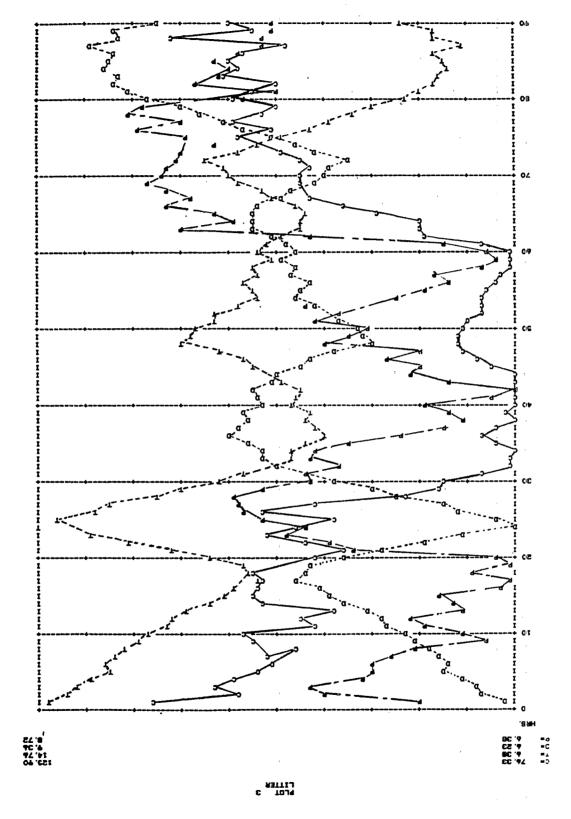
APPENDIX B

A computer program was written to automatically plot conductivity, temperature, dissolved oxygen and pH for data recorded on mag tape from the Metro-data units. Several scans were run during initial stages of the project. Large short-term variations in monitored parameters were found, particularly in the forest litter and soil A horizons. For the scans used as an example in Figures B-1 to B-4, O hours were mid-day on a warm day in mid-June. An irrigation with sludge and water took place at about 18 hours. Temperature was depressed, conductivity increased, dissolved oxygen increased and pH was significantly depressed in the litter layers.

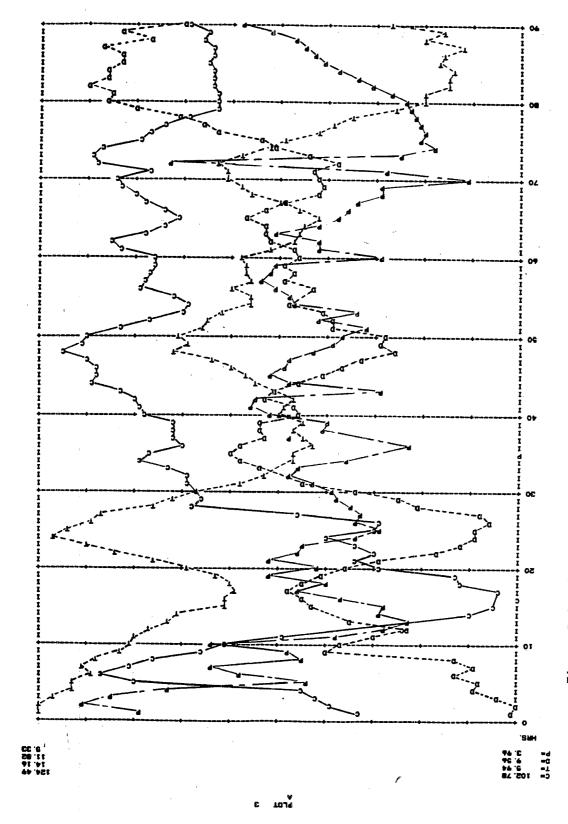
FIELD RECORDING OF SOIL SOLUTION PARAMETERS

Soil solution extracts from the lysimeter plates at various depths were passed through sensing cells where temperature, dissolved oxygen, specific conductivity and pH were recorded at 10-minute intervals. Data were recorded on Metro-data magnetic tapes which were programmed for display of average hourly values through computer printout. A complex scaling format provided for scaling each variable between its maximum and minimum range for the duration of a scan. Computer output for simultaneous runs of the litter, A, B and C horizons is shown in figures in Appendix B. Figure B-1 (forest litter) shows diurnal variation in soil solution temperature (T), scaled from 6.380 to 14.76° C. The line D shows parts per million of DO cencentration varying from minimum values of 6.23 to 9.36 mg/l. The C line is micromhos of specific conductivity with variation from 76.3 to 125.9. Soil solution pH (P) is scaled from 6.38 to 8.72. The inverse relationship between DO and temperature shows daily periods of maximum temperature, show minimum DO levels, and inversely hours of coldest temperature show maximum DO. pH and conductivity generally show similar trends, though different magnitudes of change, suggesting that bicarbonate is probably influential in pH measurement with increasing quantities of free ions releasing additional cations to solution.

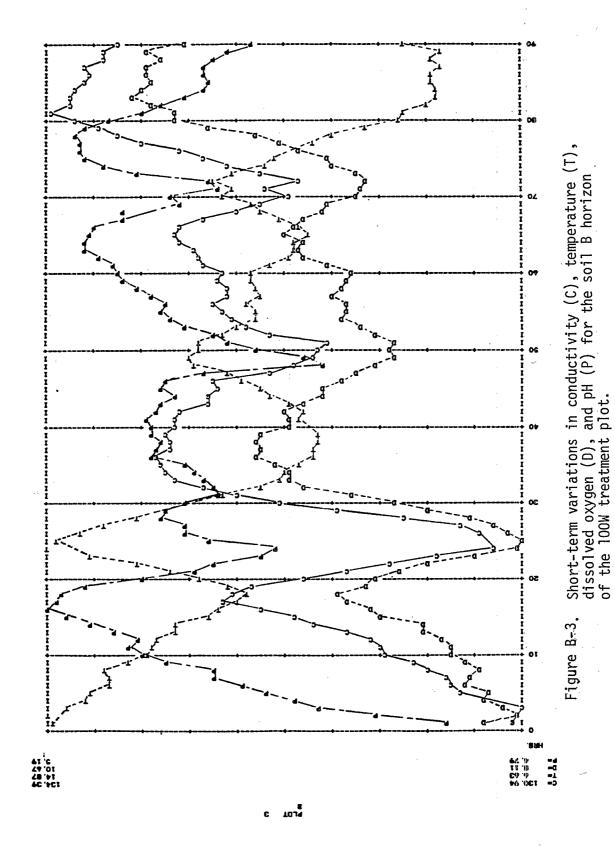
The four figures show declining variability in each parameter with increasing depth, as upper soil horizons and litter layers are the most dynamic. At lower depths, there is less variation in temperature, DO and conductivity.

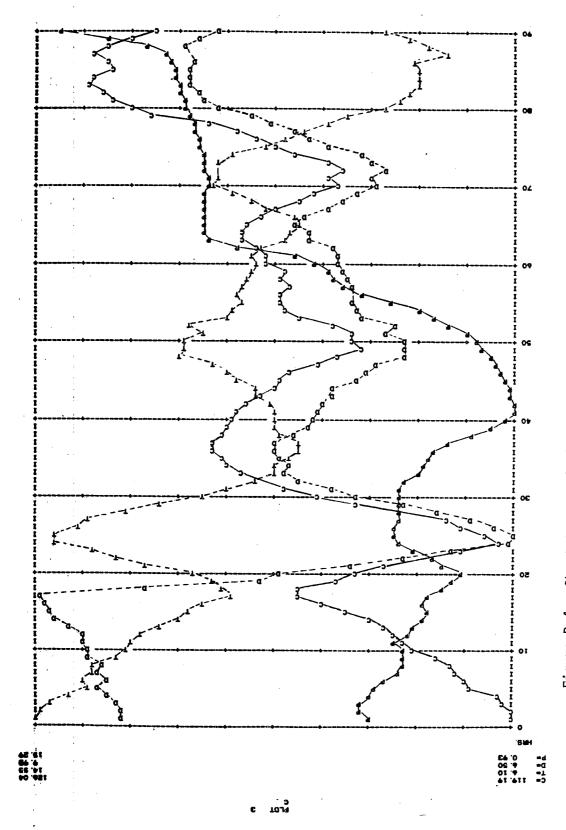


Short-term variations in conductivity (C), temperature (T), dissolved oxygen (D), and pH (P) for the forest litter layer of the 100W treatment plot. Figure B-1.



