AMBIENT AIR ANALYSIS OF BUNKER HILL LEAD EMISSIONS

Prepared for:

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1.0 EXECUTIVE SUMMARY

This study utilizes computer-assisted cartographic modeling to empirically relate ambient lead and cadmium particulate observations to known emission estimates in the Silver Valley of northern Idaho. Through this technique, the many factors influencing ambient lead concentrations were simultaneously considered and quantified. Particular attention was paid to the multiplicity and inconstancy of particlate sources, the confounding effects of complex terrain on local meteorology, and the impact of windblown dusts.

All of the known particulate sources in the valley were included in one of the following five categories, chemical consituency and magnitude estimates were developed from existing reports or direct or inferred measurements: (1.) Industrial Point Sources, (2.) Industrial Process Fugitive Sources, (3.) Industrially-related Active Fugitive Sources, (4.) Transportation-related Active Fugitive Sources, and (5.) Passive (windblown) Fugitive Sources. The first three categories represent emissions from current industrial activities. The latter two categories' emissions are residual in character and result from cumulative effects over the years.

In the first portion of the analysis, cadmium measurements were used as a tracer to quantify the important atmospheric dispersion characteristics through the use of a Gaussian plume analogy. The results indicate that mountain-valley drainage phenomema and associated nocturnal inversions dominate dispersal activity on the majority of days. Under this situation.

the standard Gifford-Pasquill dispersion parameter estimates hold for unstable conditions. As neutral and stable conditions are encountered, dispersion becomes more and more inhibited and Gifford-Pasquill parameter esimates are grossly inadequate. This situation is often observed in complex terrain.

Stability is the critical variable in estimating pollutant dispersion in the Silver Valley. However, there are several meteorological and operational situations that require special treatment. They have to do with special synoptic conditions, stable layers aloft, or smelter operations. Stagnant high pressure areas, limited mixing depths, low wind speeds, and synoptic drainage winds result in severe pollution episodes on about thirty to fifty days per year. Inefficient smelter operations and frequent upsets and malfunctions create severe episodes twenty to thirty days per year.

In the second portion of this study, the dispersion model was applied to all lead sources and model estimates were related to observed ambient levels. Two years of data were used. The result was a self-calibrated empirical expression of lead impacts at monitored locations. Four source groups were found to significantly impact ambient lead concentrations. They were:

Low-level smelter sources, or those that emanate from below forty feet, and exhibit their greatest impact within one-half mile of the smelter. Most are associated with the sintering operation or the ore preparation area. These sources constitute over seventy-five percent of the lead impact in this zone and absolute model estimates are as high as 6.5 ug/m3 on a quarterly basis. Mid-level smelter sources or those that emanate from forty to one hundred forty feet and exhibit their greatest impact at about two miles from the smelter. Blast furnace upsets and pellet dryer emissions constitute the bulk of this category. Their maximum modeled estimate was 3.5 ug/m3 at two miles.

Active fugitive sources, generally road dust and sinter storage, and Passive fugitive sources (windblown dusts) exert small but significant lead impact at several stations. The former accounts for less than five percent of total impact and the latter as much as twelve percent in summer months at non-attainment locations.

Current smelter emissions from the low and mid-level sources account for at least eighty-five percent of total lead impact within eight miles of the smelter. Tall stack emissions are insignificant in comparison. The highest ambient concentrations occur in winter months and are associated with prolonged periods of atmospheric stability. It is most important to remember that the different sources' impacts vary with location and season and that their combined effect determines the limiting situation for standard attainment. Control of upsets and malfunctions is also critical to the development of any implementation strategy. These items are discussed in detail in the final section of the report.

2.0 CONCLUSIONS

- Meeting the NAAQS in the Silver Valley may be extremely difficult for the following reasons:
 - a. The standard is small (less than 10%) when compared to current and historical lead levels.
 - b. There are several types of sources capable of significantly impacting the NAAQS concentration.
 - c. The maximum effects of these different sources occur at different locations and during different seasons, consequently all must be treated.
 - d. Ambient concentrations are very sensitive to severe events, both meteorological and operational; any control strategy must be capable of accomodating serious upsets or prolonged stagnations.
 - e. There are numerous residual pollutant effects that are not well understood.
- 2. Source reductions will have to be accomplished in several source categories and all to high degrees.
 - a. 80% to 90% reductions in all low and mid-level smelter point and process fugitive sources.
 - b. A comprehensive program of controlling and/or eliminating upset conditions, especially at the smelter blast furnace.
 - c. A program of reducing transportation related and sinter storage related active fugitive emissions in and around the smelter.
 - d. A program of surface stabilization of contaminated soils in the airport area, fairgrounds-lumberyard industrial corridor, and Bunker Hill tailings pond embankment.
- 3. Evidence suggests that configuration changes, both in emissions and real property, may be viable alternatives to emission reductions in some locations.
 - a. As tall stack emissions do not have a significant impact in comparison to low and mid-level sources, venting these sources through the main baghouse would significantly reduce impacts. This would be especially pertinent to blast furnace and pellet dryer emissions.

- b. As low-level emissions reductions are almost wholly dependent on their impact at Silver King and Smelterville, this requirement could be significantly reduced were the company to purchase these properties.
- 4. An updated emissions inventory and a detailed analysis of OSHA compliance activities are needed.

3.0 LEAD CONTAMINATION IN SHOSHONE COUNTY, IDAHO

Attainment of the National Ambient Air Quality Standard (NAAQS) for lead will requre substancial control and capture of current lead emissions in the Silver Valley. There are numerous sources of lead to the ambient air in this area and their impact varies considerably with location and season. In depth analyses of the complex source-receptor relationships are necessary. Such studies usually involve diffusion model applications. This area, however, is particularly ill-suited for diffusion modeling.

The study area is an enlogated, narrow, deep valley encompassing the South Fork of the Coeur d' Alene River. Figure 3.1 is a map of the area. Appendix A contains computerized representations of the Valley's industrial and topogapghic features. The valley is subject to adverse meteorological conditions and the regular formation of surface-based inversions and diabatic winds. There are large denuded areas where deposition of water and windborne contaminants have combined with industrial wastes to create large area sources associated with transportation activities and inconstant industrial processes combine with large point sources to effect a complex source spectrum with considerable spatial and temporal variation. The smelting complex is the hub of the industrial community. There are five principal processes within the industrial area: (1.) a galean (Pbs) ore mine, (2.) a lead/zinc milling and concentrating operation, (3.) a lead smelter, (4.) an electrolytic zinc plant, and (5.) a phosphate fertilizer production plant. Vegetation levels and soil characteristics are important to this study. Denuded areas subject to deposition of heavy metals could contribute to lead exposures through

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	Pinehurst	PNH 2		
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	Smelterville	2117 2		
	Silver King School	SKS 4		
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reentrainment of the exposed soils. A variety of vegetation is found in the area. A considerable change in the natural vegetation has resulted from prevalent industrial activities. Much of the original coniferous forest was harvested for timber and fuel. Sulfur dioxide gas from the smelter has damaged or destroyed much of the native vegetation near Kellogg, and has increased soil acidity to levels intolerable for many species. Following a 1910 fire and installation of the present smelter in World War I, natural plant invasion did not occur and young trees did not survive. Most of the burned area outside the smelter influence has been retimbered by second growth lodgepole pine. Bushy plants and grasses cover the hillsides in uneven distribution dependent on moisture availability (University of Idaho, 1974).

Soils in the area vary with their location. On the mountain slopes parent materials consist of decomposed rocks and a thin layer of forest litter. In scattered areas at lower elevations, small amounts of volcanic ash and loess are found (University of Idaho, 1974). In an area of high rainfall and steep slopes devoid of protective vegetation, these materials erode rapidly. Accelerated erosion destroyed many of the acid-resistant plants in the smelter zone. The loss of surface soil and decreased water-holding capacity increased runoff. The result is the bare, severely eroded hills surrounding the smelter. Little or no vegetation is found there. The soils are surface hardened remnants of the native materials.

The valley floor is partially filled with alluvial deposits varying in thickness from a few inches to several feet. The alluvium is principally unconsolidated sand and gravel. Mine and mill tailings have formed a metalladen veneer of silt over large areas of the valley. The use of mine wastes

for fill and construction activities have also resulted in redistribution of soil metals throughout the area (State of Idaho, 1974a, 1975a, 1978b). In Appendix B are representations of vegetation cover and soil metal distributions in the valley.

The climatic conditions show a mean annual temperature of 47.4°F, a mean number of days (138) above 32°F, with average frost dates of May 12 and September 27. Mean annual precipitation is 31 inches ranging from 9.4 inches in January to <0.1 inch in August. Most precipitation is in the form of snow with snow cover prevalent from December to late February. Thunder showers persist from mid-June through July with short rainy seasons in early spring and late September (NOAA, 1976).

4.0 SOURCE IDENTIFICATION AND EMISSIONS EVALUATION

4.1 GENERAL

There are several sources of lead to the atmosphere in the Silver Valley. These sources vary in magnitude, frequency, chemical constituency, and configuration. Some categorization scheme is necessary for purposes of discussion and analysis. Developing reasonable source categories requires certain knowledge of the different sources' behavior and the responsible environmental and industrial processes. Lead/zinc ores have been mined in several locations in the Silver Valley for nearly one hundred years. All currently active mines are east of the smelting complex. Galena (PbS) is the predominant lead ore. The ore is concentrated by wet-chemical milling processes near the mines. Both lead and zinc concentrates are produced. Concentrates are typically 60% metal in the sulfide form and are the consistency of coffee grounds. They are transported to their respective smelters by rail in a wet form.

Direct air pollution from these activities is insignificant. The milling operations are wet and the material is transported before it dries. Concentrates spilled on the roadways constitute an active source of heavy metal particulate after drying. Other residual aspects of the milling operations can have significant air quality effects. Large quantities of tailings containing significant amounts of heavy metals are produced by the milling operations. Currently these tailings are stored in large ponds. The dikes of these ponds are usually made of dried tailings. These metal-laden, low pH, fine sandy materials are not conducive to vegetation,

and the dikes and abandoned ponds can be significant sources of reentrained fugitive dusts.

In the first eighty years of mining, tailings were discharged directly into the river or dumped at convenient locations. Periodically, the river would flood and deposit the waste material on the valley flood plain. These silts are also a source of particulate reentrainment. Tailing sands are typically 1 to 4% lead by weight.

After arriving at the smelters, the concentrates are dried and mixed with other concentrates and residual materials in preparation for pyrometallurgical treatment. Zinc concentrates are roasted at the zinc plant to produce a powdered oxide of zinc. Lead concentrates are "sintered" to produce a lead oxide compound, similar to furnace clinkers, called sinter product. Both of these processes burn the sulfur from the sulfide concentrations as a fuel. This roasting of the ores produces the tremendous quantities of sulfur dioxide associated with non-ferrous smelters. There are several process-fugitive sources associated with the materials handling aspects of these processes. Those before roasting are in association with the ore-preparation and crushing plants in the lead smelter. Drying of the concentrate compounds is a significant point source of particulates. These are all fine particulate emissions, about 30% lead by weight in the sulfide form. Materials handling of zinc concentrates seems to be an insignificant particulate source. Roasting of the ores is a significant particulate source in both smelters. Over two-thirds of the total zinc plant emissions result from ore-roasting. In the smelter, over a quarter of the total lead emissions are associated with the "Lurgi" sinter operations. Most of these are point source emissions from scrubbers attached to the process. Others are

fugitive emissions associated with materials handling of the highly abrasive sinter product. Point source emissions are generally fine particulate. The fugitive source emissions vary from very fine to very large settléable particulate. Lead concentrations are 10% or less in the zinc smelter and 45 to 60% lead in the sinter operations emissions. Both are in the oxide form.

After roasting, base metals are recovered from the metallic oxides. In the zinc plant, the powdered zinc oxide is dissolved in sulfuric acid and zinc is recovered electrolytically. In the lead smelter, the sinter product is combined with coke and smelted in a blast furnace. Oxygenenriched air is "blasted" into the furnace and the coke is burned as a fuel. The coke also supplies carbon as a reducing agent in an oxidation-reduction reaction that results in molten metallic lead. This is a violent operation that produces over one-half of the total lead emissions in the entire complex. Much of it is emitted from the facility's main stack after the particulates are removed in the smelter's primary particulate control facility, the main baghouse. A significant amount is also discharged directly to the atmosphere as a result of upsets in blast furnace operation. Emissions from this stage of the process are fine particulates and metal fume of about 60% lead content in the oxide and pure metallic form.

Following the smelting processes, the base metals move to their respective refineries where they are concentrated to 99.99% pure form. Some fine particulate emissions are associated with refinery processes.

In this analysis the sources are first separated into the gross regulatory categories, point sources, and fugitive sources. Point sources are defined here as controlled sources that emanate from stacks or equivalent

devices. Fugitive sources are all those sources that are not classified as point sources. Four sub-categories of fugitive sources, based on the source's suspension energy, have been developed for these analyses. They are (1) process fugitive sources that are attendant to and derive their suspension energy from industrial processes, (2) active fugitive sources associated with gross materials handling in industrial areas, (3) active emissions arising from transportation activities, and (4) passive fugitive emissions that are reentrained by the wind. All of the sources in the valley were located on computerized maps prepared especially for these analyses. These maps are presented and discussed in Appendix A. References to the appropriate maps are made here.

4.2 POINT SOURCES

All point sources in the valley are inventoried by the NEDS classification system designation number. Table 4.1 shows the source characteristic information available for the 32 point sources identified. The variables in that table are as follows:

NEDS--the National Emission Data System identification number

UNIT--identifies the associated industrial process unit (i.e., 1 crushing plant, 5 sinter or Lurgi operation)

PBFR, CDFR--percent lead and cadmium, respectively, in the particulate emission

EXITVEL--exit velocity of the stack emission (m/s)

STHGHT--physical stack height above ground (m)

STDIAM--physical stack diameter (m)

EXITTEMP--stack gas exit temperature (°K)

VOLFLOW--stack gas volumetric flow rate (m³/s)

YROPPCT--percent of annual operation time.

Table 4.1 Point Source Inventory Information

OBS	UNIT	NEDS	PBFR	CDFR	EXITVEL	STHCHT	STDIAM EXITTEMP VOLFLOW YROPPCT NAME		PSNUH			
1	1	2	31	0							Crushpl Dryer	1
2	1	3	31	1						50	Crushpl Collect	1
3	1	4	31	1	•					50	Crushpl Rodmill	1
4	1	5	31	1		3.1		289	6.62	50	Crushpl Baghouse	1
5	2	6	31	1		9.8		291	3.70	80	Oreprep Baghouse	2
6	3	7	31	1		24.8		310	9.93	75	Pellet Dryer	3
7	5	8	45	1	5.20	6.1	1.07	294	4.72	75	Lurgi D Scrubber	4
8	5	9	45	1		3.0		297	1.66	100	Lurgi N Rotoclon	4
9	5	10	45	1	10.50	6.1	1.07	294	9.45	75	Lurgi B Scrubber	4
10	5	12	4.5	ĩ	10.90	6.1	3.14	312	84.46	85	Lurgi A Scrubber	4
11	5	11	45	ī	10.50	6.1	1.07	294	9.45	75	Lurgi C Scrubber	4
12,	7	16	5	ī	10.90	53.4	3.14	312	84.46	8.5	ZN Fume Main St.	5
13	7	17	5	ĩ	10.90	24.4	0.91	344	7.08	85	ZN Fuming Gran	. 6
14	Ř	32	•	•	10.70		0.74			85	Pbreein Scrubber	7
15	8	1				9.1		292	4.32	75	Reverb Baghouse	7
16	9	15				19.8		323	1926.00	ii	EF Gran Scrubber	7
17	11	í	5.5	7	14.30	217.9	4.15	327	193.10	100	Smelter Hain St.	Ŕ
18	20	18	10	í	9.00	186.0	1.83	311	23.62	50	Zinc Main Stack	9
19	21	33	10	1		100.0	1.05	711	13.02	30	Concent Dryer .	10
20	21	19	10	î							Concent Silo	10
21	22	22	20	1							Rosconv Scrubber	11
22	23	25	10	1							Meltdrs Scrubber	11
23	23	23	10	1							Rodross Baghouse	11
24			20	1	2.07	18.3	0.76	472	0.94		-	11
-	23	20.		1	2.07	10.3	0.76	4/2	0.94		AlWedge Scrubber	12
25	24	21	10	1							Residue Dryer	
26	25	28	5	1			0.77	2.0	• ••	•	Scrap Furnace	12
2 7	25	26	5	1	17.90	6.1	0.46	30	1.32	0	#3Helt Scrubber	12
28	25	27	5	1							#2Melt Scrubber	12
29	26	24	5	1							ZN Pure Baghouse	12
30	31	30	0	0	20.60	13.4	0.91	347	14.17		AMP Reactor	13
31	32	29	0	0							AMP Dryer	13
32	33	31	0	0							Doyle Reactor	13

^{*} variables defined on page 26

NAME--emission point name

PSNUM--point source identification number

With the exception of the last variable, these data were gathered from previous reports of PEDCo (1975), Valentine and Fisher (1975), PES (1978), and

EPA (1975a,c). The last variable, PSNUM, is a categorization variable
developed for the purposes of this study. It is explained in a later section
of the report. Point sources are located on the Map PTSOURCE in Appendix A.

4.3 PROCESS FUGITIVE SOURCES

These emissions are associated with the metals industry. These are pollutants that escape to the atmosphere from industrial processes. They may be leaks, vents, and overflows from production and pollution control equipment or buildings. They may escape from uncontrolled portions of processes exposed to the atmosphere, such as conveyor belts or by-product dumps. Many of these fugitive sources are attendant to the processes and exhibit regularity in their location and strength (e.g., building fans). Others, particularly leaks and overflows associated with process upset conditions and malfunctions, are erratic in both frequency and magnitude.

Process fugitive particulate sources were identified and mean emission rates were estimated in previous studies (PEDCo, 1975c; Valentine and Fisher, 1975; PES, 1978; PEDCo, 1979). Those sources have been identified in Table 4.2, together with percentage lead and cadmium components measured in the same surveys.

All industrial process-related lead sources were combined in categories for later analyses. Those sources and mean emission rate estimates and rankings can be found in Table 4.3. This table represents the best

Table 4.2 Process Fugitive Sources Ordered by Lead Source Strength	Table 4.2	Process	Fugitive	Sources	Ordered by	Lead	Source	Strength
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Name	EMR	PBFR	CDFR	PBEMR	CDEMR
Blast furnace	30.00	61	9	18.3000	2.7000
CPP exhaust fams	34.00	31	1	10.5400	0.3400
CRE con exhaust fans	25.00	34	1	8.5000	0.2500
Cast roof fans	12.40	31	1	3.8440	0.1240
PB ref roof vent	6.70	37	1	2.4790	0.0670
Elec fur roof	4.30	10	0	0.4300	0.0000
Sinter prod dump	0.54	31	1	0.1674	0.0054

EMR--total particulate emission rate, lb/hr; PBFR--percentage lead in emission; CDFR--percentage cadmium in emission; PBEMR--lead emission rate, lb/hr; CDEMR--cadmium emission rate, lb/hr.

estimates for comparing industrial sources that could be developed with the available data. It was developed from information collected between late 1974 and early 1979 and some updating may be required.

4.4 ACTIVE FUGITIVE SOURCES

These sources are varying and intermittent pollutant sources whose suspension energy is provided by agents other than steady state industrial processes. They may be industrial sources related to activities such as stockpiling, truck and train loading and unloading, or materials handling. They may be related to land use such as ground working, construction, or surface mining. Or they may be related to transportation sources such as reentrainment by vehicular traffic, from open carriers, or mobile source

Table 4.3 Point Source Category Components and Lead Emission Rates (EMR)

			9 / - 5	EMB
NEDS	Name	Mean Lead [.] EMR lb./hr.		
2	Crushing Plant Dryer	3.4	6	6
3	Crushing Plant Collector	. 4	1	12
4	Crushing Plant Rodmill	. 4	1	13
5	Crushing Plant Baghouse	. 4	1	14
6	Oreprep Baghouse	. 3	1	16
8	Lurgi D Scrubber	1.4	2	10
9	Lurgi N Rotoclon	16.0	28	1
10	Lurgi B Scrubber	4.1	7	4 5
11	Lurgi A Scrubber	3.4	6 5	2 7
12	Lurgi C Scrubber	2.7	3	9
17	Zinc Fume Granulator	1.8	3 0	9
32	PB Refinery Scrubber		0	
15 FUG	Electric Furnace Scrubber	10.5	18	2
	Oreprep Exhaust Fans Ore Conc Exhaust Fans	8.5	15	3
	PB Refinery Roof Vent	2.5	4	8
	BF Roof Vents	.9	2	11
	Electric Furnace Roof	. 4	1	15
	Sinter Product Dump	. 2	i	17
	·		•	• •
	Total Low Smelter PB EMR	57.3		
	II. Smelter Mid-L	evel Sources		
NEDS	Name	Mean Lead	% of	EMR
		EMR lb./hr.	Category	Rank
7	Pellet Dryer	8.0	2 4	2
16	Zinc Fume Main Stack	2.3	7	4
FUG	Blast Furnace	18.3	54	1
FUG	Casting Roof Fans	3.8	11	3
FUG	Fuming Furnace Roof	1.3	4	5
	Total Mid Smelter PB EMR	33.7		

Table 4.3 Continued

	III. Smelter High	never sources		
NEDS 		Mean Lead EMR 1b./hr	% of Category	EMR Rank
1	Smelter Main Stack	43.0	100	1
	IV. Zinc Plant Low	-Level Sources		
NEDS	Name	Mean Lead EMR 1b./hr.	% of Category	EMR Rank
33	Concentrate Dryer	.02	1	5
19	Concentrate Silo	0	0	
22	Rosconv Scrubber	0	0	
25	Melt DRS Scrubber	.07	3 1	4 6
23 20	Rodross Baghouse	1.36	6 7	1
21	#1 Wedge Scrubber Residue Dryer	0	1	7
28	Scrap Furnace	.07	3	4
	#3 Melt.	.37	18	2
	#2 Melt.	.11	. 5	3
24		0	. 0	
	_		•	
	Total ZN Plant Low Level	2.02		
	V. Zinc Plant High	Level Sources		
NEDS		Mean Lead EMR 1b./hr.		
18	Zinc Plant Main Stack	2.28	100	1
	VI. Ammonium-Phospha	te Plant Sourc	es	
NEDS	Name	Mean Lead	% of	EMR
		EMR 1b./hr.		
	AMP Reactor	0		
32	AMP Dryer Doyle Reactor	0 0		
33				

combustion emissions. They are distinguished from the previous category in that their frequency and magnitude are not attendant to industrial processes in the steady state time frame, and from the following category in that their suspension energy is independent of kinetic meteorological factors.