Short-term variations in conductivity (C), temperature (T), dissolved oxygen (D), and pH (P) for the soil A horizon of the 100W treatment plot. Figure B-2.





Short-term variations in conductivity (C), temperature (T), dissolved oxygen (D), and pH (P) for the soil C horizon of the 100W treatment plot. Figure B-4.

APPENDIX C

Pretreatment diameter equations based on increment cores.

Initial diameter measurements of the Pack Forest Sludge Project began after the 1974 growing season. At that time, increment cores were taken from ten trees per plot. The sample of ten was to be representative of the plot: four site trees, two trees of intermediate diameter, and four trees representing the lower diameter classes.

X-ray photos were taken of each increment core and measurements were obtained from these photos. The measurements, to the nearest 0.02 in., represented the last 2, 5, and 10 years of growth for each tree.

Marshall Murray's Inside and Outside Bark DBH Equations for Plantation Douglas-fir (shown below) were used to transform the increment core measurements to DBH outside bark (o.b.).

Shown below are the steps taken to calculate DBH (o.b.) for the years 1964, 1969 and 1972. These steps were repeated for each individual tree.

- Step 1. 1974 DBH (o.b.) * .920 = 1974 DBH (i.b.)
 - 2. 1974 DBH (i.b.) (2-yr. increment * 2) = 1972 DBH (i.b.)
 - 3. 1972 DBH (i.b.) * 1.088 = 1972 DBH (o.b.)
 - 4. 1974 DBH (i.b.) (5-yr. increment * 2) = 1969 DBH (i.b.)
 - 5. 1969 DBH (i.b.) * 1.088 = 1969 DBH (0.b.)
 - 6. 1974 DBH (i.b.) (10-yr. Increment * 2) = 1964 DBH (i.b.)
 - 7. 1964 DBH (i.b.) * 1.088 = 1964 DBH (o.b.)

Table C-1
Regression equations for all trees by irrigation treatment

Equation		100 Low level (n = 97)	100W Low + Water (n = 51)	200 Mid level (n = 144)	200W Mid + Water (n = 68)	300 High level	300W High + Water	Water	Unirrigated control
(DBH75 - DBH70)/5 - a + b(DBH70)	SE1	0.015444 0.091371 0.0011 0.69	0.022054 -1.234813 0.0013 0.86	0,023834 -1,116251 0,0009 0,83	0.024498 -1.392205 0.0003 0.99	0.015520 -0.031551 0.0008 0.75	0.021632 -1.270902 0.0006 0.95	0.0008 0.66691 0.666	0.019772 -1.299486 0.0019 0.75
(DBH76 - DBH75)/2 = a + b[(DBH75 - DBH70)/5]	~ 4 83 ° × ·	2.098700 -1.549147 0.1732 0.61	2.367967 -1.385977 0.2343 0.68	1,729283 -0,613604 0,1032 0,66	2.143671 -0.480665 0.1269 0.81	2.075038. -1.352659 0.1523 0.62	2.323145 -1.326979 0.1626 0.78	_2.303655 _2.135855 0.2166 0.55	1.850338 -1.077760 0.1615 0.78
(DBH77 - DBH75)/3 = a + b[(DBH75 - DBH70)/5]	10 M M M	1,922543 -1,331659 0,1713 0,57	2.223593 -1.375565 0.2140 0.69	1.616477 -0.677416 0.0843 0.72	1.808484 -0.846618 0.1569 0.80	2.322421 -1.274568 0.1666 0.54	2.322421 -1.274558 0.1666 0.77	2.231804 -2.199403 0.2186 0.53	1.868898 -1.040446 0.1383 0.84
(BA75 - BA70)/5 = · & + b(BA70)	_	0,032181 0,000021 0,0010 0,91	0.038438 -0.000154 0.0018 0.91	0.043387 -0.000134 0.0014 0.88	0.041636 -0.000153 0.0007 0.98	0.033469 -0.000036 0.0008 0.94	0.037304 -0.000151 0.0010 0.96	0.024466 0.000130 0.0011 0.83	0.036193 -0.000229 0.0020

Table C-1 (continued)

4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4		100 Low level	100W Low + Water (n = 51)	200 Mid level (n = 144)	200W M1d + Water (n = 68)	300 High level (n = 115)	300W High + Water (n = 59)	Water (n = 96)	Unirrigated control (n = 38)
(BA76 - BA75)/2 = a + b[(BA75 - BA70)/5]	T SE T	1.648437 0.000076 0.0972 0.75	2.398586 -0.000256 0.1865 0.77	1,772201 -0,000094 0,0772 0,79	2,270126 -0,000058 0,1122 0,86	2.064710 -0.000215 0.0766 0.87	2,213354 -0,000212 0,1324 0,83	2.061667 -0.000266 0.1384 0.70	1.800101 -0.000252 0.1479 0.80
(BA77 - BA75)/3 = a + b[(BA75 - BA70)/5]	7 # 82 F	1.504326 0.000123 0.0986 0.71	2.227958 -0.000235 0.1697 0.78	1.613280 -0.000083 0.0620 0.83	2.214242 -0.000510 0.1104 0.86	1.859052 -0.000123 0.0826 0.82	2,292921 -0,000227 0,1409 0,82	2.034952 -0.000314 0.1412 0.69	1.875561 -0.000259 0.1096 0.89
(Vol76 - Vol75)/2 = a + b(Vol75)	T S B T	0.074716 0.002766 0.0032 0.85	0.119257 -0.007834 0.0048 0.93	0.112099 -0.004221 0.0025 0.93	0.110014 -0.003667 0.0036	0.099606 -0.003252 0.0020 0.96	0.111691 -0.006706 0.0041 0.93	0.090916 -0.001882 0.0032 0.90	0.081969 -0.005481 0.0051 0.88
(Vol77 - Vol75)/3 = a + b(Vol75)	. 2 a 8 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4	0.059414 0.003991 0.0032	0.104051 -0.006525 0.0050 0.90	0.093390 -0.003329 .0.0022 0.93	0.097248 -0.002288 0.0043 0.89	0.081868 -0.001415 0.0027 0.89	0.102272 -0.005460 0.0050 0.88	0.081602 -0.002952 0.0030 0.89	0.079673 -0.006712 0.0043

1SE - Standard error of b coefficient.

Table C-2 Regression equations for paired trees by irrigation treatment.