Active Sources (Industrial)—Particulate fugitive sources were identified in the same surveys cited in the process fugitive discussion. Similarly, mean emission rates and chemical constituencies were estimated in those studies. These sources are identified in Table 4.4. They can be located on Map ACTIVES in Appendix A.

Table 4.4 Active Fugitive Source Characteristics

Name	MapID	EMR	PBFR	CDFR	PBEMR	CDEMR
Silica slag pile	73	17.8	5	0	0.890	0.000
Sinter storagè	45	1.6	42	1	0.672	0.016
Slag storage	33	20.2	2	0	0.404	0.000
State highway piles	s 74	30.3	1	0	0.303	0.000
Sinter-storage	49	0.4	42	1	0.168	0.004
Coke storage	45	0.6	0	0	0.000	0.000

EMR--total particulate emission rate, lb/hr; PBFR--percentage lead in emission; CDFR--percentage cadmium in emission; PBEMR--lead emission rate, lb/hr; CDEMR--cadmium emission rate, lb/hr.

Active Sources (Transportation) -- Several researchers have investigated those factors most contributory to road dust impacts. The condition of the road surface seems most important. Unpaved roads result in orders of magnitude greater particle suspension than paved roads. In fact, some studies have ignored the effect of particle resuspension from paved roads altogether (PES, 1978). However, others have found considerable resuspension of tracer materials from paved roads (Sehmel, 1973). The PEDCo (1975a) study recognized three categories of roads: unpaved, paved, and dusty paved. In general, it seems that in comparison to unpaved roads, paved road contributions to total suspended particulate may be negligible. However, with respect to hazardous materials or trace elements deposited on roadways, traffic induced resuspension may be significant (Sehmel, 1973). Suspension from unpaved roads has been related to particle size in the roadbed (Becker and Takle, 1979). PEDCo (1975a) has developed formulas for total particulate suspension from unpaved roadways based on traffic characteristics and the silt content of the roadbed.

Various studies identify differing ambient impacts associated with vehicle weight and speed and the number of vehicle passes (Becker and Takle, 1979; PEDCo, 1976; Sehmel, 1973; Smith, 1976). Vehicle characteristics affect both suspension rates and initial dilution through mechanically induced turbulence. The latter results in a shallower dilution gradient away from highways and a lesser dependence on atmospheric stability and source height (Becker and Takle, 1979; Sistla et al., 1979).

Other factors found to be important in road dust impacts are the angle between the wind and road (Calder, 1973; Sistla et al., 1979) and the surface roughness of the roadside environment. Little and Wiffen (1979),

Smith (1971), and Heichel and Hankin (1976) found the characteristics of roadside flora significantly affected ambient lead concentrations. Becker and Takle (1979) suggest that sampling height is an important consideration due to the differing particle sizes and settling rates of the suspended material.

Each of the roads identified in this study was characterized as paved, unpaved, or dusty paved. Annual vehicle miles traveled (VMT) for each road were obtained from previous reports and traffic studies (PEDCo, 1975c; PES, 1978; State of Idaho, 1974a). Emission Rate Factors developed by the EPA as described by PEDCo (1975c) were then applied to get per unit distance emission rates for these roads. Both road dust and gasoline combustion factors were included.

Active Sources (Urban)--Urban active fugitive source strengths were considered to be proportional to the traffic volume on city streets. The total VMT estimates for each city were obtained and proportionately allocated over the area of the community. No absolute estimates of emission rates were determined for the last two sub-categories as they were treated proportionately in later regression analysis correlating these estimates to observed concentrations. The details may be found in the parent document. However, these sources are located on the Maps TOWNS and ROADS in Appendix A. Specific emissions estimates could be developed for these sources by using the final regression coefficients in a calibration procedure. However, this would require additional work.

4.5 PASSIVE FUGITIVE SOURCES

These sources are reentrained by the wind. They include open areas of bare soil and exposed industrial areas. Their magnitude depends on wind speed and direction and surface conditions. The air quality aspects of these sources have attracted significant attention in recent years due to the need for predicting urban air pollution levels, and in the nuclear industry where resuspension of spilled nuclear materials is a possible health hazard. Like roadway sources, passive source strengths are greatly dependent on surface conditions. Soil particle size distribution, surface roughness, orientation, cover, and moisture conditions are the principal surface variables. Those factors that contribute most to wind suspension have been investigated for many years in relation to soil erosion. Those studies have been modified to develop air pollution impact estimates (PEDCo, 1973; Wilson, 1975).

The basic technique utilized in assessing passive source impacts was developed in a series of publications by Chepil and associates and was reviewed by Woodruff and Siddoway (1965). In that publication, they modified the techniques to develop "A Wind Erosion Equation." Those definitions and techniques have been extended for use as air quality predictors commonly termed the "Modified Wind Erosion Equation Method." The Modified Wind Erosion Equation Equation is as follows:

$$E = I + K' + C' + L' + V'$$

where

E = total mass soil movement, tons/acre/yr

I = soil erodibility (or the potential to erode), tons/yr

K' = soil ridge roughness factor

C' = climatic factor (soil moisture and wind velocity)

L' = field length factor

V' = vegetation cover factor

The strategy of this equation is that the erodibility. I, is defined as the potential loss of soil in tons/acre per annum from a wide, unsheltered, isolated field with a bare, smooth surface. It is based on both wind tunnel and field observations (Woodruff and Siddoway, 1965). The others are mitigative variables varying in value from 0 to 1.0 that reduce the final erosion estimate.

This equation has been used to estimate emission rates from area sources by multiplying E by the area of the source in acres times a conversion factor. That conversion factor represents that fraction of the total soil movement (E) that is observed as ambient particulate. Literature values for the conversion factor are found from 0.3 to 10% (Wilson, 1975).

The physics of erosion involves three transport processes: saltation, surface creep, and suspension. Saltation is the movement of soil by a short series of bounces usually rising no more than a foot above the ground. Surface creep is the rolling or sliding of the larger particles along the surface. Suspension is the entrainment of fine particles up away from the surface (Buckman and Brady. 1972). A significant gradient of particle size with height results from these processes. Only the suspension process produces long range air pollutants.

Erosion rates are also sensitive to meteorological variables.

Total soil movement varies with the wind speed cubed and inversely with soil moisture (Woodruff and Siddoway, 1965). Rain and snow cover eliminate

wind erosion when in sufficient quantity. In this area, diffusion of suspended materials by wind erosion is practically independent of atmospheric stability. Significant wind speeds are required to initiate suspension. In the Silver Valley such wind speeds are observed only in near neutral conditions and only in the up or down valley direction (PES, 1978). It seems that significant reentrainment from these sources occurs in the Silver Valley under meteorologically limited circumstances.

All of the available data for passive sources have been accumulated from three studies conducted in the Silver Valley (PEDCo, 1973, PES, 1978; State of Idaho, 1978). The sources can be found in Table 4.5 and located on Maps found in Appendix A, by using the value VAL as described in Part E of that Appendix. No absolute emission rate estimates were prepared for these sources because of the nature of the later analyses as detailed in the parent document. However, they are listed in their relative order of lead source strength in Appendix C. Estimates could be accomplished by a calibration procedure utilizing the final regression coefficients. Such an analysis would be interesting but is beyond the resources of the report.

Table 4.5 Passive Fugitive Source Characteristics

	VAI ER(re L 1s DDE, ROUGH	ferenced a number LENGTH, t the Win	de numbers in the tex used to k VEG ndblown Dus	t. ey the ma are the t Equatio	p input in source dev in as defin	nformation relopment ned by Che	factors	fi ARNOTE, PB da PE au	nes fraction NOTE, CDNOTE ta for the vi S report, P (of the soi The nariables at to the Pedo calth and h	its samples note denote: nove. The no reports, felfare Dep imilar sam	
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5.0 AMBIENT ANALYSIS

5.1 GENERAL

The heavy-metal contamination problem in the Silver Valley is extremely complex. The air pollution component is especially difficult. Considerable care must be exercised in any analysis of the support data and the particulars of the analytical procedures. In any type of modeling study where mathematical expressions are used to represent physical phenomena, certain assumptions have to be made to accommodate the analysis. The adequacy of those assumptions most often determines the quality of the results and conclusions. The types of assumptions that are inherent to simulation diffusion models (at the practiced state-of-art) could possibly make such analyses unreliable for the Silver Valley situation. However, given the wealth of data available, an empirical application in this situation could result in a more reliable analysis. Literature citations and a support argument for this position can be found in von Lindern (1980b).

This is not to say, however, that there are no problems in an empirical analysis. Analyses where observed pollutant concentrations are related to atmospheric and emissions indices, in the ignorance of the physical phenomena involved, are particularly prone to erroneous conclusions. Great care must be exercised in the design of the model and the interpretation of results. In any modeling analysis, it is important to understand the physical and anthropogenic factors involved. This background information is

necessary to select and to evaluate the assumptions discussed above. Most of this background material has been summarized in earlier sections. However, there are three areas of specific difficulty that should be discussed. The problems are presented in detail in von Lindern (1980b). However, they are important enough to repeat briefly in this presentation. They are:

- 1. Meteorological factors associated with complex terrain.
- 2. Spatial and temporal valuation in source strength and multiple source configurations.
- 3. Interdependency of source strength and configuration with meteorological variables.

Two atmospheric phenomena common to complex terrain literature are especially important in the Silver Valley: the frequent formation of surface based nocturnal inversions and the mountain-valley drainage wind. Radiative cooling of the slopes of the valley causes air temperature to decrease near the surface. As it cools, it becomes more dense and flows downslope to the valley floor and subsequently down valley. Extreme diurnal shear and low level isothermal structures can result from this drainage. This capping phenomena inhibits pollutant diffusion and enhances terrain channelling. After sunrise the slopes of the valley heat more rapidly than the floor. This results in the descent of the inversion layer as insolation proceeds and fumigation phenomena can return residual pollutants trapped aloft overnight to the ground. Following this period flow up the valley predominates. This flow is typically associated with the prevailing synoptic winds augmented by upslope winds resulting from differential heating. The valley narrows and deepens considerably in this direction. As a result, significant terrain channelling is expected with up-valley winds, even in the absence of stable layers aloft.

Conventional modeling analysis dealing with long-term effects requires some degree of uniformity and homogeneity in both atmospheric and source behavior. The second difficulty in analyzing the Silver Valley via modeling concerns the spatial and temporal irregularities in source behavior. Source descriptions were provided in the last section. Point sources are generally uniform in their behavior. Industrial processes are designed to operate at particular rates and capacities. Control equipment is correspondingly designed and emission rates are consistent. When dealing with process fugitive sources, emission rates become irregular, depending on process fluctuations and outside stimuli. These two categories are also subject to upsets and malfunctions that can result in orders of magnitude changes in emission rates for short periods of time. Active fugitive sources are sporadic and their emission rates depend on the frequency of their parent mechanical activity and local meteorological conditions. Passive fugitive source emission rates are wholly dependent on meteorological factors and vary not only in magnitude and frequency but in configuration as well. Wind speed, wind direction, and surface conditions dictate these sources' contribution to atmospheric levels.

This last factor, the interdependency of source strength and configuration with meteorological variables is, perhaps, the most confounding factor in applying modeling techniques to this problem. Ironically, it is this same dependence that allowed solution of this problem through the techniques employed. As an excellent meteorological data base was available through the Bunker Hill Company's Supplementary Control System (SCS) monitoring network, source behavior could be quantified through meteorological indices. However, doing so in an empirical format utilizing over thirty variables

and two years of data from hundreds of sources was a tremendous computational demand. In this study, the problem was addressed by using the geographic information system described in von Lindern (1980b).

5.2 THE MODELING PROCEDURES AND ASSUMPTIONS

5.2.1 The Basic Source-Receptor Model. The objective of this modeling analysis is to quantify those factors most important in air lead contamination in the Silver Valley and to ascertain the relative contribution of different sources to excess atmospheric concentrations. Considering the wealth of data available and the particular difficulties of applying standard diffusion models, an empirical approach was selected. The first step in the solution strategy was to develop the basic source-receptor model describing the fundamental relationship between a receptor and any source of exposure. The design of that model should account for those particular topographical emissions, and meteorological phenomena suspected of confounding air pollution impact analysis in this area. The most important of these considerations have just been discussed. Cursory examination of the wind spectra in the Silver Valley suggests that the mountain-valley drainage/nocturnal inversion scenario dominates the local meteorology on the majority of days. The model developed sought to quantify the basic source-receptor relationship in a quasi-Gaussian fashion by accounting for the peculiar flows associated with this diurnal phenomenon. It is also known that certain meteorological conditions can void or considerably modify the basic relationship. The Bunker Hill Company SCS has identified those critical meteorological situations. Their data are used to modify the basic relationship where appropriate by dummy variable additions to the regression procedure.

5.2.2. <u>Variable Construction</u>. A derivation of the modeling equations and construction of the variables and regression analyses are not contained in this report but can be found in von Lindern (1980b). It is most important in this report to point out and discuss the assumptions inherent to the modeling strategy and how those assumptions affect the quality of the results and limit the conclusions that may be drawn from these techniques.

The basic modeling equation was the simple Gaussian plume model

$$x = Q \cdot \frac{1}{2\pi u} \cdot \frac{1}{\sigma y \sigma z} \cdot \exp(\frac{-y^2}{2\sigma y^2}) \cdot \exp(\frac{-H^2}{2\sigma z^2})$$

where

x = the pollutant concentration at the receptor

Q = the initial pollutant source strength

u = the wind speed in the x direction

σy and σz = the standard deviations in pollutant concentration in the y and z directions

x,y,z = the Cartesian coordinates

H = the difference in elevation between source and receptor

No provision is made for differential plume rise or particulate settling.

This model was derived in a surrogate form to accommodate those variables that were available from the emissions, meteorological, and geographic data bases. Three basic requirements were involved in the surrogate derivations:

(1) the equation had to be derived using variables that were both available from the emissions and meteorological inventories and that were amenable to

manipulation in the geographic information system, (2) the model had to be capable of empirical parameterization via linear regression, and (3) the model, to the maximum extent possible, should reflect the concepts of accepted air pollution meteorology. Simultaneously accommodating these constraints involved making several assumptions and adjustments to the solution strategy. Those basic assumptions are discussed below.

OBSERVATION TIME PERIOD BASIS. The dependent variable in the equation is the mean 24-hour pollutant concentration observed at the various stations. These data were obtained from the State's hi-vol network shown on page 20. These data are believed to be of excellent quality and were used as is. However, this daily periodicity defines the base time unit for the dependent variables.

ADEQUACY OF EMISSION DATA. This represents what could be called the weakest data base in the procedure. Only single observations were available for many of the sources. Quarterly, self-reported, emission rates were available for some smelter sources, but the reliability of these data is unknown. These estimates were reduced to 24-hour emission rates. It is suspected that many of these sources can vary more than an order of magnitude on a daily basis, especially when malfunctions and upsets are considered. Two elements of strategy were developed to deal with this deficiency. The first mitigative factor was to introduce on-off criteria as described in the parent document. Many sources were known to be inoperative on particular days for either operations or meteorological reasons. Zero emission rates were assigned to the appropriate sources for these days. The second mitigative element

concerns the basic philosophy of the entire project. The source contributions and rollbacks eventually developed are discussed in relative rather than absolute terms. Pseudo-dummy variable analysis, as described by Draper and Smith, is utilized throughout. This technique is especially useful in determining the relative significance of different variables in regression analysis.

Source estimates were eventually grouped into eight categories for final analysis. The grouping was based on the spatial configuration of the sources. The basic assumption is that significant particulate lead sources emanate from eight predominant locations in the valley. By appropriately characterizing the emissions (by rate estimates and on-off criteria) and the atmosphere (by wind and stability indices), relative contributions could be ascertained by regression analysis.

THE VIEWFIELD ASSUMPTION. Downwind concentration dilution in the traditional Gaussian formula is inversely proportional to the wind speed. Doubling the wind speed doubles the space between particles and halves observed concentrations. However, this assumes a constant wind direction and speed. In the Silver Valley wind direction and speed change continuously and usually reverse themselves daily in association with the mountain-valley drainage phenomenon. In order to compensate for this difficulty in the daily time basis, two wind flow directions were assumed in conjunction with the mountain-valley drainage phenomenon. Up-valley and down-valley winds were defined. The mean daily wind speed, direction, duration, and standard deviation in wind direction were calculated for each flow. Each receptor's face was turned into the wind and its upstream "view" was considered. This is the

VIEWFIELD assumption. How many sources a receptor can see depends on its "viewfield" or a narrow upwind sector that contains those sources that could possibly impact that receptor. The depth of view depends on wind speed and duration of the wind flow from that direction. The width of the sector depends on the variation in wind direction. An expression was derived to express the number of hours per day that a receptor was exposed to a source and the assumption was made that the downwind component of dilution was proportional to that duration of exposure and inversely proportional to wind speed and the width of the view-sector. These terms are all calculable in the geographic information system and could be easily accomplished for each receptor location. In this way, both a proportioning factor and an additional on-off criteria were developed. If, because of wind speed or direction, a receptor is not exposed to a source during the day the exposure reduces to zero. If the wind blows consistently all day from source to receptor the downwind dilution becomes the 1/u term in the traditional equation. This variable is not meant to simulate wind flow in the valley. It is an index meant to represent downwind dilution from sources in the upand down-valley directions and is designed for use in a regression equation.

THE STABILITY ASSUMPTION. The surrogate term for the $\frac{1}{\sigma_y \sigma_z}$ term in the Gaussian model equation was designed so that the Gifford-Pasquill form of these variables could be recovered from the eventual regression coefficients. First the traditional form of the Gifford-Pasquill plots were derived in a system of linear equations developed through logarithmic transformations. It was then assumed that the linear coefficients for this system were a function of atmospheric stability and would be specific for this situation.

Next the potential temperature gradient at Spokane International Airport was selected as the daily indicator of atmospheric stability. This has been found, through the Company's SCS, to be the most appropriate stability index available. It was then assumed that the linear coefficients were first degree functions of the potential temperature gradient. This assumption is consistent with the form of the Gifford-Pasquill charts and results in a convenient form of the expression for both parameterization and geographic manipulation.

The term $\exp(\frac{-y^2}{2\sigma y^2})$ in the Gaussian plume THE PLUME MEANDER ASSUMPTION. analogy represents the dilution effect of being remote from the mean wind vector. In the Gaussian formulation, a uniform wind direction is assumed and the degree of lateral dilution is a function of atmospheric stability. In the Silver Valley, wind direction fluctuates continuously and straight line flow is not expected. Thus, it was assumed that exposure reductions associated with being remote from the mean wind vector are more a function of variation in the mean wind vector than of the stability criteria per se. Two lines of reasoning justify this assumption. The variation in the mean wind vector (plume meander) is likely a function of stability itself. And the maintenance of a crosswind Gaussian distribution in the Silver Valley is unlikely. Plume meander, when considered on a daily basis, would predominate any such distribution. As a result the standard deviation in wind direction was selected as a measure of crosswind remoteness and dilution was assumed to be a function of the number of standard deviations a receptor was from the mean wind vector.

THE VERTICAL DISTRIBUTION ASSUMPTION. Differential plume rise, particulate settling, and plume depletion are all ignored in this model. A constant plume rise and constant plume elevation are assumed. The main reason for ignoring differential plume rise was the insufficiency of the data base for developing appropriate variables. In order to account for plume rise, an equation utilizing the atmospheric surface temperature and wind velocity would have to be included in the source height factor. These measurements are used in other independent variables in the modeling equation. Including them in another term would increase the possibility of undesirable effects associated with inter-variable correlation. Moreover, from a practical point of view, these are low temperature plumes and only a single morning surface temperature was available. The benefit of using that single observation in predicting an effective daily plume rise was considered not worth the confounding effects of adding a possibly redundant variable. Considering plume depletion, no data were available for developing settling estimates. Making unnecessary assumptions in the development of regression variables should be avoided. Practically, plume depletion must be significant with respect to certain sources. Fugitive sources in particular contain large particle emissions subject to considerable fallout. Point sources, on the other hand, are fine particle emissions that may approximate gaseous behavior. explained later, cadmium emissions are used to parameterize the basic source receptor model. Cadmium emissions are predominantly fine particulate. noring settling phenomena is, likely, permissible in defining parameters for the model. Later, when developing source estimates for lead and particulates, depletion from fugitive sources must be considered important. However, in the division of source categories, active and passive sources appear in

separate independent variables. When linear regression analyses are applied to those variables it can be inherently assumed that some constant amount of plume depletion is accounted for in assigning the regression coefficients. In practical terms, the assumption is that a particular percentage of fugitive emissions fall out between the source and nearest receptor and that gaseous behavior is observed beyond. This assumption is adequate for the empirical format of the study.

THE EMPIRICAL MODIFIERS. The Bunker Hill Company meteorologists operating the Supplementary Control System (SCS) have long recognized those difficult meteorological factors discussed earlier in this section. They have, through years of experience, developed semi-quantitative indices to represent the onset of certain meteorological and operational conditions. For the most part, these variables identify the onset of non-routine conditions where "normal" assumptions do not apply. As such they are quite appropriate for use in regression analysis as dummy variables to account for the effects of special conditions where the base model may not apply. These variables were transformed to act as empirical modifiers in the regression analyses. Details of the variables and transformations can be found in the parent document.

5.2.3 <u>Finding the Model Parameters</u>. The model was assigned parameters through a stepwise regression procedure that forced inclusion of the variables developed as the initial surrogate model. The several empirical modifiers developed from the SCS were also offered for forward selection. There are several assumptions inherent to regression analysis that are not

necessary to include. However, there are two assumptions made in solving the regression equation that are important to discuss here.

The first is the use of cadmium data to find the model parameters. The number of sources and the difficulties with defining configurations make it impossible to solve the modeling equation for lead or total particulates. Cadmuim emisions, however, seem to predominatly arise from within the lead smelter or, more to the point, from the same geographic location. This considerably reduces the complexity of the modeling equation for cadmuim. Because cadmuim emissions do emanate from three distinct source heights, one unknown function still precludes solution of the equation. An assumptive constraint has to be added to the regression matrix. That constraint was developed by assuming that the ratio between $\alpha_{\rm y}$ and $\alpha_{\rm z}$ at neutral stabilities at 0.1 miles from the source will be the same as the ratio found under there conditions in the traditional Gifford-Pasquill charts.