Equation		100 Low level (n = 35)	100W Low + Water (n = 27)	200 Mid level (n = 37)	200W Mid + Water (n = 27)	300 High level	300W High + Water	Water	Unitrigated control
(DBH75 - DBB70)/5 a + b(DBH70)	1, 2, 2, 3, 4, 4, 4, 4, 4, 4, 4, 4, 4, 4, 4, 4, 4,	0.018855 -0.712716 0.0028 0.59	0.021239 -1.231359 0.0014 0.90	0.025284 -1.609345 0.0020 0.82	0.024451 -1.403748 0.0005 0.99	0.020106 -1.172949 0.0008 0.95	0.023085 -1.588449 0.0010 0.94	0.01727 -0.546116 0.0013 0.88	0.021642 -1.743038 0.0017 0.83
(DBH76 - DBH75)/2.= a + b[(DBH75 - DBH70)/5]	# 83 # 고 # 83 54	1,951750 -0,318911 0,2436 0,66	2. 252687 -0. 537911 0. 2965 0. 70	1.761001 0.636425 0.2261 0.63	2.749107 -2.038610 0.1912 0.89	2.021593 -0.377513 0.1575 0.82	2.417146 -1.615903 0.2439 0.76	2.153193 -1.538215 0.2644 0.73	1.859155 -1.048305 0.1585 0.80
(DBH77 - DBH75)/3 = a + b[(DBH75 - DBH70)/5]	4 8 8 4 A	1.829662 -0.257320 0.2194 0.68	2,114383 -0,688487 0,2838 0,69	1,707798 0,086809 0,1957 0,69	2.722826 -2.062345 0.1980 0.88	1.762675 0.072813 0.1642 0.77	2.352021 -1.354091 0.2514 0.74	1.986358 -1.404181 0.2853 0.66	1.878875 -1.007116 0.1321 0.85
(BA75 - BA70)/5 = a + b(BA70)	,D 약 및 N	0.036275 -0.000121 0.0033 0.79	0.036420 -0.000115 0.0018 0.94	0.042227 -0.000175 0.0028 0.87	0.041106 -0.000180 0.0012 0.98	0.035357 -0.000147 0.0010 0.97	0.038329 0.000169 0.0012 0.97	0.032967 -0.000055 0.0023 0.89	0.037290 -0.000282 0.0019 0.91

Table C-2 (continued)

Equation		100 Low level (n = 35)	100W Low + Water (n = 27)	200 Md level (n = 37)	200W M1d + Water (n = 27)	300 High level (n = 37)	300W High + Water (n = 33)	Water (n = 27)	Unirrigated control (n = 37)
(BA76 - BA75)/2 =	م	1,963537	2,444350	2,110571	2,832737	2.058824	2.343178	1,783088	1.792263
a + b[(BA75 - BA70)/5]	4	0.000012	-0.000232	-0.000015	-0.000419	-0.000045	-0.000417	-0.000098	-0.000235
	S	0.2090	0,2786	0.1815	0,1626	0.1158	0.1952	0.2226	0.1502
	r ₂	0.73	0.75	0.79	0.92	0.90	0.82	0.72	0.80
(BA77 - BA75)/3 =	م	1.966495	2, 238534	1.887084	2,759162	1,836350	2.368171	1.629902	1.862848
a + b[(BA75 - BA70)/5]	~	-0.000048	-0.000215	-0.000019	-0.000411	-0.000092	-0.000369	-0.000037	-0.000232
	SE	0.1796	0.2569	0.1529	0.1702	0.1338	0.2110	0.2362	0.1096
	47	0.78	0.75	0.81	0.91	0.84	0.80	99.0	0.89
.(Vo176 - Vo175)/2 =	م	0.107828	0,116060	0.122271	0.125976	0.099637	0.113689	0.098879	0.082687
a + b(Vol75)	•	-0.005095	-0.006246	-0.004896	-0.007524	-0.003679	-0.007172	-0.003111	-0.005905
	SE	0.0065	0,0060	0.0067	0.0066	0,0040	0.0064	0.0073	0.0053
	r ₂	0.89	0.94	0.90	0.94	0.95	0.91	0.88	0.87
(Vo177 - Vo175)/3 =	م	0.093941	0.099226	0.102549	0.121253	0.085581	0.103725	0.079926	0.080438
a + b(Vol75)	43	-0.003671	-0.005336	-0.004371	-0.007423	-0.001791	-0.006173	-0.002224	-0.007163
	SE	0.0059	0.0055	0.0051	0,0080	0.0049	0.0080	0.0071	0.0044
	714	0.88	0.93	0.92	0.90	06.0	98.0	98.0	06.0

1SE - standard error of b coefficient,

Key to Tables C3 &C4

	Variable	Labels	Parameter	4
	1	DB70	DBH 1970	
· ·	2	DB75	DBH 1975	e e
	3	DB76	DBH 1976	
	4	DB77	DBH 1977	
•	5	BA70	Basal Area 1970	
•	6	BA75	Basal Area 1975	
	7	BA76	Basal Area 1976	
	. 8	BA77	Basal Area 1977	
	9	V75	Volume 1975	
	10	V76	Volume 1976	,
	11	V77	Volume 1977	
	12	D5-0	(DBH75 - DBH70)/5	
	13	D7-5	(DBH77 - DBH75)/3	
	14	D6-5	(DBH76 - DBH75)/2	!
	15	B5-0	(BA75 - BA70)/5	
	16	B7-5	(BA77 - BA75)/3	
	17	B6-5	(BA76 - BA75)/2	
	18	V7-5	(Volume 77 - Volume 75)/3	
	19	V6-5	(Volume 76 - V-lume 75)/2	•

Table C-3
Treatment means all trees

Trea	tment: Low	Level Sludge	100		
VAR 1 2 3	LABEL DB70 DB75	MEAN 164.9691 178.1649 186.1443	STD-DEV 59.2251 63.8735 68.7330	MIN 7.6000E 01 8.3000E 01 8.4000E 01	MAX 4.9000E 02 5.2700E 02 5.4500E 02
4 5 6 7 8	DB76 DB77 BA70 BA75 BA76 BA77	187.3918 0.0242 0.0281 0.0309 0.0321	70.7252 0.0215 0.0250 0.0274 0.0283	8.4000E 01 5.0000E-03 5.0000E-03 6.0000E-03	5.4800E 02 1.8900E-01 2.1800E-01 2.3300E-01 2.3600E-01
9 10 11 12 13	V75 V76 V77 D5-0 D7-5	0.2753 0.3220 0.3363 2.6392 3.7423	0.3016 0.3471 0.3564 1.1044 2.8122	3.2000E-02 3.4000E-02 3.4000E-02 6.0000E-01	2.7030E 00 3.0350E 00 3.0690E 00 7.4000E 00
14 15 16 17 18 19	D6-5 B5-0 B7-5 B6-5 V7-5 V6-5	3.9897 0.0008 0.0013 0.0014 0.0203 0.0233	2.9747 0.0007 0.0013 0.0014 0.0202 0.0244	0. 0. 0. 0. 3.3333E-04 5.0000E-04	1.2000E 01 5.8000E-03 6.0000E-03 7.5000E-03 1.2200E-01 1.6600E-01

Treat	ment: Low	Level Sludge +	Water 100W		ı
VAR	LABEL	MEAN	STD-DEV	MIN	MAX
1	DB70	155.7451	53.4202	9.0000E 01	3.0300E 02
2	DB75	166.7451	59.3599	9.3000E 01	3.3100E 02
3	DB76	174.3922	65. 8076	9.4000E 01	3.4700E 02
4	DB77	177.2941	68.3580	9.4000E 01	3.5400E 02
5	BA70	0.0212	0.0158	6.0000E-03	7.2000E-02
6	BA75	0.0245	0.0189	7.0000E-03	8.6000E-02
7	BA76	0.0272	0.0221	7.0000E-03	9.5000E-02
8	BA77	0.0283	0.0234	7.0000E-03	9.8000E-02
9	V75	0.2466	0.2169	4.6000E-02	9.0600E-01
10	V76	0.2898	0.2690	4.6000E-02	1.1080E 00
11	V77	0.3040	0.2855	4.6000E-02	1.1430E 00
12	n5-0	2.2000	1.2731	6.0000E-01	6.2000E 00
13	D7-5	3.5163	3.4133	0.	1.2000E 01
14	D6-5	3.8235	3.6672	0.	1.3000E 01
15	B5-0	0.0007	0.0006	0.	2.8000E-03
16	B7-5	0.0012	0.0016	٥.	6.3333E-03
17	B6-5	0.0013	0.0017	0.	6.5000E-03
18	V7-5	0.0191	0.0238	-3.333E-04	9.1333E-02
19	·V6-5	0.0216	0.0269	-5.0000E-04	1.0900E-01
					(cont d)