5.3 THE IMPORTANT SYSTEM VARIABLES

Using these two assumptions, the equation was solved and the selected model is shown below. An extensive discussion of the results is found in von Lindern(1980b).

The regression statistics for this model indicate that pollutant dispersion can be successfully quantified by this model form. Seventy-three percent of the variablity in observed concentrations is explained at strong significance levels. The initial model seems particularly strong (R2 = .71 at p .0001). This is especially encouraging considering the difficulties with cadmuim source estimates discussed earlier.

Table 5.1 Regression Statistics for Selected Stepwise Model

Initial Hodel: DEPZ = β_0 + β_2 LNVWF + β_3 POTEMP + β_4 lnX + β_5 POTEMInX + β_6 NSDSQ

Step	Variable Added	Sum of Squares Hodel	F-Hodel (P>F)	F-Variable on Entry (P>F)	R ² Hodel
0	(Initial Hodel)	7749.4	673.3(.0001)	(#11 .0001)	.712
1	HD 5 O L N X	7836.9	583.3(.0001)	39.0(.0001)	.720
2	REGVAR	7870.1	507.3(.0001)	15.0(.0001)	.724
3	BFDOWN	7892.5	448.2(.0001)	10.2(.0015)	.726
4	HD36	7914.2	402.1(.0001)	10.0(.0016)	.728
5	W15	7930.8	364.4(.0001)	7.6(.0059)	.729

Total SS = 10877.5

Parameter Statistics on Final Sta	Parameter	. Statisti	ca on	Final	Ste
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Source	Parameter Estimate	Sum of Squares	F (P>4)	
INTERCEPT	2.31	(SAS Type II)		
LNVWP	.17	31.0	14.2(.0002)	
POTEMP	. 65	1511.5	694.5(.0001)	
LHX	1.51	1999.4	918.7(.0001)	
POTEMP*LNX	.15	898.2	412.7(.0001)	
NSDSQ	04	91.8	42.2(.0001)	
HD50LHX	.16	33.0	15.2(.0001)	
REGVAR	.20	33.2	15.3(.0001)	
BFDOWN	. 75	40.9	18.8(.0001)	
HD36	2.06	25.4	11.7(.0006)	
W15	.084	16.5	7.6(.006)	
SS (Hodel)		7930.8	F-Model = 36.4	
SS (Error)		2446.7	R^2 Model = .729	
SS (Total)		10877.5		

^{*} variable descriptions can be found in Table 5.3

Table 5.2 Regression Statistics for Final Logarithmic Model

DEPZ =
$$\beta_0$$
 + β_2 LNVWF + β_3 POTEMP + β_4 LNX
+ β_5 POTEMP*LNX + β_6 NSDSQ + β_7 BFDOWN
+ β_8 MD36 + β_9 MD50LNX + β_{10} W15 + β_{11} REGVAR

Source	DF	Sum of Squares (SAS Type IV)	F-Value (PR>F)
LNVWF	1	40.4	22.0(.0001)
POTEMP	1	1327.9	720.8(.0001)
LNX	1	2016.8	1094.6(.0001)
POTEMP*LNX	1	784.4	425.7(.0001)
NSDSQ	1	80.7	43.8(.0001)
BFDOWN	1	23.4	13.0(.0003)
MD 36	1	18.4	10.0(.0016)
MD50LNX	1	22.2	12.0(.0005)
W15	1	17.4	9.4(.0022)
REGVAR	1	36.4	19.8(.0001)
Model	10	7499.4	407.0(.0001)
Error	1428	2631.0	•
Total	1438	10130.4	$R^2 = .740$
Parameter	Estimate	T-Value (PR>T)	Std. Error
R -	2.26	6.50(.0001)	.347
β ₀	.192	4.69(.0001)	.040
B2	612	-6.62(.0001)	.023
β2 β3 β4	-1.48	-26.85(.0001)	.045
β ₅	.138	-33.08(.0001)	.007
β ₆	039	-6.62(.0001)	.006
β ₇	569	-3.61(.0003)	.157
β8	1.75	3.16(.0016)	.554
β ₉	124	-3.47(.0005)	.036
β ₁₀	.085	3.07(.0022)	.027
β11	.205	4.45(.0001)	.046

^{*} variable descriptions can be found in Table 5.3

As in von Lindern (1980b) the best way to discuss these model results is in terms of the regression coefficients. Parts of those discussions are included here.

Ten variables were selected as important in predicting observed cadmium levels. The first five variables comprise the initial surrogate model. Five of the empirical modifiers offered were found to be significant. Briefly, they are:

BFDOWN--the on-off indicator of blast furnace operation.

MD36--an on-off indicator of the most severe limitation in mixing depth.

MD50LNX--the on-off variable for situation of uninhibited mixing depth times the logarithm of distance.

W15--an indicator of suppressed wind speeds in the middle atmospheric layers of the valley.

REGVAR--the severity code indicating adverse dispersal conditions associated with peculiar synoptic situations.

The model as taken from the parent document is as follows:

$$\ln(x_T) = \beta_0 + \beta_1 \ln QHF + \beta_2 \ln VWF + \beta_3 NSDSQ + \beta_4 POTEMP$$

$$+ \beta_5 \ln X + \beta_6 POTEMP * \ln X + \beta_7 BFDOWN + \beta_8 MD36$$

$$+ \beta_9 MD50LNX + \beta_{10} W15 + \beta_{11} REGVAR$$

The parameters, their associated independent variable, and the parameter values are shown in Table 5.3.

Table 5.3 Parameter Values for the Logarithmic Model

Parameter	Independent Variable	Factor Par Description	ameter Value
β _O	Intercept	Initial dispersion	2.26
β	ln (QHF)	Source height function	1.00
β ₂	ln(VWF)	Receptor View function	.192
β ₃	NSDSQ	Lateral position factor	039
β4	POTEMP	(Stability)	612
β ₅	LNX	(and)	-1.48
β ₆	POTEMP*LNX	(distance factors)	.138
β ₇	BFDOWN	Operations factor	569
β ₈	MD 36	Limited mixing depth fact	or 1.75
β ₉ ·	MD50LNX	Uninhibited mixing depth factor	124
^β 10	W15	Inhibited mid-level winds factor	.085
β ₁₁ -	REGVAR	Severe synoptic factor	.204

^{*} for a detailed description and derivation of these variables please see von Lindern 1980b

These parameters and associated variables can be grouped for discussion relative to their contribution to quantifying pollutant dispersion in this valley.

- β_0 , β_4 , β_5 , and β_6 are the dispersion parameters for the plume centerline dilution effect associated with the mean wind as derived from the traditional Gaussian form. These parameters were used to derive the familiar σ_y - σ_z plots of Gifford (1961) and direct comparisons of the dispersal conditions in this situation are made relative to the standard modeling assumptions.
- β_1 is the unit coefficient for lnQHF or the source strength-source height term. This term reflects the initial source strength reduced by a factor dependent on the relative source-receptor height. The latter is developed from the same basic component parameters as β_0 , β_4 , β_5 , and β_6 above.
- β₂ and β₃ are associated with the terms ln(VWF)and NSDSQ. These two variables, as they were developed, serve to mitigate the standard model predictions with respect to the topographically induced wind conditions. VWF is an exposure factor that accounts for the reduced receptor "view" of the source associated with up- and down-valley wind shifts. NSDSQ is associated with the lateral variance in the mean wind and accommodates reduced exposures associated with the cross-valley wind shifts. In the form offered in the modeling analysis, they become empirical modifiers of the more traditional dispersion equation characterized by the above parameters and variables.

- β_7 is the parameter for BFDOWN and is a direct empirical modifier associated with shutdown of the largest single cadmium source. The β estimate for this term is (-.569). When the blast furnace is down (BFDOWN = 1), the predicted effect is $\exp(-.569 * 1) = .57$ times the model prediction. This suggests that when the blast furnace is nonoperative at least 16 hours per day, ambient cadmium levels are reduced 43%.
- β_{8} and β_{q} are associated with extreme mixing depths. β_{q} is the parameter for MD50LNX. This variable allows for greater dispersion under uninhibited vertical dispersion conditions. Because a limited mixing depth associated with nocturnal inversion is the "normal" situation in this valley, the standard diffusion parameters (β_0 , β_4 , β_5 , β_6) are calculated under that circumstance. The value of β_9 is 0.125. The significance of this variable is as follows: when mixing depth is great (i.e., MD50 = 1, MD50LNX = ln(x)) the value of the coefficient of ln(x) or p + q = -(1.48 + .125) = -1.61. This is nearly the value supposed in the traditional Gifford-Pasquill form as discussed in von Lindern (1980b). This supports the idea that the "normal" situation in the Silver Valley has an associated limit to vertical dispersion probably related to the surface based nocturnal inversions. Uninhibited vertical dispersion is an "abnormal" situa-Similarly, when mixing depth is severely inhibited, an opposite "abnormal" effect is present. The variable MD36 has a value of 1 when mixing depth is most shallow and 0 at other times. The $\boldsymbol{\beta}$ value for this variable is 1.75. This suggests that when the lowest level inversion structure exists, the model predictions are increased

by exp(1.75) = 5.75 times. This represents a severe condition treated here by a simple empirical modifier.

 β_{10} and β_{11} are empirical modifiers associated with special synoptic situations. β_{10} -W15 accounts for reduced wind speeds in the midlevel valley atmosphere. β_{11} is associated with REGVAR, an indicator of severe synoptic conditions. The W15 variable has greatest effect when the mean wind value is less than 1 mph. At that value the model estimate may be increased as much as $\exp(.08*5) = 1.5$ times. This situation implies extreme calm or shear in the middle atmospheric levels. REGVAR is a severity code associated with some peculiar synoptic conditions. Two conditions are especially important. They are the valley drainage wind (=3) and stagnation (=5) that are both associated with high pressure areas in the mountain range vicinity. The former can increase model estimates by $\exp(3*.20) = 1.8$ times and the latter by $\exp(5*.20) = 2.7$ times.

The strength of the initial model indicates that the mountain valley drainage phenomena dominate pollutant dispersion in the Silver Valley. The two mixing depth variables selected in the stepwise process suggest that nocturnal inversions are also part of the "normal" dispersion picture for the valley. Four levels of mixing depth were offered in the stepwise procedure. The non-significance of the two middle levels indicates that they are accounted for in the remainder of the model. In practical terms this means that the basic source-receptor model reflects a diurnal capping inversion between 3600 and 4800 feet. Special modifiers to the basic model are required only when greater or lesser mixing depths are present.

It also means that the dispersion parameters derived from the regression coefficients for this situation reflect this diurnal phenomenon. Some important aspects of the montane air pollution meteorology for this area can be explained by comparing these derived dispersion coefficient estimates with the standard Gifford-Pasquill parameters.

Figures 5.1 a and b show the derived dispersion parameters for this situation plotted as solid lines. The dotted lines are the corresponding plots taken from the subroutine distributed by EPA (1976b) to estimate Gifford-Pasquill dispersion parameters. (The units have been converted as indicated in the axes labels.)

In discussing the differences in these two sets of curves, it is important to remember that the standard curves represent the expected standard deviations in the horizontal and vertical distributions of pollutants calculated for different stability criteria and downwind distances. Both sets assume a normal distribution around a plume centerline defined by the mean wind vector and have been developed from field observations over flat terrain for relatively short averaging periods (<30 min).

The curves offered in this study are derived in a totally different manner. The Gaussian form is present, but much modified in an effort to accommodate the majority of wind fluctuations in the NSDSQ and VWF terms. These two terms account for, respectively, the daily cross-valley variation in wind direction and the variation in wind speed and flow up and down the valley. In a sense they normalize the dispersion curves by accounting for the gross fluctuations related to the local wind phenomena.

The dependent variable in this model development was a twenty-four hour average. As a result, all independent variables were constructed on a

Figure 5.1a Comparison of Standard and Derived

Horizontal Dispersion Parameter Estimates

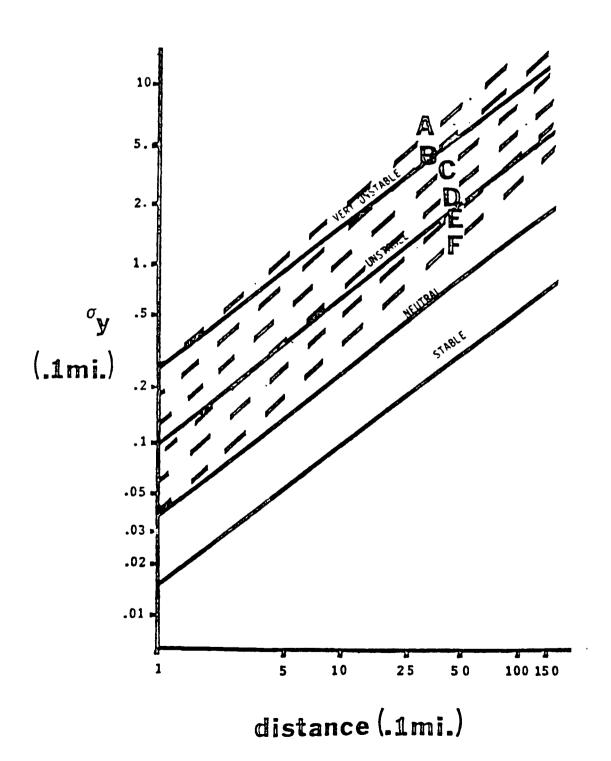
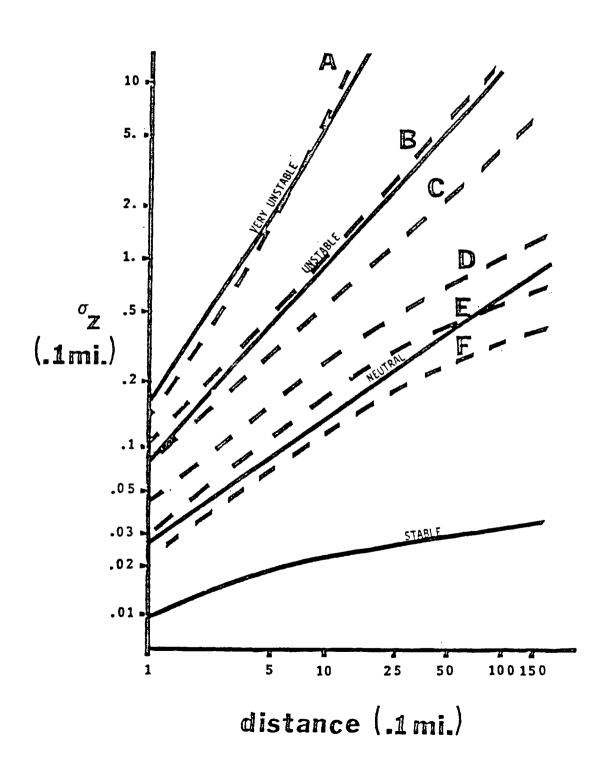


Figure 5.1b Comparison of Standard and Derived

Vertical Dispersion Parameter Estimates



twenty-four hour basis. This represented no great inconvenience because the mountain-valley drainage wind is a diurnal phenomenon. However, it is not obvious what the $\sigma_{_{\boldsymbol{V}}}$ and $\sigma_{_{\boldsymbol{Z}}}$ terms in the above derivations and charts represent. Examination of Table 5.2 shows that a certain amount of the variance in pollutant concentrations is explained by the gross wind variation terms, NSDSQ and VWF. Other empirical factors related to operations and "abnormal" meteorology explain a small percentage. However, the greatest portion of the sums of squares is explained in the three terms from which the $\sigma_{\rm v}$ and $\sigma_{\rm p}$ charts are derived. Those terms most likely represent the downwind pollutant distribution for the component wind period, averaged over twenty-four hours. The component averaging period is the one-hour mean wind. It is suspected that the $\boldsymbol{\sigma}_{\boldsymbol{V}}$ and $\boldsymbol{\sigma}_{\boldsymbol{z}}$ values derived are the expected standard deviations in downwind pollutant concentrations for one hour for a given stability category. However, that value necessarily reflects an average for all the hours of the day. This is a most important point to remember in discussing the differences in these and the standard curves.

Turner (1979) pointed out that any modeling effort has to consider the pertinent averaging period with respect to both the prevalent meteorological phenomena and the ambient standard in question. The same logic prevails here. These charts and the other significant model variables can illustrate many of the difficulties encountered in applying Gaussian form models to complex terrain, provided the pertinent averaging time is considered.

There are three obvious differences in the form of these two sets of curves. The first difference is that the estimates are similar for unstable conditions but considerably less dilution occurs as neutral

conditions are approached, and that effect is exacerbated toward stable conditions. The second inconsistency is that the slopes of the curves in the horizontal dispersion chart are notably less than their standard counterparts. The third difference is that under very stable conditions a nearly uniform distribution in the vertical with downwind distance is predicted in this study's σ_{τ} chart.

The mitigation of terrain effects under unstable conditions has been noted by several researchers (Hinds, 1970; Fosberg et al., 1976; Reid, 1979). It is likely that when unstable conditions prevail, no capping phenomena are present and uninhibited vertical dispersion would persist. Similarly, the tendency to develop calm and stable layers aloft is reduced over the twentyfour hour period. As a result, the only inhibition to normal diffusion present would be terrain channeling. Under unstable conditions terrain channeling would likely exercise its influence in the horizontal, but not for some distance downwind. The effect of terrain channeling on the horizontal dispersion parameter may be seen in the reduced slopes noted above. Other complex terrain researchers have made similar findings as reviewed by Miller (1979). However, the result in these cases is usually curved $\boldsymbol{\sigma}_{_{\boldsymbol{V}}}$ lines starting out at or near standard slopes and decreasing in slope with distance. This is, perhaps, a more appropriate form than that presented in this study, as the effect of terrain channeling would become more pronounced with plume growth relative to the valley width. Unfortunately, this model form can only accommodate straight lines in the horizontal.

Essentially the mitigative effects of instability on complex terrain dispersion may be accounted for in the absence of those phenomena that produce the confounding situations. As neutral conditions are approached the

nocturnal inversion and drainage wind phenomena become routine. Over a twenty-four hour period, a considerable period of time is spent under an inverted temperature structure, downslope winds develop, and at least two directional changes in valley flow occur. In addition, during inversion breakup a double dosage of pollutants can occur. As stable conditions develop these phenomena become more intense with increased duration. These hours or, more appropriately, the dispersion observed in these hours is included in the "average" that produces the above charts.

As very stable situations are encountered, calm conditions, intense inversions aloft, and severe limitations in mixing depth are likely. Most air quality models of the Gaussian form have recognized that, under limited mixing depths, uniform vertical distribution may develop some distance downwind. That observation may be seen in the σ_{τ} curves at stable conditions.

Several researchers have noted that it is stable conditions that are difficult to simulate in complex terrain situations. In this case, the frequency and duration of particular phenomena (that become more frequent as stability increases) are ultimately responsible. It seems that with the inclusion of mitigating or normalizing variables that account for those phenomena, and with proper consideration of the averaging period, the complex terrain situation may be discussed in an empirical Gaussian format.

5.4 APPLYING THE BASIC SOURCE RECEPTOR RELATIONSHIP

5.4.1 Methodology. Thus far, the modeling procedure has concentrated on defining the basic relationship between a receptor and a source. Having developed a satisfactory model, the next step was to apply it to all the source-receptor combinations in the valley. In practice this was a

mammoth task. However, it was greatly facilitated by employing the Geographic Information System. The details of this application are complex and can be found in von Lindern (1980b).

As the relationship was applied to each source, the impact estimates were accumulated by source category at each of the valley's nine monitoring locations. These categorical estimates were then regressed against observed ambient concentrations. This was accomplished for each day of the two-year study and done simultaneously for lead, cadmium, and TSP. The result is a calibrated model that reflects the most significant particulate sources and weights the relative impacts of the various categories. This, again, was an exhaustive and complex procedure that is detailed in von Lindern (1980b). Over 4300 observations were analyzed. The regression statistics are shown below.

The eight source categories are SMLOWEST (low-level smelter sources), SMMIDEST (mid-level smelter sources), SMHIEST (smelter tall stack). ZPLOWEST (zinc plant tall stack), AMP (ammonium phosphate plant), ACTEST (active fugitive sources), and PASEST (passive fugitive sources). They are described in detail in von Lindern (1980b) and the emissions inventory section of this report. Four of these source categories were found to be significant in predicting particulate concentrations in the Silvery Valley. They are low-and mid-level smelter sources, and both active and passive fugitive sources. The other source categories likely do contribute, but are insignificant in magnitude when combined with these sources. Final prediction statistics can be found in Tables 5.4a and b. BKGROUND refers to TSP background levels (BKGROUND = 0 for lead and cadmium).

Table 5.4a Regression Statistics for the Model:

TPOL =
$$\beta_1$$
 SMLOWEST + β_2 SMMIDEST + β_3 SMHIHEST
+ β_4 ZPLOWEST + β_5 ZPHIHEST + β_6 AMP
+ β_7 PASEST + β_8 ACTEST + β_9 BKGROUND

			with the control of t
Source	DF	Sum of Squares (SAS Type IV)	F-Value (PR>F)
SMLOWEST SMMIDEST SMHIHEST ZPLOWEST ZPHIHEST AMP PASEST ACTEST BKGROUND	1 1 1 1 1 1 1	136963 27801 696 971 993 1196 52130 5246 576519	379.9(.0001) 77.1(.0001) 1.9(.1646) 2.7(.1008) 2.8(.0970) 3.3(.0685) 144.6(.0001) 14.6(.0001) 1598.9(.1646)
Model Error Total		3771996 1527706 5299703	$1162.4(.0001)$ $R^2 = .712$
Parameter	Estimate	T-Value (PR> T)	Std. Error
β123456789	.425 4.62 27. 27.8 -109.1 -7.16 19.9 9.28 35.5	19.5(.0001) 8.8(.0001) -1.4(.1646) -1.6(.1008) -1.7(.0970) -1.8(.0685) 12.0(.0001) 3.8(.0001) 40.0(.0001)	.021 .53 5.14 17.0 65.7 3.93 2.43 1.65 .887

Total ambient particulate concentration
Low-level smelter sources estimated ambient impact
Mid-level smelter sources estimated ambient impact
Smelter tall stack estimated impact
Low-level zinc plant sources estimated impact
Zinc plant tall stack estimated impact
Ammonium phosphate plant estimated impact
Passive fugitive sources estimated impact
Active fugitive sources estimated impact
TSP estimated background concentration

Table 5.4b Regression Statistics for the Final Relative Source Impact Model:

TPOL = β_1 SMLOWEST + β_2 SMMIDEST + β_3 PASEST + β_4 ACTEST + β_5 BKGROUND

Source	DF	Sum of Squares (SAS Type IV)	F-Value (PR>F)
SMLOWEST SMMIDEST PASEST ACTEST BKGROUND	1 1 1 1	144129 33016 45310 7458 774061	402.(.0001) 92.(.0001) 126.(.0001) 21.(.0001) 2159.(.0001)
Model Error Total	5 4310 4315	3793243 1545150 5338393	$2116.(.0001)$ $R^2 = .711$
Parameter	Estimate	T-Value (PR> T)	Std. Error
β1 β2 β3 β4 β5	.423 4.20 17.9 10.9 35.2	20.1(.0001) 9.6(.0001) 11.2(.0001) 4.6(.0001) 46.5(.0001)	.021 .437 1.59 2.39 .757

This model in Table 5.4b was used to predict ambient concentrations for each quarter in the study period. Those predictions can be found in Appendix D and a quarterly summary is presented in Table 5.5. Those predictions are used to evaluate relative source contributions to ambient lead concentrations and to estimate required source reductions for achieving the NAAQS. The individual source results and a "residuals" analysis can be found in von Lindern (1980b). An observed/predicted concentrations summary follows.