Table C-3 (continued)

Trea	tment: Mid	Level Sludge	200	1	
Trea VAR 1 2 3 4 5 6 7 8 9 10 11 12 13 14 15	LABEL DB70 DB75 DB76 DB77 BA70 BA75 BA75 BA77 V75 V76 V77 D5-0 D7-5 B5-0	MEAN 142.3472 153.7292 160.3750 162.7361 0.0180 0.0213 0.0234 0.0242 0.2029 0.2400 0.2498 2.2764 3.0023 3.3229 0.0006	STD-DEV 52.7329 59.0864 64.2980 66.1230 0.0135 0.0164 0.0188 0.0196 0.1747 0.2141 1.3812 2.6284 2.9313 0.0006	MIN 5.2000E 01 5.2000E 01 5.3000E 01 2.0000E-03 2.0000E-03 2.0000E-03 6.0000E-03 7.0000E-03 0.000E-03	MAX 2.8300E 02 3.1300E 02 3.3500E 02 3.4000E 02 6.3000E-02 7.7000E-02 9.1000E-02 7.7900E-01 9.6000E-01 9.7000E-01 6.6000E 00 1.0333E 01 1.1000E 01 2.8000E-03
16 17 18 19	87-5 86-5 V7-5 V6-5	0.0010 0.0011 0.0156 0.0185	0.0011 0.0012 0.0169 0.0203	0. 0. -3.3333E-04 -5.0000E-04	4.6667E-03 5.5000E-03 6.3667E-02 9.0500E-02

Trea	tment: Mid	Level Sludge +	Water 200W		
VAR	LABEL	MEAN	STD-DEV	MIN	MAX
1 - 1	DB70	140.7500	60.1918	4.5000E 01	2.9800E 02
2	DB75	151.0294	67.5700	4.5000E 01	3.2800E 02
3	DB76	158.8824	74.0146	4.7000E 01	3.4600E 02
4	DB77	162.6471	77.0628	4.8000E 01	3.5500E 02
- 5	BA70	0.0184	0.0161	2.0000E-03	7.0000E-02
. 6	BA75	0.0214	0.0194	2.0000E-03	8.4000E-02
7	BA76	0.0241	0.0226	2.0000E-03	9.4000E-02
8	BA77	0.0253	0.0240	2.0000E-03	9.9000E-02
9	V75	0.2092	0.2141	6.0000E-03	9.0000E-01
10	V76	0.2479	0.2615	7,0000E-03	1.0600E 00
11	V77	0.2633	0.2775	8.0000E-03	1.1160E 00
12	D5-0	2.0559	1.4842	0.	6.0000E 00
13	D7-5	3.8725	3.4804	0.	1.2667E 01
14	D6-5	3.9265	3.5305	0.	1.3500E 01
15	B5-0	0.0006	0.0007	0.	2.8000E-03
16	B7-5	0.0013	0.0016	0.	6.3333E-03
17	B6-5	0.0013	0.0017	0.	6.5000E-03
18	V7-5	0.0181	0.0221	3.3333E-04	8.5000E-02
19	V6-5	0.0193	0.0244	5.0000E-04	1.0150E-01
					(cònt'd)

Table C-3 (continued)

Treats	ent: High	Level Sludge	300	*	
VAR 1 2 3 4 5 6 7 8 9 10 11 12 13 14 15 6 17	LAREL DB70 DB75 - DB76 DB77 BA70 BA75 BA76 BA77 V75 V76 V77 D5-0 D7-5 D6-5 B6-5	MEAN 161.4757 173.8696 181.4348 184.7565 0.0233 0.0271 0.0297 0.0308 0.2640 0.3101 0.3246 2.4748 3.6290 3.7826 0.0007 0.0013 0.0013	STD-DEV 60.3884 65.1306 70.1706 72.0171 0.0181 0.0212 0.0238 0.0248 0.2319 0.2783 0.2896 1.0821 2.6618 2.8478 0.0006 0.0013 0.0014	MIN 7.3000E 01 7.9000E 01 8.1000E 01 8.4000E 03 5.0000E-03 5.0000E-03 3.2000E-02 3.3000E-02 3.7000E-02 4.0000E-01 0.	MAX 3.3400E 02 3.6000E 02 3.7900E 02 3.8500E 02 8.8000E-01 1.0200E-01 1.1300E-01 1.1170E 00 1.3160E 00 1.3070E 00 6.0000E 00 9.3333E 00 1.0500E 01 3.2000E-03 5.3333E-03 6.0000E-03
18 19	V7−5 V6−5	0.0202 0.0230	0.0202 0.0236	0.	8.4333E-02 1.0100E-01

Treat	ment: High	Level Sludge +	Water 300	W	•
VAR	LABEL	MEAN	STD-DEV	MIN	MAX
1	DB70	159.3559	54,9259	6.7000E 01	2.8100E 02
2	DB75	170.2373	60.8810	6.8000E 01	3.0500E 02
			66.5959	6.8000E 01	3.2200E 02
3	DB76	177.6949		7.0000E 01	3.2800E 02
4	DB77	181.5763	69.5352		6.2000E-02
5	BA70	0.0222	0.0147	4.0000E-03	
6	BA75	0.0256	0.0175	4.0000E-03	7.3000E-02
7	BAZÓ	0.0282	0.0200	4.0000E-03	8.1000E-02
é	BA77	0.0296	0.0215	4.0000E-03	8.4000E-02
	V75	0.2522	0.1903	2.3000E-02	7.6300E-01
9	• • -	, • •	0.2331	2.3000E-02	8.9500E-01
10	V76	0.2964			9.5400E-01
11	U77	0.3132	0.2496	2.5000E-02	
12	D5-0	2.1763	1.2172	2.0000E-01	4.8000E 00
13	บ7−5	3.7797	3.2149	0.	1.0333E 01
14	D6-5	3.7288	3.1980	0.	1.0500E 01
15	B5-0	0.0007	0.0006	0.	2.2000E-03
16	B7-5	0.0013	0.0014	0.	4.6667E-03
17	B6-5	0.0013	0.0014	0.	4.5000E-03
			0.0207	0.	7.3000E-02
18	V7-5	0.0203			7.7500E-02
19	V6-5	0.0221	0.0221	0.	(cont'd)

Table C-3 (continued)

VAR	LABEL	MEAN	STD-DEV	MIN	MAX
1	DB70	180.1042	69.0003	5.3000E 01	3.8300E 02
2	DB75	192.9583	72.9698	5.6000E 01	4.0300E 0 2
3	DB76	200.5313	78.1973	5.7000E 01	4.2100E 02
4	DB77	203.5729	80.7160	5.7000E 01	4.3000E 02
, 5	BA70	0.0292	0.0226	2.0000E-03	1.1500E-01
. 6	BA75	0.0334	0.0254	2.0000E-03	1.2800E-01
7	BA76	0.0363	0.0281	3.0000E-03	1.3900E-01
8	BA77	0.0376	0.0295	3.0000E-03	1.4500E-01
9	V75	0.3381	0.2807	1.0000E-02	1.3600E 00
10	V76	0.3958	0.3322	1.0000E-02	1.5620E 00
11	V77 -	0.4120	0.3503	1.0000E-02	1.6860E 00
12	D5-0	2.5708	0.9629	6.0000E-01	5.6000E 00
13	D7-5	3.5382	2.9635	0.	1.0333E 01
14	D6-5	3.7865	3.0015	0.	1.1000E 01
15	B5-0	0.0008	0.0006	0.	3.0000E-03
16	B7-5	0.0014	0.0015	0.	5.6667E-03
17	B6-5	0.0015	0.0015	ō.	5.5000E-03
18	V7-5	0.0246	0.0243	-6.6667E-04	1.0867E-01
19	V6-5	0.0289	0.0269	-1.0000E-03	1.0100E-01
Treat	ment: Unir	rigated control			
		-,			
.VAR			STD-DEU	MTN	MAV
1	LABEL DB70	MEAN	STD-DEV	MIN 1.0700F 02	MAX
	LABEL		58.0644	1.0700E 02	3.0900E 02
1 2	LABEL DB70	MEAN 203.0789 216.6579	58.0644 63.8928	1.0700E 02 1.1200E 02	3.0900E 02 3.3400E 02
1 2 3 4	Label DB70 DB75	MEAN 203.0789	58.0644 63.8928 68.6226	1.0700E 02 1.1200E 02 1.1200E 02	3.0900E 02 3.3400E 02 3.4800E 02
1 2 3 4 5	LABEL DB70 DB75 DB76	MEAN 203.0789 216.6579 224.5526	58.0644 63.8928	1.0700E 02 1.1200E 02 1.1200E 02 1.1300E 02	3.0900E 02 3.3400E 02 3.4800E 02 3.5600E 02
1 2 3 4 5 6	Label DB70 DB75 DB76 DB77	MEAN 203.0789 216.6579 224.5526 228.7632	58.0644 63.8928 68.6226 71.1414	1.0700E 02 1.1200E 02 1.1200E 02 1.1300E 02 9.0000E-03	3.0900E 02 3.3400E 02 3.4800E 02 3.5600E 02 7.5000E-02
1 2 3 4 5 6 7	LAREL DB70 DB75 DB76 DB77 BA70	MEAN 203.0789 216.6579 224.5526 228.7632 0.0348	58.0644 63.8928 68.6226 71.1414 0.0187	1.0700E 02 1.1200E 02 1.1200E 02 1.1300E 02 9.0000E-03 1.0000E-02	3.0900E 02 3.3400E 02 3.4800E 02 3.5600E 02 7.5000E-02 8.8000E-02
1 2 3 4 5 6 7 8	LABEL DB70 DB75 DB76 DB77 BA70 BA75	MEAN 203.0789 216.6579 224.5526 228.7632 0.0348 0.0400 0.0432 0.0450	58.0644 63.8928 68.6226 71.1414 0.0187 0.0221	1.0700E 02 1.1200E 02 1.1200E 02 1.1300E 02 9.0000E-03 1.0000E-02 1.0000E-02	3.0900E 02 3.3400E 02 3.4800E 02 3.5600E 02 7.5000E-02 8.8000E-02 9.5000E-02
1 2 3 4 5 6 7 8 9	LABEL DB70 DB75 DB76 DB77 BA70 BA75 BA76 BA77 V75	MEAN 203.0789 216.6579 224.5526 228.7632 0.0348 0.0400 0.0432	58.0644 63.8928 68.6226 71.1414 0.0187 0.0221 0.0246	1.0700E 02 1.1200E 02 1.1200E 02 1.1300E 02 9.0000E-03 1.0000E-02	3.0900E 02 3.3400E 02 3.4800E 02 3.5600E 02 7.5000E-02 8.8000E-02 9.5000E-02
1 2 3 4 5 6 7 8 9	LABEL DB70 DB75 DB76 DB77 BA70 BA75 BA76 BA77 V75	MEAN 203.0789 216.6579 224.5526 228.7632 0.0348 0.0400 0.0432 0.0450	58.0644 63.8928 68.6226 71.1414 0.0187 0.0221 0.0246 0.0260	1.0700E 02 1.1200E 02 1.1200E 02 1.1300E 02 9.0000E-03 1.0000E-02 1.0000E-02	3.0900E 02 3.3400E 02 3.4800E 02 3.5600E 02 7.5000E-02 8.8000E-02 9.5000E-01 9.2000E-01
1 2 3 4 5 6 7 8 9 10 11	LABEL DB70 DB75 DB76 DB77 BA70 BA75 BA76 BA77 V75	MEAN 203.0789 216.6579 224.5526 228.7632 0.0348 0.0400 0.0432 0.0450 0.4075	58.0644 63.8928 68.6226 71.1414 0.0187 0.0221 0.0246 0.0260 0.2391	1.0700E 02 1.1200E 02 1.1200E 02 1.1300E 02 9.0000E-03 1.0000E-02 1.0000E-02 1.0000E-02 8.4000E-02	3.0900E 02 3.3400E 02 3.4800E 02 3.5600E 02 7.5000E-02 8.8000E-02 9.5000E-02 1.0000E-01 9.2000E-01
1 2 3 4 5 6 7 8 9 10 11 12	LABEL DB70 DB75 DB76 DB77 BA70 BA75 BA76 BA77 V75 V76 V77 D5-0	MEAN 203.0789 216.6579 224.5526 228.7632 0.0348 0.0400 0.0432 0.0450 0.4075 0.4633	58.0644 63.8928 68.6226 71.1414 0.0187 0.0221 0.0246 0.0260 0.2391 0.2787	1.0700E 02 1.1200E 02 1.1200E 02 1.1300E 02 9.0000E-03 1.0000E-02 1.0000E-02 1.0000E-02 8.4000E-02	3.0900E 02 3.3400E 02 3.4800E 02 3.5600E 02 7.5000E-02 8.8000E-02 9.5000E-02 1.0000E-01 9.2000E-01 1.0400E 00 1.0960E 00
1 2 3 4 5 6 7 8 9 10 11 12 13	LABEL DB70 DB75 DB76 DB77 BA70 BA75 BA76 BA77 V75 V76 V77 D5-0 D7-5	MEAN 203.0789 216.6579 224.5526 228.7632 0.0348 0.0400 0.0432 0.0450 0.4075 0.4633 0.4848 2.7158 4.0351	58.0644 63.8928 68.6226 71.1414 0.0187 0.0221 0.0246 0.0260 0.2391 0.2787 0.2968	1.0700E 02 1.1200E 02 1.1200E 02 1.1300E 02 9.0000E-03 1.0000E-02 1.0000E-02 1.0000E-02 8.4000E-02 8.6000E-02	3.0900E 02 3.3400E 02 3.4800E 02 3.5600E 02 7.5000E-02 8.8000E-02 9.5000E-01 9.2000E-01 1.0400E 00 1.0960E 00 5.0000E 00
1 2 3 4 5 6 7 8 9 10 11 12 13 14	LABEL DB70 DB75 DB76 DB77 BA70 BA75 BA76 BA77 V75 V76 V77 D5-0 D7-5 D6-5	MEAN 203.0789 216.6579 224.5526 228.7632 0.0348 0.0400 0.0432 0.0450 0.4675 0.4633 0.4848 2.7158 4.0351 3.9474	58.0644 63.8928 68.6226 71.1414 0.0187 0.0221 0.0246 0.0260 0.2391 0.2787 0.2968 1.3296	1.0700E 02 1.1200E 02 1.1200E 02 1.1300E 02 9.0000E-03 1.0000E-02 1.0000E-02 1.0000E-02 8.4000E-02 8.6000E-02 8.8000E-02	3.0900E 02 3.3400E 02 3.4800E 02 3.5600E 02 7.5000E-02 8.8000E-02 9.5000E-02 1.0000E-01 9.2000E-01 1.0400E 00 1.0960E 00
1 2 3 4 5 6 7 8 9 10 11 12 13 14 15	LABEL DB70 DB75 DB76 DB77 BA70 BA75 BA76 BA77 V75 V76 V77 D5-0 D7-5 D6-5 B5-0	MEAN 203.0789 216.6579 224.5526 228.7632 0.0348 0.0400 0.0432 0.0450 0.4675 0.4633 0.4848 2.7158 4.0351 3.9474 0.0010	58.0644 63.8928 68.6226 71.1414 0.0187 0.0221 0.0246 0.0259 0.2787 0.2787 0.2968 1.3296 2.7188	1.0700E 02 1.1200E 02 1.1200E 02 1.1300E 02 9.0000E-03 1.0000E-02 1.0000E-02 8.4000E-02 8.6000E-02 8.6000E-01 3.3333E-01 0.	3.0900E 02 3.3400E 02 3.4800E 02 3.5600E 02 7.5000E-02 9.5000E-02 1.0000E-01 9.2000E-01 1.0400E 00 1.0960E 00 5.0000E 00 9.3333E 00
1 2 3 4 5 6 7 8 9 10 11 12 13 14 15 16	LABEL DB70 DB75 DB76 DB77 BA70 BA75 BA76 V75 V76 V77 D5-0 D7-5 B5-0 B7-5	MEAN 203.0789 216.6579 224.5526 228.7632 0.0348 0.0400 0.0432 0.0450 0.4075 0.4633 0.4848 2.7158 4.0351 3.9474 0.0010 0.0017	58.0644 63.8928 68.6226 71.1414 0.0187 0.0221 0.0246 0.0260 0.2391 0.2787 0.2968 1.3268 1.3296 2.77188 2.7773 0.0007	1.0700E 02 1.1200E 02 1.1200E 02 1.1300E 02 9.0000E-03 1.0000E-02 1.0000E-02 8.4000E-02 8.6000E-02 6.0000E-01 3.3333E-01 0.	3.0900E 02 3.3400E 02 3.4800E 02 3.5600E 02 7.5000E-02 8.8000E-02 1.0000E-01 9.2000E-01 1.0400E 00 1.0960E 00 5.0000E 00 9.3333E 00 9.5000E 00
1 2 3 4 5 6 7 8 9 10 11 12 13 14 15 16 17	LABEL DB70 DB75 DB76 DB77 BA70 BA75 BA76 BA77 V75 V76 V77 D5-0 D7-5 D6-5 B7-5 B6-5	MEAN 203.0789 216.6579 224.5526 228.7632 0.0348 0.0400 0.0432 0.0450 0.4075 0.4633 0.4848 2.7158 4.0351 3.9474 0.0010 0.0017 0.0016	58.0644 63.8928 68.6226 71.1414 0.0187 0.0221 0.0246 0.0260 0.2391 0.2787 0.2968 1.3296 2.7188 2.7773 0.0007 0.0014	1.0700E 02 1.1200E 02 1.1200E 02 1.1300E 02 9.0000E-03 1.0000E-02 1.0000E-02 8.4000E-02 8.6000E-02 8.6000E-01 3.3333E-01 0.	3.0900E 02 3.3400E 02 3.4800E 02 3.5600E 02 7.5000E-02 6.8000E-02 1.0000E-01 9.2000E-01 1.0400E 00 1.0960E 00 5.0000E 00 9.3333E 00 9.5000E 00 2.6000E-03
1 2 3 4 5 6 7 8 9 10 11 12 13 14 15 16 17 18	LABEL DB70 DB75 DB76 DB77 BA70 BA75 BA76 BA77 V75 V76 V77 D5-0 D7-5 B6-5 B7-5 V7-5	MEAN 203.0789 216.6579 224.5526 228.7632 0.0348 0.0400 0.0432 0.0450 0.4075 0.4633 0.4848 2.7158 4.0351 3.9474 0.0010 0.0017 0.0016 0.0258	58.0644 63.8928 68.6226 71.1414 0.0187 0.0221 0.0246 0.0260 0.2391 0.2787 0.2968 1.3296 2.7188 2.7773 0.0007 0.0014 0.0014	1.0700E 02 1.1200E 02 1.1200E 02 1.1300E 02 9.0000E-03 1.0000E-02 1.0000E-02 8.4000E-02 8.6000E-02 6.0000E-01 3.3333E-01 0.	3.0900E 02 3.3400E 02 3.5600E 02 7.5000E-02 8.8000E-02 9.5000E-01 9.2000E-01 1.0400E 00 1.0960E 00 5.0000E 00 9.3333E 00 9.5000E 00 2.6000E-03 4.6667E-03
1 2 3 4 5 6 7 8 9 10 11 12 13 14 15 16 17	LABEL DB70 DB75 DB76 DB77 BA70 BA75 BA76 BA77 V75 V76 V77 D5-0 D7-5 D6-5 B7-5 B6-5	MEAN 203.0789 216.6579 224.5526 228.7632 0.0348 0.0400 0.0432 0.0450 0.4075 0.4633 0.4848 2.7158 4.0351 3.9474 0.0010 0.0017 0.0016	58.0644 63.8928 68.6226 71.1414 0.0187 0.0221 0.0246 0.0260 0.2391 0.2787 0.2968 1.3296 2.7188 2.7773 0.0007 0.0014	1.0700E 02 1.1200E 02 1.1200E 02 1.1300E 02 9.0000E-03 1.0000E-02 1.0000E-02 8.4000E-02 8.4000E-02 8.6000E-01 3.3333E-01 0.	3.0900E 02 3.3400E 02 3.5600E 02 7.5000E-02 8.8000E-02 9.5000E-01 9.2000E-01 1.0400E 00 1.0960E 00 9.3333E 00 9.5000E 00 2.6000E-03 4.6667E-03 4.5000E-03