Model predictions for quarterly lead means range from .26 to 7.2 $\mu g/m^3$ over the entire study area. Actual observed concentrations range from .25 to 12.5 $\mu g/m^3$. These results are summarized in Table 5.5 along with the ratio of predicted to observed values. This ratio on an annual basis consistently falls between .5 and 2.0 (a kind of unofficial measure of model quality). Moreover, on a quarterly basis the model does well in predicting means for most of the stations. This is especially true in consideration of the range of values and the fact that predictions are based on quarterly mean emissions averages.

However, the model does characteristically underpredict in certain situations. The most important are those where a few extremely high individual readings inflate the observed mean. Many of these outlying values do not seem to be meteorologically based, but do occur at different stations on the same day and always downwind from the smelter. It is most likely that these extreme values are the result of severe emissions excursions at the smelter. It is important to note that the model does not effectively predict the impact these days have on the quarterly means. This is important both in terms of utilizing the model output and developing an attainment

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Table 5.5 Summary of Mean Quarterly Ambient Air Lead Impact
Estimates, Predicted Values, and Observed/predicted Ratios

S T A T		KEY	≡ 085 PRE	ERVED DICTED	(RAT10)	,	μg P8	/m ³								
1 0 N	QUAR 3-	TER 77	4-	·77	1-	78	2-71	3	3-7	18	4-7	8	1-79		2-79)
(1) A T	.47	(.5)	. 86 . 67	(1.3)	. 76	(.8)	.37 1.22	(.3)	.20 .89	(.2)	.54 .56	(1.0)	.60 1.01	(.6)	.21 1.13	(.2)
P (2) N H	2.33	(1.2)	4.02	(3.4)	1.73	(1.6)	.99 1.50	(.7)	.85 1.62	(.5)	2.81 1.75	(1.6)	2.38 3.08	(8.)	.77 1.88	(.4)
(3) K V	6.9 3.4	(2.0)	8.7 3.8	(2.2)	4.2	(1.8)	3.4 2.8	(1.2)	2.4 2.9	(.8)	6.1 3.3	(1.9)	4.8	(1.1)	2.6 3.1	(.8)
(4) S S	12.5	(2.9)	12.3 7.2	(1.7)	3.0	(3.9)	5.2 3.0	(2.1)	3.6 3.6	(1.0)	6.0 5.5	(.9)	5.5 5.7	(1.0)	3.7 6.2	(.6)
(5) K C	6.5	(2.2)	7.1 3.2	(2.2)	4.2	(1.5)	2.4	(1.0)	2.7	(.7)	5.2 3.7	(1.4)	4.5 6.3	(.7)	2.5 5.4	(.5)
(6) K H	2.7	(1.5)	6.9 3.2	(2.1)	4.0	(1.6)	2.9	(1.2)	1.4 2.5	(.6)	5.6 3.6	(1.6)	4.6 6.0	(.8)	3.8	(.6)
(7) S B	1.5	(2.5)	3.5 .6	(6.4)	. 9	(2.2)	.5	(1.4)	.4	(1.0)	.7	(1.8)	1.9	(2.1)	.4	(8.)
(8) A L	.9	(1.9)	1.8	(.5)	. 9	(1.6)	.5	(1.0)	. 3	(.80)	.6	(2.1)	1.0	(2.7)	.3	(1.0)
(10) į					. 65	(1.5)	. 31	(.5)	.21	(.3)	1.12	(1.0)	. 70	(.5)	.31	(.3)

strategy. The effect of these days is great enough that they deserve special treatment in attainment considerations; and they are separately discussed in the next section. What is important to remember, at this point, is that the model predictions are based on average emission rates. The resultant predictions are then "average predictions" and any attainment strategy based on the model applies reductions to average emission rates. These reductions will not guarantee compliance with the ambient standard in and of themselves. Simultaneous control of the severe excursions must be accomplished as well. Descriptions of the model results by station and source category follow.

5.4.2 Model Results by Station

CATALDO. Observed ambient lead concentrations range from .008 to 2.8 $\mu g/m^3$ on a twenty-four hour basis. Quarterly means vary from .20 to .86 $\mu g/m^3$ or from 13 to 57% of the quarterly standard. Model estimates range from .56 to 1.13 $\mu g/m^3$. In general the model overpredicts concentrations for this site. According to the model, passive sources are major contributors to lead levels at Cataldo. Percentage contributions from the passive source category run as high as 44% of the quarterly mean in the third quarter (July-September). Summer quarterly mean passive estimates of .6 to .7 $\mu g/m^3$ are predicted. These are among the highest passive predictions for any site. The predominant sources of this passive contribution seem to be alluvial deposits on the nearby Mission Flats and along the river's flood plain north and east of the townsite.

The remaining source contributions seem to be dominated by the midlevel smelter contribution, 60%; low-level smelter sources, 12%; active sources, 6%. Passive sources amount to 20% on an annual basis. It is suspected that, although not statistically significant, the tall stack emissions constitute an important portion of the mid-level impact at this distance.

KINGSTON. Quarterly means for lead are observed from .21 to 1.11 μ g/m³. Model predictions are from .77 to 1.33 μ g/m³. The model suggests lead levels observed here are 80% resultant of mid-level smelter emissions, 20% low-level. There are no significant passive or active sources of lead affecting the Kingston location according to this model.

WALLACE. Quarterly means for lead for Wallace range from .25 $\mu g/m^3$ to 1.8 $\mu g/m^3$ observed values. Model estimates vary from .26 to .47 $\mu g/m^3$. The maximum does exceed the proposed 1.5 $\mu g/m^3$ limit. However, that exceedance can be traced to an extraordinary single observation of 6.625 $\mu g/m^3$ that occurred on 12/21/77. Passive source estimates run as high as .21 $\mu g/m^3$ and can account for up to 32% of the total exposure on a quarterly basis. On an annual basis it is suspected that tall stack and mid-level emissions combine for 75% of the total exposure, low-level smelter sources contribute 14%, passive local sources (primarily north of the city) 11%, and there is no significant active lead source in Wallace.

OSBURN. Quarterly mean lead levels in Osburn range from .42 to 3.54 $\mu g/m^3$ while model estimates range from .49 to .94 $\mu g/m^3$. The 3.54 $\mu g/m^3$ quarterly average observed in the fourth quarter of 1977 can be traced in great part to a single daily observation of 20.4 $\mu g/m^3$ (the same day that an

extraordinarily high value was observed in Wallace). Similarly, the proposed standard exceedances that were observed in three other quarters can be traced to single extraordinary days. If those extraordinary days are ignored, Osburn seems to be in compliance with the standard. Both passive and active source contributions seem minor, amounting to only 7% of the total impact in summer months. On an annual basis, high and mid-level sources account for 80% of the predicted impact, low-level 16%, passive 3%, and active sources 1%.

PINEHURST. Observed quarterly means range from .77 to $4.02~\mu g/m^3$. Model estimates range from 1.19 to $3.08~\mu g/m^3$. Although single maximum values constitute a large part of the quarterly average, Pinehurst does seem to be a bona fide noncompliance area even in ignorance of these values. Passive sources can account for up to 19% of the total lead levels observed here or a maximum of .37 $\mu g/m^3$. Active sources are absent. On an annual basis low-level smelter sources constitute 20% of the total, mid-level 70%, and passive 10%.

KELLOGG CITY HALL. Quarterly means ranging from 1.4 to 6.9 $\mu g/m^3$ are observed at Kellogg City Hall. Model predictions ranged from 2.5 to 6.0 $\mu g/m^3$. Passive and active sources combined constitute at the most a 7% contribution to these levels. On an annual basis mid-level sources contribute 66% of the impact, low-level sources 30%. The maximum quarterly mid-level impact is estimated at 3.41 $\mu g/m^3$, the maximum low-level impact is 2.52 $\mu g/m^3$.

KELLOGG MEDICAL CENTER. Observed levels at KMC range from 1.9 to 7.1 μ g/m³; model estimates range from 2.7 to 6.3 μ g/m³. Although these levels are

similar to those at Kellogg City Hall, it seems that both passive and active sources make a greater contribution to total impact at this station. Passive source contributions are estimated as high as .74 $\mu g/m^3$, active sources as high as .36 $\mu g/m^3$. The principal passive sources seem to come from the airport and tailings pond bank areas. Principal active sources are roads and transportation facilities in the vicinity of the smelting and milling complex, and sinter-storage activities. Combined, passive and active contributions can constitute as much as 1.1 $\mu g/m^3$ quarterly mean or 14% of the total estimate in summer. On an annual basis passive and active sources constitute 7% of total impact, mid-level smelter sources 59%, low-level 35%. Maximum mid-level smelter contribution is 3.48 $\mu g/m^3$, maximum low-level 2.69 $\mu g/m^3$. These do not occur simultaneously with each other or the passive-active maxima.

SMELTERVILLE. Quarterly mean lead levels in Smelterville range from 2.41 to 8.56 $\mu g/m^3$. Model estimates range from 2.80 to 4.80 $\mu g/m^3$. Both this station and Silver King School suffer from a number of extraordinarily high individual readings that cause the actual resultant means to be higher than model predictions in most quarters. Passive sources can be significant amounting to 15% of the summer total or as high as .37 $\mu g/m^3$. Active source estimates go as high as .13 $\mu g/m^3$ but amount to less than 5% of the total impact in most quarters. Passive sources surround Smelterville and impact from all directions. However, the McKinley Avenue area and smelter active sources contribute the bulk of the active source contribution. Maximum quarterly mean mid-level estimate is 2.24 $\mu g/m^3$, maximum low-level is 2.92 $\mu g/m^3$. Mean annual relative contribution is 54% low-level sources, 37% mid-level, 2% active, and 7% passive.

SILVER KING SCHOOL. The highest concentrations of heavy metals occur at this station. Quarterly means range from 3.6 to 12.5 $\mu g/m^3$, while model estimates range from 3.0 to 7.2 $\mu g/m^3$. Larger means are underpredicted because of the effect of the extreme individual readings that occur in several quarters. Passive contributions are estimated as high as .6 $\mu g/m^3$, active contributions as high as .58 $\mu g/m^3$. They can amount to as much as 30% and 16%, respectively, of the total quarterly estimate. Their simultaneous maximum is 1.0 $\mu g/m^3$ amounting to 28% of that total quarterly estimate. Low smelter sources dominate the estimates at Silver King School, amounting to 77% of the total on an annual basis, mid-level smelter sources account for 6%, active 4%, and passive sources 13%. Quarterly mean estimates attributable to low smelter sources are as high as 6.53 $\mu g/m^3$, mid-level sources only as high as .2 $\mu g/m^3$.

5.4.3 <u>Model Results by Source Category</u>. There is little effect of stack height associated with low-level smelter emissions. As a result they exert their predominant effect close to the smelter and decrease rapidly with distance. Mid-level sources, on the other hand, have little effect within one-half mile of the smelter. As distance increases the relative impact of the mid-level sources increases markedly. Figure 5.2 shows the relative impact of low and mid-level sources, by station, on an annual basis. Figure 5.3 shows the actual estimate components at each station on an annual basis.

Active sources exert small effects on estimates at several stations.

They seem to be important, in terms of attainment strategy, only during summer quarters at Silver King School, Kellogg Medical Center, and Smelter-ville. At these locations the active impacts can be traced to transportation activities around the smelter and McKinley Avenue area, and sinter product handling in areas peripheral to the smelter.

Figure 5.2 Percent Relative Impact Estimates for Ambient Lead by Source Category at each Monitoring Location

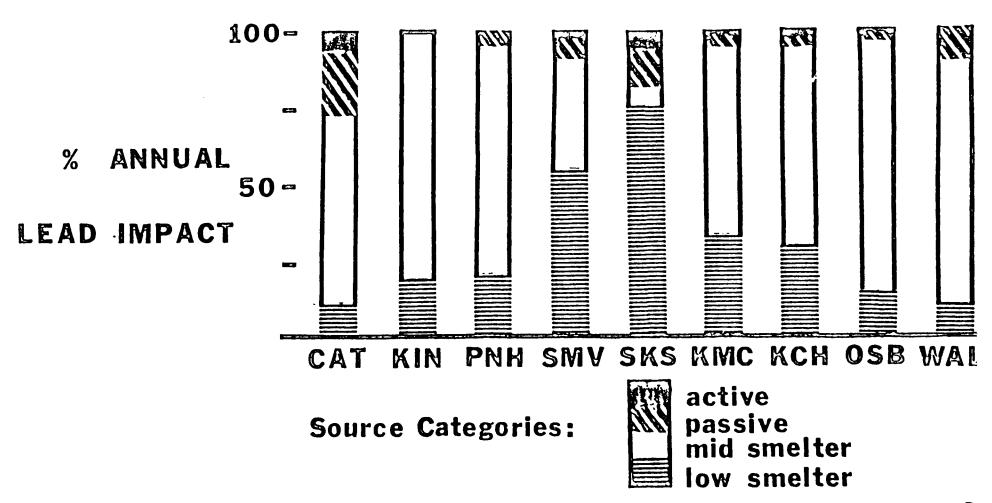
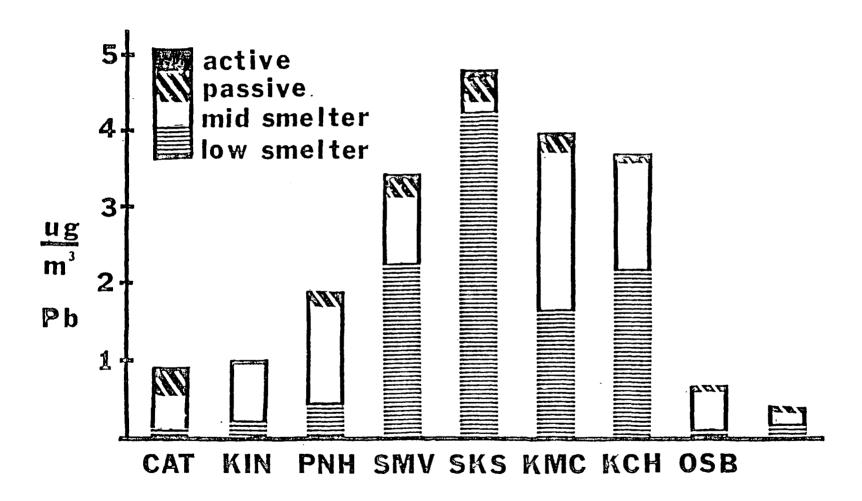


Figure 5.3 Predicted Ambient Lead Impact Estimates by Source

Category at each Monitoring Location



Source Impact Estimate Components

Passive sources exert significant impact at several monitors. In terms of percentage impact, Cataldo is most affected as a result of high lead alluvium deposited across the flood plain and the Mission Flats. In concert with other sources, however, consideration of the passive sources is most important at Silver King School, Smelterville, and Kellogg Medical Center. Both Smelterville and Silver King are surrounded by numerous high lead passive sources, especially in the immediate vicinity of the smelter. Kellogg Medical Center is exposed in the predominant wind direction to the airport area and the massive tailings pond impoundment area and features.

5.5 DISCUSSION OF MODELING RESULTS

Perhaps the most important way to discuss the modeling results is in terms of seasonal impacts of the different source categories. In analyzing the basic source receptor relationship earlier, it was evident that certain meteorological conditions are critical in determining dispersal conditions and ambient concentrations in the Silver Valley. Because these conditions vary with season and because the NAAQS is a quarterly standard this becomes an extremely important consideration. Table 5.6 shows the major components, critical season, and impact area and ambient impact estimate for each significant source category. It is evident that the critical seasons for the several source categories do not coincide. Low- and mid-level smelter sources have their maximum impact under stable conditions exacerbated by light winds and high pressure synoptic patterns that inhibit dispersion. These conditions prevail in the late fall and winter. Active fugitive sources, on the other hand, have their maximum impact under unstable conditions with light winds in the absence of moisture. This type of weather

Table 5.6 Source Categories' Critical Seasons and Impact Areas

Source category	Largest component sources	Critical season (quarter)	Critical impact area	Maximum ambient aestimate ug Pb/m quarterly mea		
Low-level smelter	Lurgi 50%	Fall, winter	< 1 mi.	≈ 7.0 μg/m ³		
	OrePrep 35%	(4,1)	Silver King, Smelterville	≃ SKS		
	Crushing 10%			~		
Mid-level smelter	Blast furnace 55%	Winter	2-4 mi.	≃ 3.5 µg/m ³		
	Pellet dryer 25%	(1)	Kellogg,	≃ KMC, KCH		
	Bujlding vents 10%		Pinehurst	≃ 2.5 µg/m ³ PNH		
Active futitive	Smelter roads	Spring, summer	< 1 mi.	≃ .6 µg/m ³ SKS		
	McKinley Avenue	(2,3)	NW Kellogg	≃ .4 µg/m ³ KMC		
	Sinter handling		Silver King	~		
Passive fugitive	Airport	Summer, fall	< 1 mi.	≃ .6 µg/m³ SKS		
	Smelter property	(3,4)	Smelterville, Silver King,	≃ .4 µg/m ³ SMU		
	Fairgrounds- lumberyard		NW Kellogg	≃ .75µg/m ³ KMC		

occurs in the spring and early summer. Finally, <u>passive fugitive sources</u> are active at neutral conditions with dry surface conditions and high winds. This weather occurs in the summer and early fall.

This situation has a tremendous impact on any strategy developed to meet the NAAQS. Table 5.7 examines the model estimates for the several non-attainment monitor locations. It can quickly be seen that the worst case situation for each of these locations occurs in the late fall and winter. Further, it is evident that the impacts during this period are nearly exclusively due to low- and mid-level smelter sources. Active and passive fugitive sources are, for all practical purposes, absent during this season.

There are some important conclusions that can be drawn at this point:

- As the critical impact season for the non-attainment area occurs in the winter when active and passive sources are practically absent, the control strategy for this season must be aimed at the smelter sources.
- As the primary impact areas for low- and mid-level smelter sources do not coincide, both must be reduced significantly in meeting the NAAQS.
- 3. A significant question remains as to the combined effect of smelter sources and passive and active fugitive sources in the summer months. Will the smelter source reductions required to meet the NAAQS in the winter be sufficient to guarantee the standard in the summer when combined with the active and passive source contributions?

The next section of this report deals with this most difficult question.

Table 5.7 Critical Quarters and Principal Sources for Non-Attainment Monitors

Location	Maximum ambient conc. μg Pb/m ³	Critical season	Component	sources	Ambient impact (% max. impact)		
	quarterly mean	(quarter)	Low-level	Mid-level	Active	Passive	
Pinehurst	3.1 μg/m ³	Winter	.9 μg/m ³	2.2 μg/m ³	0	0	
		(1)	(18%)	(82%)			
Smelterville	4.2 μg/m ³	Winter	2.9 µg/m ³	1.2 μg/m ³	<.1 μg/m ³	<.1 µg/m ³	
		(1)	(69%)	(29%)	(1%)	(1%)	
Silver King	7.2 µg/m ³	Fall	6.9 μg/m ³	.1 μg/m ³ .	.1 μg/m ³	.1 μg/m	
		(4)	(94%)	(2%)	(2%)	(2%)	
Kellogg Medical Center	6.3 μg/m ³	Winter	2.6 µg/m ³	3.5 µg/m ³	.1 µg/m ³	<.1 μg/m	
		(1)	(.23%)	(75%)	(1%)	(1%)	
Kellogg City Hall	6.0 µg/m ³	Winter	2.1 µg/m ³	3.4 μg/m ³	0	0	
		(1)	(24%)	(76%)			

6.0 STRATEGIES FOR ATTAINING THE NAAQS

6.1 ATTAINMENT CURVES

In this modeled representation, the primary sources of lead impact have been grouped into four categories. Even with this considerable simplification, development of a sufficient attainment strategy will be difficult. Each of the four main source categories can make a significant contribution to ambient levels when applied to the $1.5~\mu\text{g/m}^3$ proposed standard. Moreover, those meteorological situations that cause the greatest source impact for one category are not necessarily the most significant in other categories. As a result, great care must be exercised in selecting the proper conditions under which to evaluate an attainment strategy.

In order to determine those combinations of source reductions that will result in attainment of the $1.5~\mu g/m^3$ standard, the concept of the "limiting situation" must be introduced. The "limiting situation" is that period (quarter) and location (monitor) that requires the greatest source reduction to meet the proposed standard. The "limiting situation" for any source category is determined not only by the absolute magnitude of that source estimate, but also the relative magnitudes of the other source categories in that same period.

Achieving the proposed standard under the modeled representation requires that the following constraints be true under the "limiting situation" for each source.

$$(1 - C_{LOW}) * (SMLOWEST) + (1 - C_{MID}) * (SMMIDEST)$$

+ $(1 - C_{PAS}) * (PASEST) + (1 - C_{ACT}) * (ACTEST) \le 15.$

$$\begin{split} &\text{SMLOWEST = quarterly impact estimate of low smelter sources} \\ &\text{SMMIDEST = quarterly impact estimate of mid smelter sources} \\ &\text{PASEST = quarterly impact estimate of passive sources} \\ &\text{ACTEST = quarterly impact estimate of active sources} \\ &\text{C}_{\text{LOW}} = \text{fractional reduction of low smelter sources} \\ &\text{C}_{\text{MID}} = \text{fractional reduction of mid smelter sources} \\ &\text{C}_{\text{PAS}} = \text{fractional reduction of passive sources} \\ &\text{C}_{\text{ACT}} = \text{fractional reduction of active sources} \\ \end{aligned}$$

It follows that constraints for each source category can be developed as shown:

$$(1 - C_{LOW})(SMLOWEST) \leq 1.5 - (1 - C_{MID})(SMMIDEST)$$

$$- (1 - C_{PAS})(PASEST) - (1 - C_{ACT})(ACTEST) \text{ or }$$

$$C_{LOW} \geq 1 - \frac{(1.5)}{(SMLOWEST)} + (1 - C_{MID}) \frac{(SMMIDEST)}{(SMLOWEST)}$$

$$+ (1 - C_{PAS}) \frac{(PASEST)}{(SMLOWEST)} + (1 - C_{ACT}) \frac{(ACTEST)}{(SMLOWEST)} \text{ and, }$$

$$C_{MID} \geq 1 - \frac{(1.5)}{(SMIDEST)} + (1 - C_{LOW}) \frac{(SMLOWEST)}{(SMMIDEST)}$$

$$+ (1 - C_{PAS}) \frac{(PASEST)}{(SMMIDEST)} + (1 - C_{ACT}) \frac{(ACTEST)}{(SMMIDEST)}$$

$$C_{PAS} \ge 1 - \frac{(1.5)}{(PASEST)} + (1 - C_{LOW}) \frac{(SMLOWEST)}{(PASEST)}$$

$$+ (1 - C_{MID}) \frac{(SMMIDEST)}{(PASEST)} + (1 - C_{ACT} \frac{(ACTEST)}{(PASTEST)}$$

$$C_{ACT} \ge 1 - \frac{(1.5)}{(ACTEST)} + (1 - C_{LOW}) \frac{(SMLOWEST)}{(ACTEST)}$$

$$+ (1 - C_{MID}) \frac{(SMMIDEST)}{(ACTEST)} + (1 - C_{PAS}) \frac{(PASEST)}{(ACTEST)}$$

These equations can be used to determine the required control for any source category by defining a control regime for the other three categories. Given a proposed control regime for three of the sources, the appropriate equation is solved for each of the monitors for each quarter in the study. The maximum value of C for the fourth source as determined by this method then represents the "limiting situation" for that control strategy. If that maximum value exceeds 1.0, then attainment of the ambient standard is impossible, and further controls must be imposed on at least one of the other sources. By iterating this process for all interesting control strategies under all possible situations, attainment curves can be developed that illustrate those combinations of source reductions capable of achieving the proposed standard.

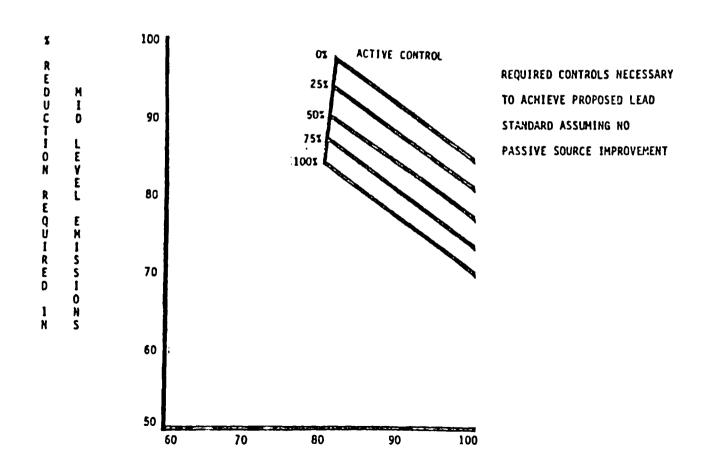
A primary concern of this study has been to evaluate the "background" contribution to ambient lead levels at various locations in the valley and how those "background" contributions can affect an attainment strategy for the area. First, a definition of background is in order. If "background" means lead levels in the absence of industrial activity, then they are probably best represented by passive source category. Active sources,

although they are akin to industrial activity, are not easily addressed in an attainment strategy. The primary active source contributors to lead levels are materials handling of smelter by-products and intermediates in areas outside the smelter proper, and dusts raised by transportation activities in the vicinity of the smelter. The exact regulatory mechanism to be employed in reduction of these sources is unclear.

It is likely that the eventual definition of "background" will be the passive source contribution plus some fraction of the active source contribution. Figure 6.1 shows the results of solving the constraint equations for the limiting situations assuming no control in passive sources and 0, 25, 50, 75, and 100% control of active sources. The abscissal and ordinal values on this graph indicate the minimum combinations of low- and mid-level source reductions required. For example, given no passive control and 50% active control, possible minimal control requirements are found along the 50% line on the graph (e.g., LOW = 90, MID = 84; LOW = 82, MID = 90). Combinations of low- and mid-level control above and to the right of the solution line will achieve the standard; those to the left and below the line will not.

The vertical line in Figure 6.1 shows that no matter how much control is exercised in the active and mid-level source categories, at least 80% low-level control is necessary. This is figured assuming no passive control. However, even if 100% passive control is assumed, solving the C_{LOW} equation for the "limiting situation" yields a control requirement of 78%. The "limiting situation" in this case is the second quarter of 1978 at Silver King School. Knowing that at least 78% control of low-level sources will be required regardless of other source reductions, the constraint

Figure 6.1. Attainment Curve for No Improvement in the Passive Source Contribution



equations are solved assuming 80, 85, 90, 95, and 100% low-level control.

One set of attainment curves was then developed for each of these low-level control strategies. That family of curves is shown in Figures 6.2a through 6.2e. In these figures mid-level control requirements are plotted against a supposed passive source reduction; the lines themselves represent a constant level of active source reduction. As with Figure 6.1, values above and to the right of a subject line will result in compliance; those to the left and below will not.

Any combination of source reductions in the four categories can be evaluated by these curves. For example, suppose no additional control on active or passive sources is contemplated, and 90% control of low-level sources is proposed. From Figure 6.2c, it is determined that 91% control of mid-level sources will be necessary. Or suppose 50% active and 25% passive control were feasible but only 80% low-level control were proposed. Under this proposal Figure 6.2a shows that 84% mid-level control would be required.

6.2 EXTREME EXCURSIONS

The inability of this model to predict extreme values at all stations has been discussed. Although that inability is not thought to affect the relative impact calculations for mean emissions, these particular days must be considered in an attainment strategy because of their effect on the quarterly arithmetic means. These days were identified and examined separately. The results of that examination can be found in Table 6.1. The data for each extreme day were qualitatively examined and each day was assigned to one of five categories. Those categories were determined by four factors found common to several of these excursions. The first

Figure 6.2a. Attainment Curve for 80% Improvement in Low-level Smelter Emissions

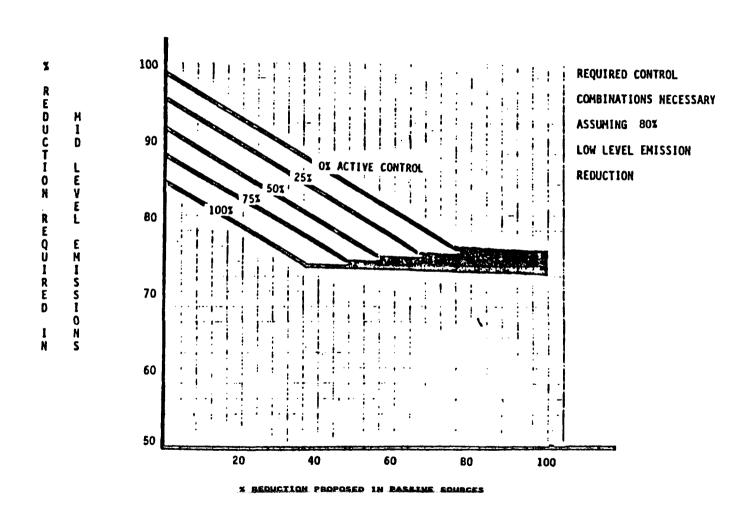


Figure 6.2b. Attainment Curve for 85% Improvement in Low-level Smelter Emissions

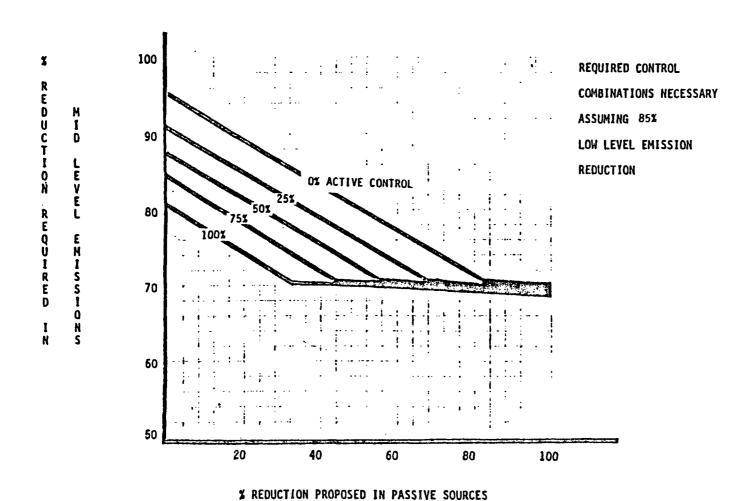


Figure 6.2c. Attainment Curve for 90% Improvement in Low-level Smelter Emissions

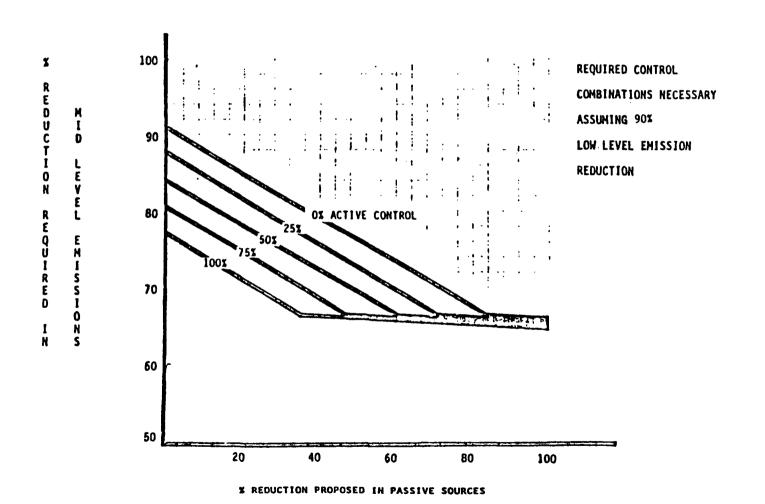


Figure 6.2d. Attainment Curve for 95% Improvement in Low-level Smelter Emissions

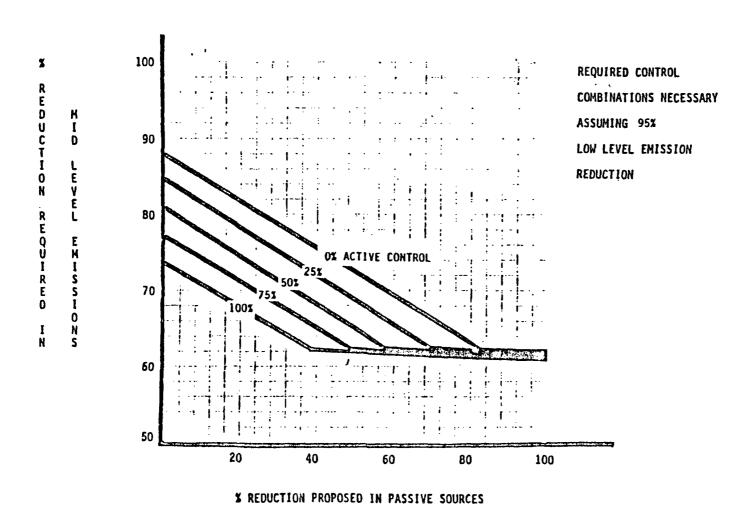


Figure 6.2e. Attainment Curves for Total Elimination of Low-level Smelter Emissions

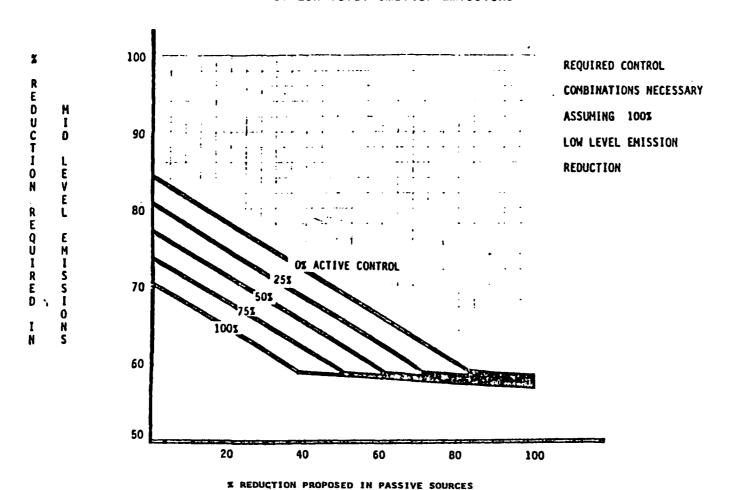


Table 6.1. Tabulation of Possible Causes for Extreme
Air Quality Excursions

SUSPECT CAUSE	ī	STATIO	N										
<u> </u>		CAT	KIN	PNH	SMV	SKS	KWC	KCH	OSB	WAL	EAST	WEST	TOTAL
DRAINAG	E WINDS	0	0	1	3	2	0.	0	0	0	0	6	6
STAGNAT	ION	2	3	7	7	6	4	3.	2	2	11	25	36
LOW WIN	ID SPEEDS	2	0	1	2	4	1	1	2	1	5	7	12
STRIKE	STARTUP	1	0	8	6	3	1	2	1	0	4	18	22
UNACCOL	INTED	1	2	8	6	6	11	6	9	9	35	23	58
# DAYS	•	4	5	25	24	21	17	12	14	12	55	79	134
SUM	HD	11.4	20.9	163.4	369.8	485.1	233.1	211.5	100.3	42.1	587	1050.6	1626.2
MEAN	E A S Y E S	2.9	4.2	6.5	15.4	23.1	13.7	17.2	7.2	3.5	10.7	13.3	12.1
# DAYS	Α	165	140	150	169	167	164	145	172	161	642	791	1433
SUM	L D	74.7	76.1	273.7	805.3	1093.3	659.6	614.2	222.0	111.6	1607	2323	3930
MEAN	Ŷ	. 45	.54	1.8	4.8	6.6	4.0	4.2	1.3	.7	2.5	2.9	2.7

condition was several days of extreme readings during the strike of 1977, and the startup period following for which no meteorological data were recorded. The second factor included days of extreme atmospheric stability or stagnation. The third situation was severe synoptic drainage winds affecting stations to the west of the smelter, and the last included periods of low wind speeds in the middle atmospheric levels. No meteorological or operational factor could be found to explain the extreme values on the remainder of the excursion days.

Three of these factors are accounted for to some degree in the model. The stagnation and drainage wind regimes and low wind speeds in the middle atmospheric levels were empirical qualifiers in the regression equation. Passive source estimates predict significant impacts with drainage wind, and extreme stability is treated by nearly a constant dispersion parameter. These factors, however, do not predict as high an impact as was observed. Perhaps the remaining two categories can suggest why. During the strike the smelter was operated intermittently by salaried personnel and considerable construction of pollution control facilities was carried on by outside contractors. Following resolution of the labor difficulties, new pollution control facilities were in use. It may be possible that less than efficient operation was practiced during these periods resulting in exaggerated emission rates. Furthermore, those days for which no meteorological explanation of the high values can be found are likely caused by abnormal emissions. Additional investigation of company operation records and upset reports may be worthwhile in establishing this point.

The remainder of the discussion here will contend with the impact, rather than the cause, of these days. The top part of Table 6.1 shows the

number of extreme days in each category for each station, for each direction, and total tabulation. It can be noted that to the west of the smelter about 8% of the severe days can be attributed to drainage winds, 32% to stagnation and extreme stability, 9% to low wind speeds, 23% to the strike/startup period, and 33% are unknown. In the easterly direction drainage winds have no effect as the monitors are upwind of the smelter. A similar percentage of extremes were attributed to low wind speeds and extreme stabilities. The strike period, however, did not seem to affect eastern stations to the degree that occurred to the west. This may be attributable to less than twenty-four operations of the smelter and the wind direction associated with the operation shift. The suspected cause of over 60% of the extreme concentrations east of the smelter is unknown.

The lower part of Table 6.1 shows the effect of these days on the total impact measured at each station. The number of extreme days at each monitor is shown together with the mean and total lead impact. These values are compared to those for all days. This shows that, at non-attainment stations, from 8 to 17% of the days account for 35 to 60% of the total impact. These days have about 4 times more impact on the quarterly mean than does an average day.

Any attainment strategy must consider these days and their extraordinary impact. Deductive reasoning concludes that those severe excursions whose cause cannot be found in the meteorology are likely due to excess smelter emissions. These excess emissions are, in turn, likely due to upsets, malfunctions, and startup/shutdown conditions. Further work in this area and the development of an effective program for preventing or minimizing is requisite to formulation of a viable attainment strategy.

6.3 DISCUSSION

There are numerous combinations of source reduction scenarios that will achieve the national standard according to results of this study. However, there are certain minimum requirements and generalizations that can be made. In order for any attainment strategy to be successful, the extreme excursions suspected resultant from process upsets, malfunctions, and startup/shutdowns must be addressed. Either additional studies should be undertaken to confirm the cause of these high concentrations, or a comprehensive program and accompanying regulation controlling these situations should be established.

Given that severe excursions are successfully addressed, the attainment curves can be used to ascertain further requirements. Under any circumstances at least 78% reduction in low-level emissions will be necessary.

Assuming an 85% reduction in low-level emissions, significant improvements in ambient concentrations can be achieved by reducing the active and passive components. Remembering that Kellogg Medical Center is the limiting situation regarding passive and active sources, improvements in those sources affecting this monitor may be worthwhile. Those sources are sinter product handling outside the smelter, McKinley Avenue, the airport area, the fair-grounds/lumberyard area, and the tailings pond embankment area. With significant improvement in these areas and 85% reduction of low-level sources, perhaps an 80 to 85% control in mid-level sources would be necessary.

Examination of Table 4.3 shows that blast furnace upsets are the dominant source in mid-level emissions. This same source is suspected as being the largest contributor to the startup-upset extreme excursions. If these upsets were eliminated, achievement of the standard would likely be much easier.

One additional point should be considered in developing an attainment strategy. All the attainment curves are based on at least a 78% low-level emission reduction required at Silver King School. Were the smelter owner to purchase all property between the Smelterville monitor and the plant, this requirement might be significantly reduced.

All these factors deserve consideration. However, in general, given the current conditions, meeting the national standard will require:

- 1. A comprehensive control program for upset, malfunction, and startup/shutdown situations.
- 2. Eighty to 85% reduction in low-level and mid-level emissions through elimination or rerouting discharges to the tall stack.
- 3. Moderate effort to reduce active emission arising from by-product handling outside the smelter, company roads, and McKinley Avenue.
- 4. Stabilization, covering, or revegetation of bare soils around Linfor Lumber, Smelterville fairgrounds, the airport, and the smelter tailings pond embankments.

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APPENDIX A

REFERENCED MAP DISPLAYS

Figure A: TOWNS, RIVER, VALLEY, MONITORS

Figure B: ELEVATION, ROADS, VALLEY

Figure C: ACTIVES, PTSOURCE, SPECSITE

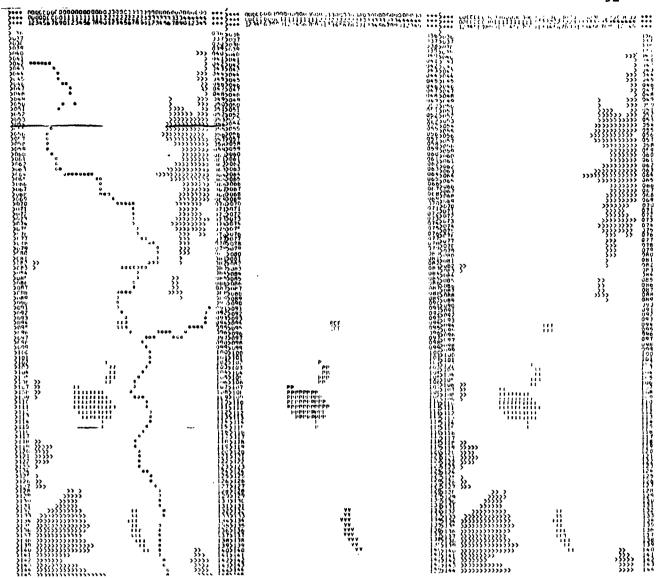
Figure D: SOILPB, SOILCD

Figure E: DOEPASS, PEDCOPAS, PESPASS,

COVERMAP, PASSIVES

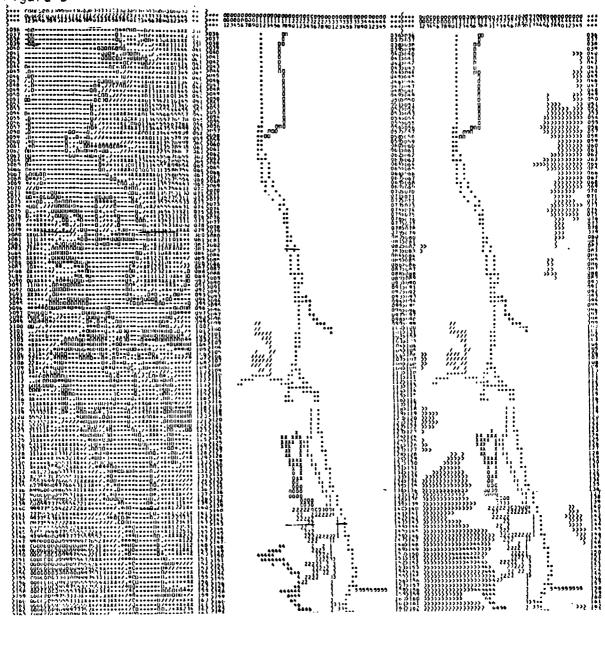
Notes for Figure A

The first map demonstrates the map VALLEY overlaid by the river and urban locations. The second map depicts the urban locations through the map TOWNS. The third map superimposes the monitor locations from the map MONITORS.