Table C-4
Treatment means paired trees

Treatm	ent: Lo	w Level Sludge	100		I
'VAR	LABEL	MEAN	. STD-DEV	MIN	MAX
1	DB70	186.6000	53.8796	8.4000E 01	2.7800E 02
2	DB75	200.6286	59.1136	9.1000E 01	3.0300E 02
3	- " DB 7 6	210.9429	64.7597	9.3000E 01	3.2200E 02
4	DB 77	215.2571	67.1015	9.6000E 01	3.2900E 02
5	BA70	0.0296	0.0158	6.0000E-03	6.1000E-02
6 🗂	** BA75	0:0344	0:0137	7:0000E-03	7.2000E-02
7.	. BA76	0.0382	0.0214	7.0000E-03	8.1000E-02
3	BA77	0.0399	0.0228	7.0000E-03	8.5000E-02 '
٠٠	775 V75	0.3339	0.2039	4.2000F-02	7.3500E-01
10.	V76	0.4017	0.2434	4.6000E-02	8.8400E-01
11	V <i>7</i> 7	0.4233	0.2622	5.1000E-02	9.1500E-01
12	. b2-0		1:3271	6.00 0 0E-01"	
13	D7 - 5	4.8762	2.9485	3.333E-01	1.0000E 31
14	D6-5	5.1571	3.1872	5.0000E-01	1.2000E 01
-15	··· 85-0	0.0010	0.0006	0:	7.2000E-03
16	87- 5	0.0018	0.0014	O.	4.6667E-03
17	8 5- 5	0.0019	0.0015	·O.	5.5000E-03
18.	٧7 ,− 5	~~~0:0282	0.0204	1.6667E-03	7.1333E-02
19	V 6- 5	0.0314	0.0233	2.0000E-03	9.1500E-02

Treat	ment: Low	Level Sludge	Water 100W		· •-
VAR	LABEL	MEAN	STU-DEV	MIN	WAX
1	ก870	170.6296	64.1172	9.0000E 01	3.0300E 02
2	DB75	182.5926	70.9617	9.3000E 01	3.3100E 02
3	DB76	192.2963	78.0320	9.4000E 01	3.4700E 02
4	DB 77	195.7037	80.8564	9.4000E 01	3.5400E U2
5	BA70	0.0259	0.0195	6.0000E-03	7.2000E-02
გ	2 A75	0.0301	0.0231	7.0000E-03	8.6000E-02
7 8	BA76	0.0337	0.0270	7.0000E-03	9.5000E-02
8	BA77	0.0350	0.0284	7.0000E-03	9.3000E-02
y	V7 5	0.3107	0.2661	4.6000E=02	9.0600E-01
10	V76	0.3703	0.3283	4.6000E-02	1.108QE 00
11	V77	0.3872	0.3460	4.6000E-02	1.1430E 00
12	D5-0	2.3926	1.4342	6.0000E-01	5.6000@ OJ
13	D7-5	4.3704	3:6519	3.333E-01	1.20008 01
14	D6-5	4.8519	3.8676	5.0000E-01	1.3000E 01
15	85 -0	0.0008	0.0007	0	2.80001-03
16	87-5	0.0 016	0.0019	υ.	6.3333E-03
17	B6 - 5	0.0018	0.0021	0.	6.5000E-03
८१	V7- 5	0.0255	0.0274	_0.	9.1333E-02
19	V6-5	0.0298	0.0319	-0.	1.0900E-01
	•		•	•	(cont'd)