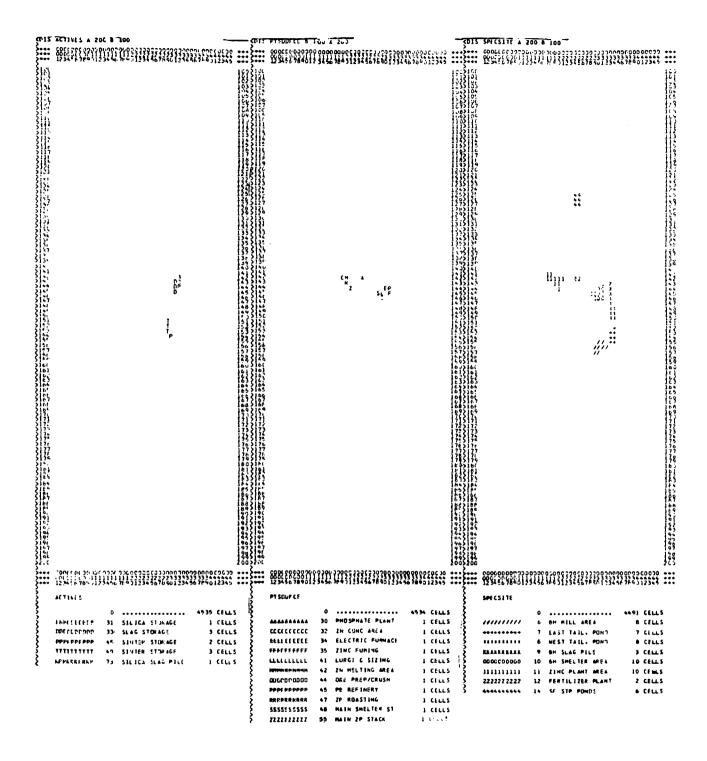
Notes for Figure B

These three maps are first the sliced version of the ELEVATION maps. It is divided into twenty levels representing 100-foot intervals from 2000 to 4000 feet in elevation. The next map shows the ROADS in the valley and in the third map values above 3200 feet elevation are masked out for VALLEY to illustrate the relative location of the roads.



Notes for Figure C

These three maps show the locations of the Active and Point Source locations, and the sites deserving special source considerations in the passive analyses. The point source labels are by process unit as defined in Chapter VIII. ACTIVES, PTSOURCE, and SPECSITE are described in Chapter VI of the parent document.



Notes for Figure D

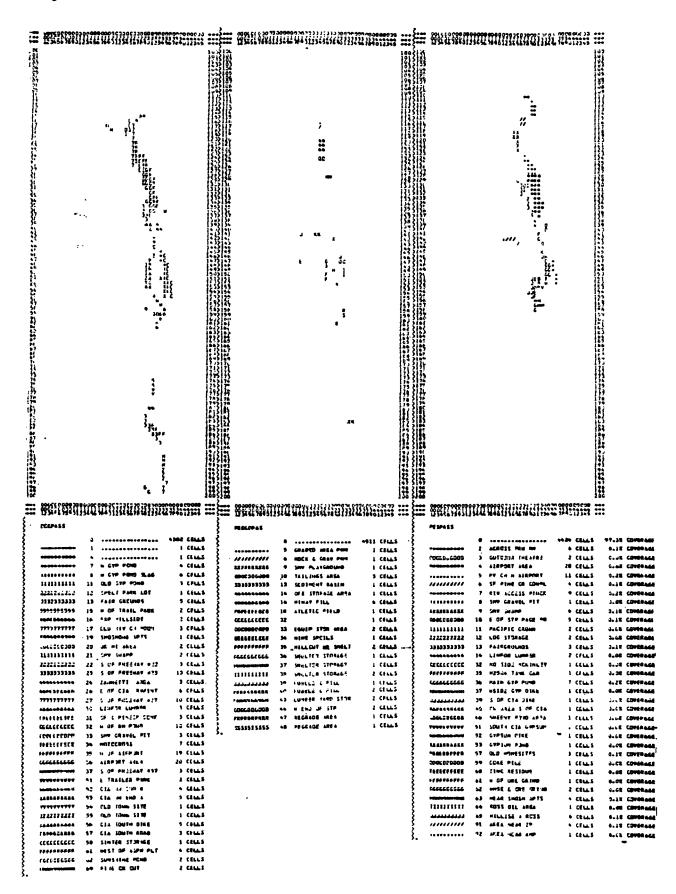
This series of maps illustrates a part of the development of the soil contamination maps SOILPB and SOILCD. The first two maps are the soil lead and cadmium estimates developed by the regression models in the report text. They are displayed here using the "slice" technique in which the different levels signified at the top of the legend are displayed via corresponding symbols below. The levels are in ppm metals. The third map is the soil contamination map synthesized from the several soil studies referenced in the report text. It shows soil lead estimates in ppm for the valley sites (river + 200 feet).

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Notes for Figures El and E2

The first three maps are the PASSIVE source locations from three studies referenced in the text. The list at the bottom of the maps refers to those in Table 2 as indicated by the variable VAL for PESPASS, (VAL - 100) for PEDCOPAS, and (VAL - 200) for DOEPASS. The next map is the vegetation COVERMAP that depicts the vegetation cover levels for the valley. Some of the areas from this map are used together with the previous three maps to produce the last in this series. That map referred to as shows all the special passive source areas considered in the valley, and the river.

Figure E.1.



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APPENDIX B

NATIONAL EMISSION DATA SYSTEM

(NEDS) Designation for Bunker Hill

Company Point Sources

(Taken From PES, 1978)

TABLE 1. SMELTER STACKS

Process Units	Process and/or Control	NEDS	Name
Sì	 Crushing Plant Dryer - Scrubber Crushing Plant Dust Collector - Receiving Bins Crushing Plant Rod Mill Scrubber Crushing Plant Conveyor - Baghouse 	02 03 04 05	Crushing Plant
S2	15. OPP Ore Preparation Plant Baghouse	06	Ore Prepara- tion Plan
\$3	16. Pelletizing Dryer - Scrubber17. Return Sinter Storage18. Pelletizing Plant Conv.	07 08	
54	 Lurgi-Strong Gas Smelter Acid Plant	01	Lurgi Sinter Machine
S5	 Lurgi "N" Rotoclone - Sinter Discharge Lurgi "B" Scrubber - Sizing Building Lurgi "C" Scrubber - Sizing Building Lurgi "A" Scrubber - Sinter Rolls - Retrofit Drum 	09 10 11 12	Sizing Building
S 6	 Lead Blast Furnace Lead Blast Furnace Feed (Sinter Tunnel) 	01	Blast Furnace
S7	Zinc Fuming Plant Main Stack - BaghouseZinc Fuming Plant Granulator - Scrubber	16 17	Zinc Fuming Plant

S8	28. 5. 24. 23.	Lead Refinery - Dross Kettles - Scrubber Reverb Norblo Flue to Main Stack Reverb Granulator Scrubber Reverb Speiss Discharge - Baghouse	32 01 17 13	Roof Vent Lead Refinery and; Reverb. Furn.
S9	6. 25.	Electric Furnace (Copper Dross) Norblo to Electric Furnace Granulator - Scrubber	01 15	
S10	7. 8. 9.	Silver Refinery Duct Retort Room Cupels Monarchs	01 01 01	Silvery Refinery

NEDS 01 Main Stack

Table 1. ZINC PLANT STACKS

Process Units		Process and/or Control			Name
Z 1	32. Concentrate Dryer - SO2 Monitor		33	33 (New)	(Concentrate
	33.	Concentrate Silo - (No Control) from Rail Car		19	Oryer)
Scrubber		Zinc Plant Acid Plant #1 #1 RSTR27 #2 RSTR28 #3 RSTR29 #4 RSTR30 #5 RSTR31		18	(Zn Roasting Plant - 5 Roasters
After B/Scrubber		Zinc Plant Acid Plant #2		18	

	36. Roasting Department Conveyor Scrubber	22	#5 Roaster = 3 Roasters 1-4 Each Acid Plant 3.5 Roasters
Z 3	34. #1 Wedge Roaster-Scrubber 39. Melting Department Dross-Baghouse 37. Roaster Dross-Baghouse	20 25 23	Zinc Dross Processing
Z 4	35. Residue Dryer - No ContTemp.	21	Residue Dryer
Z 5	Scrap Furnace Sky Vent 42. Scrap Furnace Scrubber - Not Hooked Up 40. #2 Melting Furnace Scrubber 41. #3 Melting Furnace Scrubber 41A. #3 Melting Furnace Vents 42A. Alloy Furnace	() 28 27 26 ()	
Z6	38. Purification Zinc Baghouse (Zinc Dust Prep.)	24	Purification Department
27			

Table 1. PHOSPHATE PLANT STACKS

Process Units	Process and/or Control	NEDS	Name
P.1	44. Aerotec Ammonium Phosphate Reactor	30	Separate PWR
P.2	43. Dryer S.F Ammonium Phosphate	29	Separate PWR
P.3	45. Doyle Reactor - Phosphoric Acid Reactor	31	Separate PWR

APPENDIX C

SOURCES LISTED BY SOURCE TYPES AND SOURCE STRENGTHS

POINT SOURCES ORDERED BY TOTAL SOURCE STRENGTH

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POINT SOURCES ORDERED BY LEAD SOURCE STRENGTH

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NAME	CDENR
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PROCESS FUGITIVE SOURCES ORDERED BY LEAD SOURCE STRENGTH

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ACTIVE SOURCES ORDERED BY LEAD SOURCE STRENGTH

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SINTER STORAGE CSKE STORAGE	45 45	0.6	0	0 0	C.000	0.00

PASSIVE SOURCES ORDERED BY TOTAL SOURCE STRENGTH PER UNIT AREK

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PASSIVE SOURCES ORDERED BY LEAD SOURCE STRENGTH PER UNIT AREA

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LDG STORAGE	PESPASS DOEPASS	24
NEAR SHOSH APTS SHELT PARK LOT	PESPASS DULPASS	63 12
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AIRPORT AREA	PESPASS	4
SHEENY POND AREA	PESPASS	30 46 36
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PASSIVE SOURCES ORDERED BY LEAD SOURCE STRENGTH PER UNIT AREA (CONT.)

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APPENDIX D

FINAL RELATIVE IMPACT ESTIMATIONS
BY QUARTER AND STATION

Variable Designations

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QKOUNT = study quarter (1 = 3/77 . . . 8 = 2/79)
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STATION

- 1 Cataldo
- 2 Pinehurst
- 3 Smelterville
- 4 Silver King
- 5 Medical Center
- 6 Kellogg City Hall
- 7 Osburn
- 8 Wallace
- 9 Kingston

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AMBPB = Observed Quarterly Lead Mean (\mu u/m^3)
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SMLOWX = Estimated Low Smelter Contribution (µg/m³)

SMMIDX = Estimated Mid Smelter Contribution $(\mu g/m^3)$

ACTX = Estimated Active Source Contribution (ug/m³)

PASX = Estimated Passive Source Contribution $(\mu q/m^3)$

TOTX = Estimated Quarterly Lead Mean $(\mu g/m^3)$

FRSMLOW = Fraction of Estimated Mean Due to Low Smelter Sources

FRSMMID = Fraction of Estimated Mean Due to Mid Smelter Sources

FRACT = Fraction of Estimated Mean Due to Active Sources

FRPAS = Fraction of Estimated Mean Due to Passive Sources

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Trate	2p	0.54947025	C. (570 1009	0.07274440	2.87693878	0.17416324	15.39636735	9.43166226	119.485
7 40 142 7 10 442 4 1 3 4	54 51 15	8.00075997 6.44973609 0.02721574	C. U. 12:481 C. 2:110149 C. 0857:393	0.02053035 0.15530612 0.0000000	0.21096246 1.40380379 0.38305211	0.04749152 0.01621164	1.86953405 12.34261019 0.76264078	0.001h7098 0.00315244 0.00735+68	64.763 57.012 315.200
FASX FASEL OF	2H 2H 2H	0.0337 1136	C.12640728 C.15651708 C.02422865	0.0000000 0.17#93647 0.04442CH3	0.56659393 1.61466625 0.17558360	0.07388873 0.06737540 0.00533471	0.94369419 15.91727921 3.40604796	0.01547680 0.12710443 0.00079686	375.058 62.715
FESCHIE	21: 21:	0.12154457 0.12351573 0.62701174	C. 14 110597 C. C5054472	0.34344352 0.00000000	0.89626182 0.42527212	0.02704449	23.05902979 0.75857441	0.02047932	23.206 17.377 334.231
IFPLS	2F	0.02772671	C.10191265		0.39270385 -1 OKOUNT-7 -	0.01924066	0.77634784	0.01036568	367.198
APPER CMLORA FLORA FLORA FLOME FLOME FLORA	27 27 27 27 27 27	0.00241209 0.16699369 0.711677636 0.6299778 0.17776361 1.01269146 0.13219736 0.7754329 0.67072976 0.0143066	C.6 1206718 C.78067615 C.78067615 C.77141702 C.78394934 1.74511705 C.01601255 C.15203125 C.05655405 C.18239300	0.0-743673 0.0-163553 0.72685475 0.0000000 0.00000000 0.26421278 0.04743470 0.2646714 0.0000000	2.39371429 1.52368570 5.14651056 0.23831350 1.68409787 6.67019625 0.22843191 0.86611020 0.24240849	0.17164138 0.05400653 0.17712#25 0.01374325 0.07349878 0.23962304 0.00710742 0.00695624 0.01088576	16.27781633 19.21588913 0.78130596 3.43071749 27.34266111 3.56770872 21.02847042 0.55970343 1.84411743	0.35950891 0.07875103 0.44710929 0.00509968 0.14744782 1.55031847 0.00136254 0.03687562 0.00319949	104.841 193.542 246.752 302.202 127.935 24.656 272.864
AMERI	21	3.21477714 6.06521472	0.1.7252458 0.03714609	O.COPP#796	0.44546939	0.07402905	5.55810204	0.01501227 0.00137983	57.315
SMIGHX SMIGHT ECTA	; t. ; t. ; t.	0.05521472 0.444445151 0.16643464	0.03714609 0.12126717 0.08049445	0.07 111708 0.1120*564 0.00000000	0.20811037 0.72569776 0.30399227	0.00728495 0.02378245 0.01578626	2.47568679 9.05667124	0.00137983 0.01470573 0.00647936	57.315 39.011 34.814 80.091
FESX TOT#	31 31	0.59026241	C.45015377 C.13052040	0.00000000 0.14617772	1.32447441	0.0MM28242 0.10404369	2.61309063 15.34682797 29.49227663	0.2C263842 0.2M145232	46.770
FESPLOR FESPLOR FEACT	70 21 76	0.16711097 0.41121765 0.67551627	6.05906507 6.5756744 0.07125536	0.07314408 0.09411422 0.0000000	0.19983565 0.80932960 0.30691169	0.01158440 0.05051313 0.01398021	2.784 H8509 10.69353082 1.96784171	0.00348916 0.00634099 0.00508160	55.148 62.624 94.185
FFPAS	21.	0.40591317	Ŭ. 2527A74A	0.00000000	0.81592624	0.05742035	10.55374237	0.06572451	72:131

				STATIO	N=2 QKQUNT=1				
VAR 1AI-LE	•	41. VA	STAND OF D DEVIATION	MINIMIM Valuf	MAXIMUM VALUF	STD ERROR OF MEAN	SUH	VARIANCE	C+A
AMPPS SMLOWX SMMIGX ACTX PASX TOTX FRSMLOW FRSMLOW FRACT FRACT	555555555555	2.33440730 0.34766975 1.46 th 1563 0.365309090 0.1632341 1.64623420 0.16292281 0.76568983 0.76568983	1.45274113 0.26160405 0.76442249 0.0000000 0.2556612 1.15545869 0.04010436 0.1166746 0.0000000 0.0000000	0.93790600 0.05692743 0.34743942 0.00040000 0.00046000 0.44466224 0.12771673 0.56229766 0.00000000	4.77300000 0.65661813 2.46635880 0.00000000 0.55006991 3.35011002 0.21552731 0.87221127 0.00000000 0.20217481	0.64968559 0.11699289 0.34199143 0.0000000 0.10087624 0.51673684 0.01793566 0.05214708 0.00000000 0.03711892	11.68200000 1.73834374 7.04297417 0.00000000 0.95011907 9.73144049 0.81461405 3.82634916 0.000000000	2.11045680 0.0/843668 0.57444676 0.00000000 0.05069008 1.33506480 0.00160845 0.01354659 0.00000000 0.00688907	62-17 75-24 54-27 118-70 59-36 24-61 15-22 116-23
AMNOH SMLOHX ACTX FASX TOTX FESMLOH FESSMLOH FESSMLOH FESSMLOH FESMLOH FESSMLOH FESSMLOH FESSMLOH FESMLOH FES	10 10 10 10 10 10 10 10	4.621 1000) 0.264 [5262 0.121 22729 0.121 22729 0.111 22119 1.1990 1428 0.2640 [94] 0.712 772 74 6.601000 0.0771 7766	2.6474.289 2.15071659 2.4775016 6.40002000 6.14327100 6.7471.317 C.63621539 6.1062265 0.0006000 6.11984130	0.37400000 0.37702670 0.31572449 0.00000000 0.00000000 0.34241519 0.13148474 0.50232623 0.00000000 0.00000000	10.60400000 0.65322385 1.49736757 0.000000 0.37584885 2.40526319 0.27158103 0.80391110 0.00000000	0.898H6253 0.76031083 0.13516742 0.0000000 0.04531259 0.27363412 0.01145731 0.03349188 0.030000000 0.03789715	40.710c0c00 2.66152H17 8.21227790 0.00000c00 1.11703176 11.99083282 2.08919606 7.12072730 0.000000000	8.07953F44 0.03637396 0.1F275639 0.00000000 0.02053231 0.50012218 0.00131155 0.01128414 0.00000000 0.01436194	70.69 71.65 52.05 128.27 58.97 17.33 14.91
AMRER SUMITX ACTX FCESX TCTX FESUMITO FESCT FESCT FESCT FESCT	1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1	2.75227273 6.32-44740 1.40-44446 6.56640300 0.66640600 1.75754459 0.1124617 C.1744773 0.6664400	2.23677273 0.2223.682 0.75325030 0.60000600 0.6000600 0.4737 393 0.0155 360 0.0155 360 0.0000000	0.3310C000 0.13491160 0.66012975 0.0000000 0.0000000 0.79504036 0.16464152 0.7831467 0.00000000	7.11600000 0.94567907 3.41980142 0.00000000 0.00000000 4.36648949 0.21680598 0.83030848 0.00000000	0.67441220 0.06702903 0.27711351 0.0000000 0.0000000 0.79365726 0.00470316 0.00470316 0.00400000	30.77500000 3.61163436 15.50233499 0.00000000 0.00000000 19.11396935 2.01181390 8.95818610 0.00000000	5.00315002 0.04942032 0.54738401 0.00000000 0.00000000 0.94454042 0.00024332 0.00024332 0.00024332	81. 67. 53.4 56.05 8.52 1.90
LVETH IN A COMPLEX FOR SAME IN A COMPLEX FROM I IN FERMI IN THE PROPERTY I	? 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4 4	0.95.36167 0.205111 0.9566211 0.66692000 0.27978999 1.5649320 0.17225192 0.6433706 0.17927940	C. 24397638 C. 24397638 C. 45294165 G. 60006000 C. 2776134 C. 73554768 C. 0. 456231 C. 1713920 U. 00006000 U. 19442466	0.16400000 0.64744919 0.25204457 0.0000000 0.0000000 0.50614494 0.06113440 0.31016490 0.00000000	2.13620700 0.84655223 1.69404783 0.00000000 0.96551138 3.14436008 0.44186523 0.83804908 0.00000000	0.17167755 0.04480025 0.04277261 0.0000000 0.05673862 0.15014292 0.01726131 0.03497447 0.000000000	23.88100000 6.83066675 22.66805062 0.00000000 6.71495953 36.21367689 4.13476376 15.56253062 0.00000000 4.30270562	0.35533022 0.05952155 0.26434161 0.60000000 0.07726251 0.54102950 0.00715087 0.0235712 0.000000000	59.90 85.72 47.86 99.34 48.74 49.08 26.42

				STATION	I=2 QKOUNT-5				
VARIABLE	N	MEAN	STANDARD DEVLATION	MIMIMUM VALUE	HAXIHUM VALUE	SIO ERROR OF MEAN	SUH	VARIANCE	C.V.
AME PR SMMICK ACTX FASX TOTX FFSMICH FFSMICH FFSMICH FFSMICH FFSMICH FFSMICH FFSMICH FFSMICH FFSMICH FFSMICH FFSMICH FFSMICH FFSMICH	2P 30 30 30 30 30 30 30 30	0.65713528 6.43217619 0.56227750 0.00000000 0.2347163 1.62364032 0.74363776 0.6234160 0.66341300 0.1265743	1.0712 1433 0.31764493 0.422 15397 0.00000000 0.2072 1950 0.64982963 0.13612317 0.1774111 0.00000000 0.11712974	0.11657143 0.07677745 0.70230030 0.00000000 0.00000000 0.46414009 0.0461799 0.12635530 0.00000000	5.H0571429 1.32495894 1.74889293 0.0000000 0.74663440 2.78540282 0.67711061 0.H3147234 0.00000000 0.41216886	0.70244426 0.05799377 0.07711093 0.00000000 0.03783293 0.11864215 0.0248575H 0.03610060 0.070000000	24.00539176 12.96628573 28.86871491 0.00000000 7.00420883 48.85920947 7.48511335 18.71694429 0.00000000	1.14754298 0.1C089H30 0.17838288 0.00000000 0.04293492 0.42227681 0.01852952 0.03909759 0.00000C00	124.94! 73.38! 43.89! 88.75! 39.90! 54.55! 31.69:
Thurd	75.	2.21435510	2. 47671 226	0.07341633	12.12081633	0.58534045	70.35887755	8.56558613	103-99
SYLOWA ACTA ACTA PASX TOTX FASMLOW F4 SMMIL F4 CT FAPAS	25 25 25 25 25 25 25 25	0.1/545#[4 1.347.2000 0.2000000 0.7247.4001 1.75747629 0.18301935 0.7440649 0.200010000	(1-4246924 1.18 10904 (.00000000 C.02352616 1.40582714 0.11201624 0.11201624 0.11201624 0.00000000 0.0605955	0.04544694 0.32754347 0.0000000 0.00000000 0.36407442 0.10113404 0.37594709 0.0000000	2.24255248 5.95747997 0.00000000 0.36636441 7.18621148 0.62496391 0.87211251 0.00000000 0.25377451	0.04641365 0.23660181 0.0000000 0.01670524 0.29716543 0.07240325 0.07378084 0.00000000 0.01219971	70.35887755 9.64145484 33.67550214 0.00000000 0.61995035 43.93690734 4.7048366 19.85017363 0.00000000	0.23739075 1.15739075 0.100000000 0.00647662 2.20768229 0.01254764 0.01413620 0.00000000	103.99. 124.99. 87.82. 336.82. 84.54. 59.57. 14.97!
THEOU	23	2.193.1124	2.7.91.502	STATION 0.092+1755	!=? OKDUNT=7 - 11.93142057	0.58054438	54.82742457	7 74173074	114 20
SPLOPX SCHILL ACTI FESX TOTX FESMILL FESMILL FESMILL FESMILL FESMILL FESMILL FESMILL FESMILL	21 21 23 23 23 23 23	5.696 6.276 7.17115 34 6.21 30000 0.62000007 3.67 31507 6.17765522 6.17765522 6.1765522 6.1765522	7. 4F 704743 7. 4986 9410 C. COOGOOD C. CH421734 4. 77341997 C. O. 7447159 C. U. CH46477 C. UGOOGGOO C. O. 2548767	0.104.7547 0.55777.64 0.0000000 0.0000000 0.76265210 0.13751416 0.50149723 0.0000000 0.0000000	11.57704048 11.46523740 3.09000000 0.39890904 23.24227288 0.49810777 0.86248584 0.00000000	0.49797977 0.50274014 0.0000000 0.01756053 0.98500957 0.01636579 0.01782053 0.00000000	20.62430349 49.93666488 0.0000000 0.44283825 71.04780662 4.08607008 16.7097994 0.00000000 0.20412595	7.75173076 5.76248333 5.76163868 0.06000000 0.60709256 27.31560861 0.00516030 0.00730414 0.000000000	116.79: 266.25: 110.93: 401.16: 152.92: 44.18: 10.50: 403.81
Art ha	24.	C. 171 1463	0.59495671	0.03461724	2.34974592	0.17276112	18.52355102 11.63847625	0.35874674	77.60:
SMLOP X SMLOP X ACT X	? (. ? (. ? (0.45532711 1.06513755 0.0073300	C.27027459 C.41024697 C.6000000	0.08546873 0.32~78863 0.3000000	1.02374911 2.42793939 0.0000000	0.05653165 0.10006766 0.60000000	11.63847675 27.64710675 0.0000000	0.01427153 0.26035197 0.0000000	43.75
PÁŠX 101 x	20	0 - 159 + 1794 1 - 5 1 1 1 5 1 10	(H 344 76A () . '265251H	0.00000000 0.48354686	0.85638161 J.89406767	0.0555HH66 0.16209492	9 46 16 42 637	0.0F034259 0.6F314387	76.63; 43.76
FLSMIN FLSMIN FLACT FRP4S	26 26 26 26	0.22/61241 0.55467142 0.0000000 0.15427377	(.0"71458 0.15055485 (.0000C000 C.17014123	0.07731714 0.25#34545 0.00000000 0.00000000	0.42911698 0.80103238 0.0000000 0.67433801	0.01610792 0.02952624 0.00000000 0.03336744	5.91793568 15.18642642 0.00000000 4.89563790	0.00674609 0.02266676 0.0000000 0.0289480	36.08! 25.77i 90.35

				OITATZ	4=3 QKDUNT+1				
VARIABLE	N	MEAN	OPAGNATS NUTTATV 10	MINIPUM VALUE	MAX IMUM VALUF	STO ERROR OF MEAN	SUH	VARIANCE	C.Y.
AMARA SMM IDX ACIX FASX TOIX FASM ID FASM ID FPACI FRAS	66666666666	6.911.33331 2.21.97157 0.76.92.136 0.00.12.35 0.349.3549 3.49.26.39.17 0.46.454.67 0.36.346.72 0.0157.9178 0.09942714	4.53697460 1.99400703 1.97522510 0.07047716 0.27419264 1.5746786 0.744077 0.74247751 0.07427738	1.74096000 0.10437528 0.0009000 0.0000000 0.0000000 0.63486988 0.16343124 0.0000000 0.0000000	15.63700000 5.74641053 2.38049322 0.17249170 0.36070383 6.30091571 0.9119610 0.83606876 0.03054980 0.21466938	1.97468836 0.61609120 0.41854640 0.07877259 0.13439233 0.74530341 0.12424827 0.15629806 0.0543546	41.46AC0000 13.31352744 4.59014015 0.40H77131 20.35565624 3.31802H61 2.045218592 0.09421066 0.54257482	23.39636467 3.99602912 1.05108651 0.00446717 0.10436780 3.51413587 0.05262590 0.14657451 0.00017727 0.00488728	69.986 90.089 134.012 103.469 96.669 55.035 112.318 84.794 104.250
COOPE SULTED SULTED SULTED FESULTE FES	16. 16. 16. 16. 16. 16. 16. 16.	6.57 101750 2.771 52192 0.1927.665 0.67151563 0.56161642 6.5616165165 0.5616165	6.95046650 3.71892793 0.51260584 0.17599427 0.1753612 3.7401715 0.7352508 0.746475 0.03571533 0.05674650	0.45000000 0.14140495 0.0000000 0.0000000 0.000000 0.56547444 0.76174654 0.0000000 0.0000000	26.49600000 11.74361356 1.61819557 0.48438292 0.38942634 12.33244576 1.0000000 0.75000336 0.14162618 0.21715742	1.77741662 0.92950698 0.12815171 0.03249857 0.03421653 0.91004299 0.04845626 0.0496469 0.01419712	137.10100000 44.35026413 14.26393679 1.14585009 1.45626277 61.23631879 6.10497680 7.05525248 0.35111444 0.46862225	48.3CH98456 13.82373170 0.21276577 0.01649851 0.01473234 13.24084912 0.07494015 0.07409719 0.00149646 0.00322493	81 - 114 134 - 133 57 - 419 181 - 517 150 - 375 150 - 375 63 - 376 176 - 421 165 - 955
AMP PF SML (IMA SCO) TO X AC 1 X PA ST TI(IA FF SML (IMA FF SML (I	554555445	7.72633333 1.97331541 7.22513636 6.26003090 6.2613030 4.26132477 6.46132477 6.26136030 6.26136030	6.00596558 1.8827226 1.08650744 6.00073000 0.00079000 2.6137545 6.16712114 0.16712114 0.00000000	0.69600000 0.3557495 0.4551495 0.0000000 0.0000000 0.0000000 0.2333396 0.2333396 0.00000000	30.85500000 6.39100868 4.46506958 0.0000000 0.00000000 10.10369146 0.76606012 0.76606012	2.09114392 0.4P617024 0.78053553 0.00000000 0.0000000 0.67474548 0.04185950 0.04185950 0.00000000	115.89500000 29.59973117 33.52704544 0.00000000 0.00000000 63.12677661 6.13407150 8.86192450 0.00000000	65.59324352 3.54542261 1.16050276 0.00000000 0.00000000 6.72562670 0.02628326 0.02628326 0.00000000	104.823 95.420 48.610 62.098 39.619 27.441
EMAGE SMETTA SMETTA FEGA TOTA FEGMETA FEGMETA FEGMETA FEGMETA FEGMETA FEGMETA FEGMETA FEGMETA FEGMETA	14 0 14 0 14 0 14 0 14 0 14 0 14 0 14 0	3.3445.7500 1.4717.020 1.4717.222 0.6432759 0.37315511 2.86972362 0.41314696 0.613146931 0.13143727	2. 466 1729 1. 406 1729 1. 406 1751 1. 629 1740 0. 049 1741 0. 1600 617 1. 7467 177 0. 1470 1601 0. 1241 1665 0. 1541 1459	0.22530000 0.11433710 0.00000000 0.00000000 0.00000000 0.41274473 0.10399739 0.0000000 0.0000000	11.57900000 4.73865542 2.01152069 0.16823965 1.24855424 5.66450743 0.92914714 0.73736460 0.10804857	0.46136746 0.26413650 0.12598388 0.00984288 0.07720923 0.76965145 0.04840336 0.04734451 0.0483373 0.0388373	84.66300000 34.32550498 25.58182052 1.00688236 9.328877770 70.24308556 10.37667143 10.96716390 0.38398283 0.38398283	5.32149626 1.79742960 0.39679645 0.30242206 0.14903164 1.81779767 0.05603757 0.05603757 0.0056412	68-118 97-645 61-559 122-195 103-455 47-985 58-308 53-962 157-355 117-795

				STATEO	N=3 QKOUNT=5				
ATKINUTE	N	MEAN	CPACPATZ NOTTATY 10	MIHIPUM VALUE	MAX IMUM VALUF	STD ERROR OF MEAN	SUM	VAR IANCE,	C.Y.
LAMBA TOTA TOTA	27 20 20 20 20 20 20 20 20 20 20 20 20 20	2.41114303 1.44451415 0.56413646 0.67240546 0.32495963 2.41172661 0.56771143 0.27544367 6.02513446	1.93511176 1.39744444 0.18945950 C.11033612 C.2346448 1.32647866 C.27903044 C.2772475 0.63525491 C.2763500	0.457591H4 0.00000000 0.00000000 0.00000000 0.000000	9.34816327 4.27398184 1.67474206 0.52456410 0.92538665 4.93475239 0.93276935 0.72710635 0.13073824 0.70912589	0.37760448 0.76409218 0.11147297 0.07075346 0.5451844 0.25075650 0.05273189 0.05342977 0.06666255 0.04301973	65.10059184 54.44639614 15.79583208 2.02736696 81.50874494 15.89593400 7.19162286 0.73341632 4.17902683	3.74852874 1.95265096 3.34793423 0.01217627 0.01322329 1.76060697 0.07795127 0.07993272 0.00124291 0.05181551	80.299 71.866 104.560 152.399 87.428 45.581 49.150 110.076 134.594 152.521
7~F1.b	27	6.17547101	5.24007061	0.13714296	21.54489794	0.93147379	165.30636735	23.42637252	79.016
FECT FRACT FACTA TOTA FACTA FACT FACCT FRACT	21 21 21 21 21 24 24 24	2.47.10146 6.47.525.452 0.605.4554.59 3.25.44.11.4 0.6357.704 0.6357.704 0.435.09523 0.00151.416	2. 2173.447 C. 1134700 C. 11545452 C. 1045 1111 2. 4171 1572 C. 37237696 C. 372376410 C. 04306498	0.11513011 0.00000000 0.0000000 0.0000000 0.4443155 0.1714374 0.0000000 0.0000000	14.70577782 2.50817123 0.06497477 0.46764266 14.70577782 1.09020000 0.82816026 0.01899856 0.19283865	0.60801343 0.15379507 0.0029263 0.02957935 0.55128685 0.07037263 0.07063491 0.000P8310	77.08063664 19.18724133 0.1356773C 0.95391504 92.35747032 17.79979416 9.80266785 0.04250840 0.35502559	10.15104932 0.6228183 0.06023884 0.01190072 6.50968139 0.13866460 0.13970013 0.0002184	79.016 124.977 116.938 318.938 318.862 68.439 56.577 106.761 307.802 339.170
				017476			******		
AMINON CMUCINA CCTX FACIX FACIX FACIMULU FACIA FACI FACIA FACIA	77	4.75/35/46 2.5/6//411 1.7/15/771 0.0/5/17/75 0.0/5/17/76 4.266/7664 0.5/7/211/4 0.46/6/71/6 0.0/15/71/6 0.0/15/71/6	3. 174666 4.2664-065 0.75452617 0.658-2979 3.15749713 0.99351767 0.517677 0.27516369 0.67316369	0.677.127.53 0.227.126.53 0.000.000 0.000.000 0.000.000 0.220.37.24 0.18284.571 0.300.000 0.000.000	15.95836735 10.02098336 2.63069356 0.39856110 0.70164978 10.02098336 1.00000000 0.81715429 0.03528437	0.73471029 0.60440110 0.15090724 0.01958758 0.02983975 0.75470388 0.0475956 0.05199866 0.00437753	129.17432653 81.87950699 34.47305597 0.94372271 1.51053439 118.85682005 12.33699090 0.74279253 0.37628307	14.57457675 18.11/71161 0.63764348 0.00466875 0.02493150 15.94916771 0.06340752 0.07570609 0.00053656 0.00159374	79.828 145.557 64.840 279.879 242.686 94.078 46.871 62.448 267.135 295.495
145 CE	14	1 4 1 7 13 6 14		0.28446890			44 404 3366		
AMP FR COULTY COULTY ACTY PICY TCIX BESMEIN FR SMID FR ACT FR ACT	24 24 24 24 24 24 24	2. *6179906 2. *13163120 0. *17915391 6. 27 *15 *13 3. 16111224 0. *6455649 0. 19926264 0. 03954397	1.47422176 1.15547-736 0.1537-736 0.12997617 0.1471472 1.60174649 0.73762498 0.7376271 0.01754650 0.01754650	0.1364607 0.30000000 0.00000000 0.00000000 0.007521421 0.1356653 0.0000000 0.0000000	6.17142N57 7.45AAN117 7.45AAN117 0.44HA9AAO 0.5A793174 7.995215A 0.94346739 0.69A03546 0.14102062	0.27931277 0.32466937 0.10860664 0.02549054 0.03873798 0.31569726 0.04660307 0.04755085 0.00775453	66.6 C677551 60.6 1721127 9.71146047 3.32952417 6.97069695 80.6 2891 426 17.27858313 5.16989397 1.02811960 2.52340310	2.02840620 2.14016414 0.36664144 0.61649396 0.03701641 2.55128375 0.05478816 0.05478816 0.00156345	55.595 71.008 148.262 101.498 73.675 51.909 35.757 121.937 199.994

				011412	N=4 QKNUNT=1				
VAR JARL E	N	MEAN	STANDAFO DEVIATION	VALUE VALUE	MAXIMUM VALUE	SID ERROR OF MEAN	SUM	VARIANCE	C.V.
AMPPR SMLOWX SMMLOWX ACTX FASX TOTX FASMLOW FASM	6666666666	12.48715567 3.70511064 0.02405535 0.06443079 0.55341168 4.34173451 0.77343731 0.06727834 0.01723455 0.22523680	7.17020411 4.25179065 0.07119527 0.061149527 0.66804368 4.27847520 0.37547352 0.3166448 0.0160648	4.71430000 0.00000000 0.00000000 0.0000000 0.3332093 0.0000000 0.3000000 0.0000000	23.22400000 11.63875890 0.17439208 0.14785502 1.54419416 12.30758074 1.00000000 0.52337001 0.04300994 0.95699006	2.97722357 1.73578627 0.02906535 0.02456208 0.24008416 1.71427427 0.15328658 0.08722834 0.00737520 0.15006018	74.92300000 22.23498411 0.17439208 0.39840477 3.53947009 26.34725105 4.02198188 0.52337201 0.10322720 1.35142082	51.41182697 18.07777377 0.00506877 0.00361977 0.34584242 17.71479884 0.14098066 0.04565270 0.00032636 0.13510834	57.421 114.732 244.949 90.608 99.690 95.848 56.013 24.949 105.004 163.193
AMINA	16	12.15519750	#• #5632397	0.96800000	26.99200000	2-21408077	197.53900000	78.43445630	.71 - 733
SMECKX SMECKX ACTX	16 16 16	6.471)5413 0.69313546 0.09357#31	8.67/96644 U.14231264 C.13893/23	0.27419070 0.0000000 0.0000000	28.71652612 0.48673970 0.48477040	2.16999161 0.03557821 0.03673381	109.93686608 1.48968767 1.44925290	75.34181736 0.02025294 0.01930300	71.733 126.327 152.851 153.387
PÄSX TGTX FESMLIN	17. 16 16	0.12412577 7.1/146469 6.54147999	0.70551612 0.65105557 0.17055037	0.00000000 0.44516952 0.56023927	0.64329054 28.71652612 1.00030000	0.05137953 2.16277139 0.04263759	2.01902936 114.89383500 13.47006547	0.04223770 74.84128146 0.02908743	162.946 120.474 20.258 153.286
FRSHMIO FRACT FRPAS	10 10 16	0.09969275 0.02513781 0.03329085	C. 15241460 C.04689494 C.07474618	0.0000000 0.0000000 0.0000000	0.49976473 0.15423560 0.31780846	0.03H20365 0.01170124 0.01993654	1.59507600 0.40220445 0.53265364	0.07335230 0.00219070 0.00635945	153.286 186.193 239.544
				STATIO		-			~~~~~
AMEIF SMLLHX SMMIDX	15	11.71491333 5.49119937 0.20110169	5.89791377 5.67265192	1.05300000 0.70406363	30.27100000 17.74978503	2.55563035	174.216C0000 82.37039058	97.96869698 32.17499226	85.220 103.301
ACTX PASX] <u> </u>	0.000) ეტეტ 0.000 10000	0.24454750 0.00000000 0.00000000	0.00000000 0.00000000 0.00000000	0.69524963 0.00000000 0.00000000	0.06443289 0.00000000 0.00000000	3.01652538 0.00000000 0.0000000	0.06227396 0.00000000 0.00000000	124.090
101x Fr 2~6 Un Fr 2~6 Un	5 5 5	5.69245106 0.50197311 0.69932689	5.53414172 0.12784532 0.12748532	1.12193795 0.62994620 0.00000000	17.74978503 1.00000000 0.37195380	1.42°90667 0.03301985 0.03301985	85.38691596 13.52959658 1.47040342	30.62661388 0.01635466 0.01635466	97.219 14.178 130.459
FRECT FRPAS	15	0.066443339 0.06943000	0.00000000 0.0000000	0.00000000 0.0000000	0.00000000	0.0000000	0.0000000	0.00000000	•
				\$17110					
25611% x 25611% x 56611% x	26 26 26	C.233 ()769 2.3.247737 0.53260570	4.2406 (114 2.504 16158 0.734 7602	1.13930000 0.00030000 0.00030000	18.94200000 9.70549168 0.31348639	0.44145847 0.44110649 0.01440983	162.06600000 60.37661154 0.84774822	18.40936110 6.27062522 0.00539873	68.834 107.837 225.347
ACTX PASX TUTX	26. 26. 26	0.04122734 0.49162871 1.5951916	U. 05017478 U. 76242796 2. 76437974	0.00000000 0.00000000 0.40034610	0.16027079 2.02369705 11.29907693	0.00984008 0.12991281 0.52253154	1.25391199 15.53834648 78.01661822	0.00251751 0.43481080 7.05901957	225.347 104.038 110.843
FFSMLGK FFSMMLG FFACT	26 26	0.6555753 0.6.723647	0.37623357 0.05445735 0.06586017	0.00000000	1.00000000	0.07378547 0.01067997	17.07156946	0.14155170	88.794 57.300 200.163
FAPLS	λί	0.21332322	C. 33472317	0.00000000 0.0000000	0.32560471 0.96657050	0.01370463 0.06564461	0.93265H39 7.28B403B4	0.00488324 0.11203960	194.807

				STATIO	N=4 QKOUNT=5				
VARIABLE	h	MI AN	STANDARD Di VIATIIN	HINIMUH VALUE	MAX 14U4 VALUE	STD ERROR OF HEAN	SUH	VARIANCE	C.V.
EMPPR SMEONX SCIX FESX TOTX FESCULON FESCULON FESCULON	26 26 26 27 27 27 27	3.59593105 2.67129127 0.61374912 0.67359056 0.59339674 3.57543570 0.74591426 0.61312945 0.62321145 0.20417691	2.1060f656 2.85486695 0.0644629 0.0949742 0.43403143 2.44241650 0.44241650 0.44714566 0.44714566	1.13491437 0.0000000 0.0000000 0.0000000 0.0000000	9.847142P6 10.53945020 0.323193611 1.57054313 11.89703978 1.00000000 0.16256197 0.19304849	0.39800728 0.53951914 0.01254961 0.01851561 0.10016653 0.54661534 0.07553770 0.00751546 0.00891347	100.68746939 80.39615567 0.52525574 16.77213269 99.75379957 18.16237171 0.36681343 0.92991781 e.60089755	4.43547424 8.15026532 0.00459918 0.28093332 8.36607321 0.15974527 0.00158150 0.00222460 0.13677576	58.567 99.428 353.154 88.485 61.188 61.21 303.562 142.017
*****	~				N=4 GKAUNT=6		***********		45 310
teht? teht? teht! te	74 74 74 74 74 74 74 74	5.00463757 3.67219146 0.04196604 0.01299604 0.045577446 3.0377446 0.1456734 0.11401346 0.044464 0.0465524	2.41567104 f. 12130491 C. 1735,606 C. 03150615 C. 1745 666 6. 7644 667 C. 1763 666 C. 17214766 G. 0141 464 C. 68145062	0.71/73464 0.0000000 0.0000000 0.0000000 0.000000 0.4716146 0.000000 0.0000000 0.0000000	13.96734K94 30.274K9K99 0.6757577 0.11K57749 0.67732761 30.724K9PK9 1.00000000 0.50K11592 0.28464968	0 - 79928 301 1 - 23951 377 0 - 07403008 0 - 0761 7847 0 - 03815805 1 - 27856142 0 - 03849280 0 - 03157662 0 - 06306528 0 - 01777398	143.89302041 95.47694032 2.36714619 0.245440 1.47048675 99.60051009 17.96691413 2.37324441 0.10213995 0.55766111	15.3248010 39.94625416 0.03010421 0.00049251 0.03717596 39.24344237 0.03111160 0.02465704 0.0001990 0.00663420	65.310 172.113 190.589 286.587 244.021 163.529 20.617 152.369 290.689 306.721
				STATIO	N+4 QKOUNT+7				
AMP F P SML OW X ACT X PLOX FOS PLOW FAS PMLU FAS PMLU FAS PMLU FAS PMLU FAS PMLU	26 27 27 26 26 26 21 26 26	5, 5, 6, 6, 7, 45 5, 6, 7, 2, 2, 2, 4, 1 0, 101, 2, 6, 1 0, 17, 2, 2, 1, 1, 1, 1, 1, 1, 1, 1, 1, 1, 1, 1, 1,	4. d107744 6.77427972 1.175-1494 0.13423677 1.17679057 5.12661119 1.12979149 0.17004243 1.03672434 0.01917046	0.67447714 0.0000000 0.0000000 0.0000000 0.0000000	71.51918767 25.90724736 0.52167744 1.18-46519 0.76893716 25.90374336 1.00000000 0.55544102 0.13737695 0.0851149	0.95775804 1.09122685 0.03323724 0.06316481 0.03344807 1.10112601 0.02545558 0.02545558 0.02524	144.21h55102 151.93475741 2.83483496 3.82010641 1.77511751 160.16484628 23.68326150 1.76258445 0.37365884	23. H2491 696 33.34172 HB2 0.03093199 0.11171422 D.03132566 33.94939793 0.01604765 0.01692154 0.00134868 0.00036751	2779.176 2779.176 101.861 14.250 191.686 255.536 276.148
AMC E F	21	3.45197410	Z.206Pa613	0.12734694	9.95102041	0.43280598	96 73132461	4 4703/4/1	
SPMICX ACIX PASX TOIX FREWILL FREWILL FRACT FRACT	21 21 21 21 21 21 21 26	5.1055 1439 5.01505 1439 5.01505 1439 0.5775 1939 0.435 1937 6.155 1234 0.75 17 13 193 0.17 7 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1 1	4.64773797 6.04767996 6.4644787 6.37044863 4.86303114 6.35249993 6.05286554 6.71387561 0.16285360	0.0200000 0.0000000 0.0000000 0.0000000 0.15054711 0.0000000 0.0000000 0.0000000	17.76773671 0.18315043 1.87700558 0.93690382 18.85457382 1.000000000 0.27163721 0.72601969 0.52832876	0.91992940 0.00836042 0.09117005 0.06284516 0.95371887 0.06913093 0.01036779 0.04194446	95.73132653 132.74427102 0.36571056 15.01576159 11.39809773 159.52384092 10.17242423 0.46103030 4.08417033 3.28237515	4.87034641 22.00302272 0.00181731 0.21611144 0.10268736 23.64907184 0.17475618 0.00279477 0.04574278	59.936 91.875 303.075 80.494 73.097 79.260 50.434 298.154 136.154