Table C-4 (continued)

Treatm	ent: Mid	Level Sludge	200		
VAR	LABEL	MEAN	. STD-DEV	MIN	MAX
1	D870	170.7568	48.1805	9.0000E 01	2.5400E 02
2	DB75	184.2973	54.3455	9.4000E 01	2.7500E 02
; -3·	``DB76	79571081	59:4730	9.6000E 01	2.9200E 02
4 -	. D B 77	198.4324	61.6079	9.6000E 01	. 2.9500E 02
5.	BA70	0.0246	0.0125	6.0000E-03	5.1000E-02
		- 0.0289	0.0151	7.0000E-03	5.900UE-02
· 7.	B A76	0.0326	0.0176	7.0000E-03	6.7000E-02
8	BA77	0.0338	. 0.0134	7.0000E-03	6.8000E-02
25	· V75 ····	0.2843	0.1617	"4.9000E-02"	~5.7700E-01
10	V76	0.3440	0.2016	5.3000E-02	7.0600E-01
11	V 77	0.3586	0.2119	5.3000E-02	7.31 OOE-01
12	D5-0	2.7081	1.3438	6.0000E-01	5.0000E 00
13	D7 - 5	4.7117	2 . 77 2 4	3.333E-01	1.0333E 01
14	D6-5	5.4054	2.9717	5.0000E-01	1.1000E 01
~15	B5-0°	.00009.	0:0006	0.	1.8000E-03"
ló	87- 5	0.0016	0.0012	O.	4.000UE-03
17	Bo-5	0.0018	0.0013	0.	4.5000E-03
18	√7-5``	0.0248	0.0173	6.6667E-04	5.9333E-02
19	¥6 − 5	0.0299	0.0208	1.0000E-03	6.9500E-02

Treatm	ent: Mid	Level Sludge + Wa	eter 200W	•	
VAR	LABEL	MEAH	STD-DEV	MIN	MAX
1.	DB 70	154.9630	50.8841	8.6000E 01	2.5300E 02
2.	DB75	166.8889	57.1081	8.9000E 01	
<u>2</u> . 3	DB76	175.9259	63.9945	9.0000E 01	2.7700E 02
4	DB 77	180.1852	67.3853	9.0000E 01	3.0400E 02
5	BA70			9.0000E 01	3.1500E 02
- 4	BA75	0.0208	0.0132	6.0000E-03	5.0000E-02
-·· 🛪 ··	BA76	0.0243	0.0160	ç•0000E-03	6.0000E-02
ုံ	BA 77	0.02 74	0.0191	6.0000E-03	7.3000E-02
9	V75	0.0239	0.0206	6.0000E-03	7.8000E-02
		0.2385	0.1746	_4,400 0 E-02_	6.41 00E-01
10	~~\\76 ⁻	0.2835	0.2189	4.7000E-02	8.1700E-01
11	٧77	0.3030	0.2391	4.7000E-02	8.7500E-01
12	D5-0	2.3852	1.2501	6.0000E-01	4.8000E 00
13	D7-5	~~4.4321	3.6219	3.3333E-01	1.2667E 01
14	D 6- 5	4.5185	3.6387	0.	1.3500E J1
15	B5-0	0.0007	0.0006	o.	2.0000E-03
13	67-5	0.0015	0.0016	- 5.	
17	B 6- 5	0.0016	0.0016	Ů.	6.0000E-03
18	V7-5	0.0215	0.0223		6.5000E-03
19	V6-5	0.0215		1.0000E-03	7.8000E-02
7	. 40-5	0.0225	0.0228	5.0000E-04	8.8000E-02
		• •			(cont.td)

Table C-4 (continued)

Treat	ment:	High Level	Sludge	300		
VAR	LABEL		MEAN	STD-DEV	MIN	MAX
1	DB7		2973	64.4601	8.1000E 01	3.0100E 02
2	DB7	5 206	8649	70.9570	-814000E 01	3.2500E 02
3	DB7	6 217	.0811	76.2647	8.5000E 01	3.4200E 02
4.	DB7	7 221	4324	77.9550	8.5000E 01	3.4700E 02
5	Ba7		0324	0.0195	5.0000E-03	7.1000E-02
6.	BA7	5 0.	.0374	0.0230	6.0000E-03	8.3000E-02
7	BA7	'δ '0.	.0415	0. 0259	6.0000E-03	9.2000E-02
B	BA7		.0432~	0.0269	" 6:0000E-03	9.5000E-02
9	V75		.3779	0.2508	3.8000E-02	8.7400E-01
10	Ÿ76	0	4458	0.3010	3.8000E-02	
- 11	- 'V77	0	4695	0.3159	- 3:3000E-02	
12	05-	·0 2	.7135	1.3321	6.0000E-01	5.0000E 00
13	D7-	-5 4	8559	2.6811	3.333E-01	8.6667E 00
14	D6-		.1081-	2.9654	5:0000E-01	
15	B5-	-o o	.0010	0.0007	2.0000E-04	
16	B7-	-5 0	.0019	0.0014	0.	4.6667E-03
-17-	B6-	-	:0020	0:0015	0.	5.0000E-03
18	٧7-	-5 0	.0305	0.0227	0.	8.4333E-02
19	V6-	-5 0	.0340	0.0257	0.	8.6000E-02

Treatme	ent: High	Level Sludge + Wa	ter 300W		
			STD-DEV	MIN	MAX
VAR	LABEL		53.7146	9.0000E 01	2.8100E 02
ī	D270" -		59.9348	9.3000E 01	3.0500E 02
2	DB75		66.2471	9.4000E 01	3.2200E 02
3	DB76		69:2220-	9.4000E 01	-3:2800E 02
4	DS 77	0.0283	0.0157	6.0000E-03	6.2000E-02
5	BA70	0.0283	0.0187	7.0000E-03	7.3000E-02
6	BA75		0.0216	7.0000E=03	8.1000E-02
-7	-BA76-	0.0383	0.0210	7.0000E-03	8.4000E-02
8	BA77	0.0382	0.2054	5.3000E-02	7.6300E-01
9	V75	0.3294	- 0: 2526	5.5000E-02"	-8.9500E-01
- 10	V76	0:3899	0.2707	5.5000E-02	9.5400E-01
11	V77	0.4133		6.0000E-01	4.800DE 00
12	D5-0	2.6242	1.2784	3:3333E-01	71.0333E 01
13	D7-5	4.8182	3.4991		1.050DE 01
14	D6-5	4.7273	3.5446	0.	2.200DE-03
15	<i>8</i> 5-0	0.0009	0.0006	2.0000E-04	4.6667E-03
. 16	87-5	0.0018	0.0016	0.	4.5000E-03
i7	B6-5	0.0017	0.0016	0.	
iä	V7-5	0.0280	0.0232	6.6667E-04	7.3000E-02
-19	··vò=5	0.0303	~~0.0245~	- 1:0000E=03	7.7500E-02
.,	.,,				(cont'd)

Table C-4 (continued)