				STATIO	1-5 QROUNT-1	-			
TVARIABLE	ĸ	4E 14	STARBARD DEVLATION	MINIMUM VALUF	MAX IMIH VALUE	SID ERROR OF HEAN	SUH	VARIANCE	C.Y.
CONTRACT CONTRA	* 6000000000000000000000000000000000000	6.4774.667 1.34431477 1.3443146 0.6414600 0.1174.995 2.8746177 0.7644777 0.57777214 0.01424179 0.03434563	4.87196694 1.46447257 6.86447257 6.86467161 6.11666142 1.92829924 0.72867487 6.7586675 6.01654672 6.04305553	0.21000000 0.05758635 0.34393576 0.0000000 0.0000000 0.42642211 0.14154647 0.27665744 0.00000000 0.00000000	14.72300000 4.05213504 2.73879985 0.16450569 0.30871245 6.09226784 0.66621933 0.66821933 0.65830902 0.04380870 0.10506775	1.9P897717 0.59704801 0.35377300 0.07466477 0.04735738 0.7P722487 0.09335613 0.104575724 0.00675640	38.82400000 7.84856865 8.39899175 0.29481602 0.71635431 17.25877073 2.21039219 3.46783306 0.08570157	23.73606187 2.14453346 0.74481047 0.06365011 0.01345633 3.71833794 0.05229220 0.0057389 0.00027389 0.00190581	75. 293 111.951 61.817 122.957 97.154 67.037 62.073 44.270 115.865 110.954
EMB THE SUM TO P ECT X TO T X TO T X TO S Y TO S	144444444444444444444444444444444444444	7.095/5114 1.554/1661 1.574/1663 0.005/6669 4.092/5/669 4.092/5/76 0.34/1/199 0.02/1/199 0.02/6/6/	5. 35 195713 2. 1944 134 1 C. 7573 984 2 C. 99146197 1. 09466976 2. 1994 1142 2. 1785 1348 0. 1799 2409 0. 055 13686 C. 0514 -801	1.06500000 0.12965470 0.49873599 0.00000000 0.0060000 0.51839070 0.13304441 0.23304425 0.60000000 0.00000000	22.52500000 8.43971765 2.61501259 0.29562348 0.27312146 11.00489550 0.76690575 0.79033529 0.20882543 0.19292997	1.43197428 0.61696549 0.20242324 0.07444421 0.07503430 0.77758900 0.04775246 0.04784620 0.01376075	99.356C0UC0 21.76603120 21.406454C0 0.86538976 0.77398551 44.8186047 4.76179082 8.56360431 0.37965750 0.25454737	28.70770459 5.32404978 0.57365237 0.00436527 0.00477402 E.36502509 0.03192417 0.03204962 0.0031329 0.00265102	75.498 148.462 49.534 144.622 169.432 52.531 29.261 205.253 244.393
AMILIA SMI II A LCTA LCTA LCTA ITTA ILCTA FEACT FEACT FEACT	15 15 15 15 15 15 15 15 15	6.2%5,100 1.20071145 2.56217214 0.61716100 0.6311000 0.7541124 0.74160176 0.74160176 0.6000000	4.84 10 260 9 1.11732768 1.35445015 0.000000000 2.000000000 2.664 17232 0.00486413 0.00000000 0.00000000	1.9130000 0.24126452 1.01259492 0.03000000 0.03000000 1.2735944 0.14639650 0.60667078 0.6060000	16.44300000 4.01801752 7.93487750 0.0000000 0.0000000 11.95289702 0.39117922 0.81060350 0.0000000	1.25046396 0.25849277 0.47891697 0.600000 0.00000000 0.76023379 0.01675301 0.01675301 0.0000000	93.8490000 17.01100169 44.43108209 0.00000000 0.00000000 62.44204378 3.84466865 11.15533135 0.00000000	23.45490169 1.24842114 3.43848534 0.00000000 0.0000000 0.00420995 0.00420995 0.00420995 0.00000000	77.407 93.054 62.607 70.730 25.315 8.725
AMPER SMICHA SMICA ACTA PLSX TITX FESMIC FESMIC FERMIC FERMIC FERMIC FERMIC	75. 25. 25. 25. 25. 25. 25.	2. 19940000 1-0111-1447 1-25474534 6.64592489 6.1459-014 0.31707423 0.50049045 0.0141999 0.66773334	1.43464751 1.3444394 0.70585980 0.04604195 0.13533221 1.37952229 0.72372492 0.2277442 0.02369388 0.05676581	0.2340000 0.0655073 0.0000000 0.0000000 0.0000000 0.57343441 0.10147739 0.00000000 0.00000000	5.90800000 4.31591301 2.50735772 0.15943108 0.37980538 4.91943548 0.92416908 0.81562130 0.09477516	0.32693750 0.24769819 0.14117196 0.00920639 0.02706644 0.77590446 0.04474498 0.04575684 0.04573878	59.98500C00 25.3275864 31.41963439 0.83812200 3.52100408 61.10634711 7.92650585 15.02476570 0.35499491 1.693333354	7.67220325 1.53285981 0.45823806 0.00211694 0.01831481 1.90505284 0.05188565 0.005640	68-129 122-247 56-164 137-307 96-089 56-439 70-559 37-901 166-861 142-863

				101 TAT2	1=7 QKOUNT =5				
· VARIANLE	N	HEAN	DI VIATION	HIN!HUH V4LUF	HAYIMUH Valuf	SID ERROR OF MEAN	SUM	VAR JANCE	C.V.
AMEPS SPLOWN SPHICK ACTX PASN TOTX FRESPHICH FRESPHIC FRESPHIC FRESPHIC FRESPHIC FRESPHIC FRESPHIC	27 27 27 27 27 27 27 27	0.44 1528 34 0.06.342047 0.327562 39 0.00497787 0.0227164 0.44 445441 0.17711711 0.7333211 1 0.6144 1681 0.070124 25	0.33222490 0.03212032 0.03744264 0.01494377 0.02946948 0.11612951 0.03444464 0.09374016 0.02744175 0.06993539	0.05757143 0.03%60479 0.1%660%47 9.00000000 0.00000000 0.22190143 0.12433720 0.507169F9 0.00000000	1.36914286 0.18537726 0.52810146 0.06927520 0.10476592 0.65321861 0.34162387 0.84033345 0.11131472 0.22990914	0.06393671 0.00618156 0.01691302 0.00787400 0.00569073 0.07234913 0.00564070 0.01804992 0.00528687 0.01345907	11.97526531 2.16475263 7.84418447 0.24569117 0.67134082 12.14606909 4.80917153 19.79966962 0.49779398 1.89335487	0.11037339 0.00103171 0.00177336 0.00072302 0.00017438 0.01348606 0.00201587 0.00879659 0.00075525 0.00489096	74.905 39.694 26.829 164.113 91.627 25.815 25.207 12.790 149.059
AMPAR SMIDX ACTX PASX TOTX FRSMLDH FFSMLD FFSMLD FFSMLD FFSMLD	27 27 27 27 27 27 27	1.16737037 0.09125326 0.1922463 0.603744634 0.603743762 0.13204944 0.46533142 0.16533142 0.0046194	1.1c487076 c.ur463461 c.2 333103 c.6c195470 c.60545836 c.33934403 c.33934403 c.0245-254 c.0245-254 c.00316334	0.06922449 0.02474615 0.16938850 0.00000000 0.00000000 0.21463465 0.10966596 0.74663387 0.0000000	5.55000000 0.32870927 1.33444338 0.00755185 0.02857853 1.52958398 0.25314613 0.88624545 0.01187730 0.01187730	0.72417949 0.01244856 0.05452785 0.0037772 0.0015835 0.04530871 0.00558656 0.00569278 0.00168671	31.51900c00 2.46402636 15.23406120 0.01470258 0.02857553 17.74136566 3.56560602 23.36557921 0.02327353 0.04554124	1.35692393 0.00419410 0.04927774 0.00003325 0.00003024 0.115116116 0.00084266 0.00097401	99.786 70.879 50.217 360.433 519.615 21.981 3.418 360.371 519.615
EMETR SMUTHER ACTX PESX TOTX FESMUTH FFECT FFECT FFECT	25 27 27 28 28 28 28 28 28 28	1.73753547 0.14174315 0.17324412 0.03443410 0.00142666 0.93145200 0.13012936 0.465316 0.0011735	7. 3676 1762 00414457 C.86964 120 0.01240585 0.02440607 1.67131646 C.01374155 C.01579134 C.01374750	0.13034042 0.0504077 0.31649734 0.0000000 0.00000000 0.3666910 0.11664799 0.70643257 0.0000000	8.39346939 1.12064296 4.92498674 0.05474381 9.10511128 6.04557970 0.19536771 0.06742489 0.11066347	0.43610252 0.03857991 0.16435657 0.00238228 0.00461269 0.70245676 0.00259697 0.00772827 0.00573492	54.30615306 3.96M80411 21.79094723 0.124154M8 0.20794578 26.09185601 3.91802215 23.66046930 0.15904950	5.37519136 0.04167:45 0.7:636627 0.00015#91 0.00059575 1.1476#473 0.00018884 0.00167733 0.00024537	118.981 14.750 284.293 328.656 114.965 9.821 4.839 278.001 324.027
AMEDE SMETINA SCHILLINA ACTX PACTY TOTY FACTY FACTY FREAS	27 27 27 27 27 27 27 27	6.477 17234 6.167 17347 6.46641275 6.61676346 6.66764157 0.16672161 0.7677746 0.62547341 0.01204743	0.7745657 0.02571592 0.02571234 0.01614453 0.1743676 0.017470723 0.02527581 0.02848985	0.0414775 0.06370386 0.76440441 0.00000000 0.0000000 0.3476723 0.1644853 0.67520761 0.00000000	0.98918776 0.19191946 0.72102352 0.05623220 0.05724540 0.97117517 0.27918055 0.81060028 0.09674814	0.05341579 0.00495711 0.01699573 0.00306369 0.00310682 0.02311682 0.00716053 0.00716053	11.40216327 2.76440623 10.8111444 0.34853142 0.18530781 14.15939081 5.26558483 20.71056485 0.69851448	0.07703765 0.0006213 0.0077908 0.0077908 0.00024343 0.00076061 0.01546448 0.0007134438 0.00081167	65 - 725 25 - 132 25 - 132 107 - 852 235 - 217 23 - 713 7 - 491 97 - 700 236 - 443

				STATION	1=8 QKQUNT=1				
410A19AV	N	AI-71!	STANDAPD Di VIATIUN	MINIMUM VALUF	HAX IMUH Value	STO FRROR OF MEAN	SUH	VARIANCE	C, Y
AMAPA SMAIDX ACTX FFSX TOTX FFSMICH FFSCT FFSCT FFSCT	688646484	0.85016667 0.04131429 0.26635497 0.00003000 0.14450051 0.45273267 0.10113762 0.7754350 0.050000 0.22796700	1.06143490 0.07534471 0.1554-212 0.0000000 0.74715535 0.0411-495 0.49474070 0.00000000	0.28100000 0.02302361 0.14636437 0.00000000 0.00000000 0.18660434 0.03694603 0.2329331 0.2329331 0.00000000	3.01200000 0.09032388 0.48177198 0.00000000 0.59774852 0.82117676 0.13264278 0.87622237 0.00000000 0.72816066	0.43331878 0.01022609 0.06343449 0.0000000 0.0739256 0.1000483 0.01679327 0.11834048 0.0000000	5.10100000 0.24787972 1.60181441 0.00000000 0.86400308 2.71369722 0.60682209 4.02537591 0.00000000	1.17659097 0.00062744 0.02414360 0.00000000 0.05691186 0.06109071 0.00169208 0.040000000 0.10920394	124.84 60.63 58.20 165.66 54.64 40.67 43.20
Anche	15	1.775/5667	1.74114079	0.47100000	6.62500000	0.44956062	26 .6 3500000 0 .9 8320474	3.03157174	98.05 47.37
SMC()H 4 SMC()A AC ()A	6 5 6	6.04553494 0.24240142 9.0000000	0.0310-065 0.11424157 0.0000000	0.03431105 0.14435674 0.0000000	0.15989900 0.60514783 0.0000000	0.00801709 0.02949705 0.00000000	0.96320474 3.93452123 0.0000000	0.00076411 0.01305114 0.0000000	43.55
řěšx Totx	i	0.0044)7.23	0.07132927 6.14504336	0.00000000	0.00260793 0.76504683	0.00550720	0.08260793 5.00033389	0.00045494	387 - 29 42 - 91
FEGULTO FEGULTO FEGULTO	15 15 15	0.19504154 0.7/4019 0.00033903	C.01486894 C.05431639 C.000J000	0.14935436 0.55912218 0.0000000	0.21541446 0.80799614 0.00000000	0.00383914 0.01464560 0.60000000	2.92547371 11.81670283 0.00000000	0.00022109 0.00295419 0.00000000	7.62
FEPES	is	0.01711423	0.06656973	0.0000000	0.25792346	0.0171AB23	0.25782346	0.00443153	387:29
*Aeb#	1.5	U.94:11333	C.599 1-186	STATION 0.34190000	1=8 QKNUNT=3 - 2.4480C000	0.15449120	14.10200000	0.35801298	63.64
SMILKX SMILLE ZCTX	15	0.65477241 0.46459395	(.04014353 0.19759763	0.04401654 0.72146199	0.20650425 1.01226919	0.01036501	1.45189210	0.00161150 0.03904458	40.76
FÁS) 1(1x	15	0.44 130300 0.44 13 13 03 0.44 14 13 75	0.00000000 0.0000000 0.23770666	0.00000000 0.000000 0.26590654	0.00000000 0.00000000 1.21877344	0.0000000 0.00000000 0.06137559	0.00000000	0.0000000	40.87
F# 5~1 (i);	15 15	0.16524456 0.63171144	C. GO26 1549 C. GG26 1549	0.1/332/A1 0.1/332/A1	0.17133410 0.83667119	0.00067542 0.00067542	8.72225628 2.45432842 12.50567158	0.05050445 0.0000684 0.00000684	1.57
FACT	15	0.60000000 0.6000000	0.00000000 0.00000000	0.0000000 0.0000000	0.0000000	0.0000000 0.0000000	0.0000000	0.0000000	•
				NOTTATE	I=8 QKOUNT=4 -				
X11 π 45 ΣωΓιώ z Σων t lu _,	26 25 34	0.47412309 0.14175057 0.22319142	C.27357341 C.01395309 C.06633652	0.11790000 0.3174/497 0.6975/742	1.29400000	0.05471468	11.873C0C00 1.04376420	0.07484741	57.60 33.42 31.08
ACT) PAST	25 25	0.25030006 0.21 % 1993	0.00000000 0.30978591	0.0000000	0.34773472 0.00000000 1.07724402	0.01367610 0.00000000 0.06195718	5.50003545 0.00000co0 5.26349418	0.00467590 0.00000000 0.09596731	31.08 147.13
FF 5 % (25 75 26	0.47221191 0.11577423 0.61145369	C. 31757196 C. 04849900	0.16743210	1.2399575Ā 0.17658141	0.06351439	11.80729786 2.89435507	0.10085195 0.00235215	67.24 41.89
FFACT	5e. 25	0.64933603 0.27936491	C.25751919 C.00000000 C.30517642	0.11107299 0.00606000 0.0000000	0.84805293 0.00000000 0.86884497	0.05150384 0.00000000 0.06103528	15.34652213 0.00000000 6.75912280	0.06631613 0.00000000 0.09313265	41.95 112.87
					3 6 6 6 6 6 7 7 7 7	0.0010,720	0.77712280	0.07313265	112.01

				STATION	-10 QKOUNT-3				
3 14 4 1 8 4 4	N	HEAM .	STANDARD DEVIATION	HINIPUH VALUF	HAX 1MUH VALUE	STD FRROR OF HEAN	SUH	A'S I WHCE	C,V.
AMPPA SMLOWX SMMIDA ACTX COTX COTX COTX COTX FRAMID FRACT FRPAS	5455454555	0.65213133 0.25454736 1.04945251 0.00100000 1.34303947 0.17623533 0.47570467 0.10013000 0.00013000	0.48482054 0.17178272 0.46335930 0.00000000 0.69164645 0.00733140 0.00733140 0.00000000	0.15800000 0.10100799 0.50200724 0.00000000 0.00000000 0.60745084 0.16616411 0.80557943 0.00000000	1.69400000 0.52449918 2.36133956 0.00000000 0.00000000 2.90563674 0.19047057 0.83363549 0.00000000	0.12518012 0.03325143 0.14546907 0.0000000 0.00000000 0.17859371 0.00189296 0.00189296 0.00000000	9.78500000 3.51881034 16.476768 0.00000000 0.00000000 19.79559802 2.61352798 12.38647002 0.0000000	0.28505095 0.01658486 0.31741877 0.00000000 0.00000000 0.47843568 0.00005375 0.00005375	13 32 1 00 0Ci
APP PR SPLOHX	25 25	0.41117000 0.14517143	(.22037852 C.09169140	0.02400000	0.72300000	0.04407570	7.77800000	0.04456669 0.00440731	70.83
EPHIGA ACIX PASX	25 25 25	0.72355275 0.06303000 0.06303333	0.24175296 0.00000000 0.00000000	0.14715849 0.0000000 0.0000000	1.24644A25 0.00000000 0.00000000	0.05235059	15.71361861 0.00000000 0.00000000	0.0(851461	41.66
ነበተአ የቀንሥር በሁ -	25 25	0.77365419 0.17767617	0.34410760 0.04053133	0.77144h44 0.15396251	1.64356111	0.06A82152 0.00A10627	19.34135456	0.11841004	\$9.67! 27.60
FRSMMID FRACI FRACS	25 25 25	0.70(m)3030 0.70(m)3090 0.00000000	0.04051133 0.0000000 0.0000000	0.66948459 0.00000000 0.0000000	0.84601749 0.00000000 0.00000000	0.0000000	20.55759574 0.00000000 0.00000000	0.00000000	6.631
				STATION	10 QKQUNT-5 -				
AMEPE SMLOH X SMMLOX ACTX PASX TOTX FESMLOH	26 26 26 26 26 26 28	0.21542347 0.1r352451 0.62276792 0.00000000 0.00000000 0.60629152 0.22493066	C.1937H908 C.09737457 U.214660A1 0.60000000 U.00000000 0.27563377 C.09057137	0.0348.714 0.06109911 0.25491152 0.00000000 0.00000000 0.37440195 0.16238608	0.91765306 0.43334884 1.04574650 0.00000000 0.00000000 1.42191193 0.50625328	0.03662769 0.0140301 0.04056704 0.00000000 0.00000000 0.05206969 0.01711770	6.03185714 5.13874219 17.43747646 0.00000000 0.00000000 22.57621864 6.29721835	0.03755421 0.00948278 0.04607926 0.00000000 0.00000000 0.0797398 0.00820444	99.99. 99.000 10.0000 10.0000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.0000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.0000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.0000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.0000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.0000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.0000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.0000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.0000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.0000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.0000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.0000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.0000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.0000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.000 10.0000 10.000
FESTATO FRACT FRPAS	50 50 50	0.775,7495	0.04057417 C.00000000 C.0000C000	0.44374672 0.0000000 0.0000000	0.63761392 0.0000000 0.0000000	0.00000000	21.70278125	0.00020444	11:88
TALLA	,,	1.11914709	1.47435457	STATION: 0.0271H367	•10 QKDUNT •6 · 5.99387755	0.28393308	20 211571/3		*****
5~[0]6* 5~416.x 2C1* F&SX	27	0.12729764 0.56472592 0.66372000 0.66372000	0.1/0/4/050 0.1/10/4/06 0.1/10/00/06 0.1/10/00/00	0.02743092	0.77145110 0.77147718 0.00000000 0.0000000	0.15374 0.15575013 0.16575013 0.60000000	30.21157143 5.05674997 26.05299106 0.00000000	2.17668586 0.03274309 0.45665199 0.40000000	131 - 951 03 - 901
TOTX FLEWIDS FLEWIT FLEET	(1) (1)	1.15721285 0.15375696 0.84321456 0.64321456	(0.20000755 0.11556460 0.47433258 0.000000	4.14021AA7 0.32566742 0.80439540 0.0000000	0.18676542 0.00895452 0.00895452 0.00895452	31.10974703 4.07124744 22.92875256 0.0000000	0.44179573 0.00216495 0.00216495 0.0000000	30.85. 30.85.
14115	77	6.7(10)000	(noncnoo	0.0000000	0.00000000	0,0000000	_ 0.00000cad_	1-20000000.	- •