Trea	tment: Wat	er only		. 49. ~	
WXR.	"LABEL"	MEAN	STD-DEV	MIM:	MAX
. 1	DB7J	192.5185	62.0089	8.7000E 01	2.9700E 02
2	DB75	206.3704	67.3790	9.3000E 01	3.1900E 02
3	DB73	215.2222	72.4565	9.4000E UI	3.2900E 02
4 .	DB 77	218.6567	74.5773	9.4000E 01	.3.3000E 02
5.	BA70	0.0320	0.0135	6.0000E-03	6.9000E-02
6	BA75***	0.0370	0.0216	7:0000E=03	TT810000E-02
7 ·	BA7ó	0.0404	U.0240	7.0000E-03	8.5000E-02
8	BA77	0.0418	0.0251	7.0000E-03	8.6000E-02
9	" v75"	0.3764	0.2399	4.5000E-02	8.4500E-01
.10	٧76	0.4447	O.2378	4.4000E-02	9.5500E-01
' 11	V 77	0.4600	0.2935	4.4000E-02	9.6100E-01
15	D5-0-	~ ~ ~ 2.7704 ····	1:1391	6.0000E-01	5.0000E 00
.13	D7-5	4.0988	2.7855	3.333E-01	8.6667F 00
14	D6-5	4.4259	2.8730	0.	8.50Q0E 00
_15	" B5-0"	0.0010	0.0002	0.	2.2000E-03
16	B7-5	0.0016	0.0013	0.	4.0000E-03
17	B6-5	0.0017	0.0014	0.	4.0000E-03
18	V7-5	0.0279	0.0210	-6.6667E-04	7:0000E-02
19	V 6-5	0.0341	0.0253	-1:0000E-03	7.9500E-02

Trea	tment: U	nirrigated control			
VAR	LABEL	MEAN	STD-DEV	MIN	MAX
1	DB70	205.6757	56.5843	1.0900E 02	3.0900E 02
2	DB75	219.2162	62,7699	1.1200E 02	_3.3400E 02
3	DB76	227.1892	67.5898	1.1200E 02	3.4800E 02
- 4	DB77	231.4595	7 0.1267	1.1300E 02	3.5600E 02
5	BA <u>70</u> _	0,0355	Q.0184	9.0000E-03	_7.5000E-02
6.	BA75	U.0408	0.0219	1.0000E-02	8.8000E-02
.7 8	BA76	0.0440	Ü.U244	1.0000E-02	9.5000E-02
	BA77	0.0459	0.0258	1.0000E-02	1.0000E-01
٠ 9	V75	0.4158	0.2368	8.4000E-02	9.2000E-01
10	٧7٥	0.4727	0.2764	8.6000E-02	1.0400E 00
. 11		0.4946	0.2946	8.8000E-02	1.0960E .00
12	Ŭ5 − 0	2.7081	1.3471	6.0000E-01	5.0000E 00
. 13	D7-5	4.0811	2.7413	3.333E-01	9.3333€ 00
14	D6-5	3.9865	2.8050	0.	9.5000E 00
15	85-0	0.0010	0.0007	2.0000E-04	2.6000E-03
16	B7-5	0.0017	0.0014	٥.	4.6667E-03
17	86- 5	0.0016	0.0014	0.	.4.5000E-03
18	V7-5	0.0263	0.0200	1.3333E-03	6.9000E-02
19	. V6-5	0.0285	0.0209	1.0000E-03	7.0500E-02
	-	0			(cont'd)

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4. TITLE AND SUBTITLE	5. REPORT DATE March 1980 (Issuing Date)
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7. AUTHOR(S)	8. PERFORMING ORGANIZATION REPORT NO.
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15. SUPPLEMENTARY NOTES

Project Officer: Gerald Stern (513) 684-7654

16. ABSTRACT

A sprinkler irrigation system developed for uniform applications of anerobically digested, municipal-industrial sewage sludge initially applied up to 5.8 mt/ha/wk. Reduced infiltration of sludge occurred due to physically blocking of soil pores, causing ponding of sludge in the micro-depressions. Sludge loading rates were decreased to 10, 20, 30 and 40 mt/ha/yr.

The renovating capactiy of forest soils for most suspended and dissolved constituents in sludge was very good (95 to 99+%). Nitrogen was the exception as nitrification rates increased with increased rates of sludge applications, resulting in leaching of NO3-N and concomitant cation losses in soil wa ter through the surface 2 m of soil. Leaching losses did not alter dissolved ions in ground water at 10 m. Phosphorus in all forms was never found in significant amounts in soil solutions of tested soils.

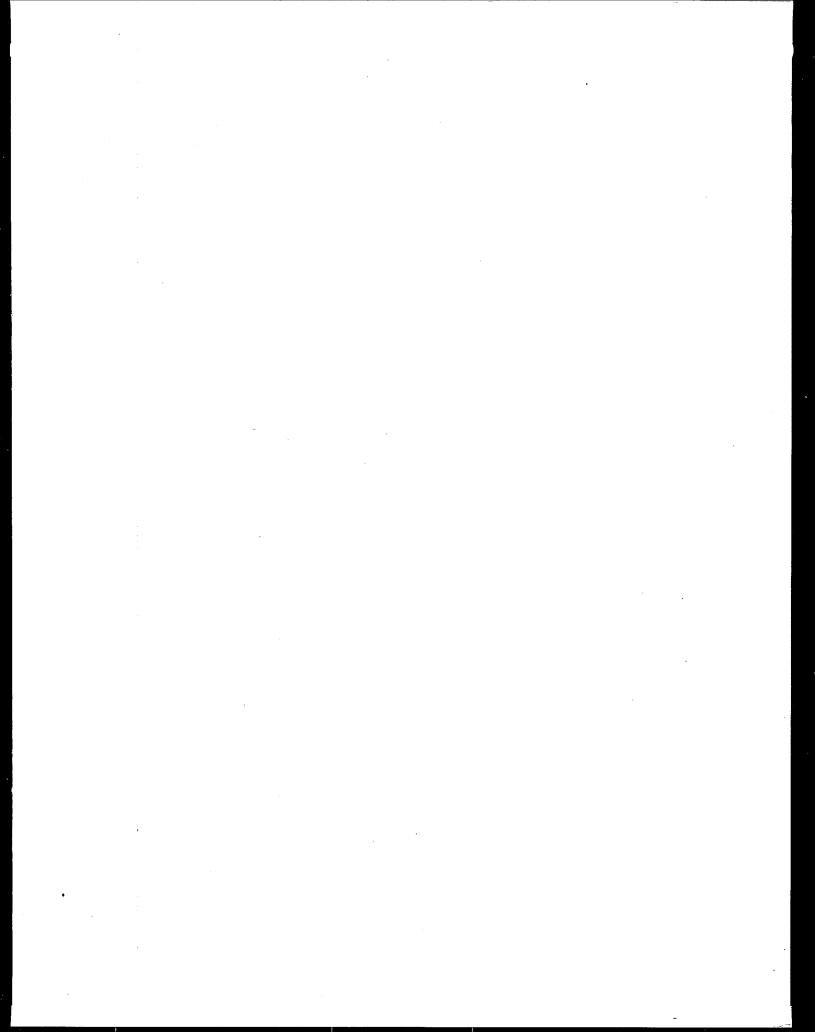
Optimum loading rates of 20 to 30 mt/ha/hr. of sludge show trends to increased

surface soil total N, organic material and cation exchange capacity.

Analyses for virus at all depths in the soil and from the soil solution at corresponding depths were negative, nor were human pathogens of the bacteria type isolated from the limited numbers of soils and soil solutions analyzed.

Sludge applications increased the growth rate of treated trees. Water applied after sludge irrigation also may enhance tree growth over sludge only applications.

17. KEY WORDS AND DOCUMENT ANALYSIS					
DESCRIPTORS	b.IDENTIFIERS/OPEN ENDED TERMS	c. COSATI Field/Group			
Sludge, Sludge disposal, Waste disposal, Forestry, Forest land, Trees (plants), Soils, Soil analysis, Soil chemistry, Irrigation, Sprinkler irrigation, Surface irrigation		13B			
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