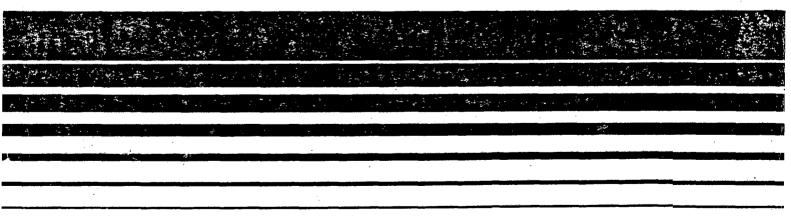
Air



A Screening Final Procedure for the Report Impacts of Air Pollution Sources on Plants, Soils, and Animals



# A Screening Procedure for the Impacts of Air Pollution Sources on Plants, Soils, and Animals

by

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## 1.1 BACKGROUND

Section 165 of the Clean Air Act<sup>1</sup> requires preconstruction review of major emitting facilities to provide for the prevention of significant deterioration (PSD) and charges Federal Land Managers (FLMs) with an affirmative responsibility to protect the air quality related values of Class I areas. Regulations<sup>2</sup> implementing these provisions require:

- An analysis of the impairment to visibility, soils, and vegetation (52.21 (o)) and
- A notice from the EPA Administrator to the appropriate FLM of any permit application from a source whose emissions would affect a Class I area (52.21 (p)).

For sources more than 10 km from any Class I areas, exemptions provide that no analysis of impairment need be done if emission increases are below specified limits.\* The analysis should address the impairment due to general secondary growth associated with the source and need not address the impacts on vegetation having no significant commercial or recreational value. For impacts in Class I areas, consultation between EPA and the FLM is required.

#### 1.2 SCOPE

The entire subject of air quality related values and impairment to these values is currently under investigation. For example, although some values related to plants, soils, and visibility are "air quality related values," the term itself remains to be defined in a fashion appropriate to the review of PSD permit applications and air quality reviews. Much of the data required to relate ambient concentrations of pollutants to impairment of these values is currently lacking. However, the requirements of 52.21 (o) and (p) need to be addressed now while additional investigations are being carried out.

<sup>\*</sup>The "de minimis" values are given in Sec. 52.21 (b)(23)(i) of the PSD regulations.<sup>2</sup>

The information and screening procedure presented here provide interim guidance:

- To aid in determining whether emissions are significant or whether there are significant air quality impacts under Sec. 52.21 (o) and
- To aid in flagging sources which should be brought to the attention of an FLM under Sec. 52.21 (p).

Impacts on vegetation and soils are the principal areas addressed by the procedure which thus takes a limited view of the possibly broad scope of air quality related values. A selected review of impacts on fauna has also been included and the odor potential of regulated pollutants is addressed.

This procedure is intended for use by air quality engineers and is not a manual for the assessment of impacts on plants, soils, and other air quality related values such as would be suitable for an ecologist. A handbook providing for such detailed assessments is being prepared for the FLMs. In keeping with the screening approach, the procedure provides conservative, not definitive results. However, a source which passes through the screen without being flagged for detailed analysis cannot necessarily be considered safe. Species more sensitive to particular pollutants than species considered in this study probably exist. Further research may indicate that averaging times different from those used here are controlling. When available, such information could be easily included in the screening procedure by changing the screening concentrations presented here.

Based on estimates of typical stack parameters, significant emission levels have been estimated. These estimates are not intended to replace source-specific screens, but do indicate what sizes of sources appear most likely to cause significant impacts on plants and soils.

The procedure presented here provides a simple method for assessing the potential a source has for adversely affecting some air quality related values. In particular, the potential for impacts on plants, soils, and animals is assessed. The approach taken is similar to the "de minimis" approach used by EPA in the PSD regulations.<sup>3</sup> In the procedure presented here, the minimum levels at which adverse effects have been reported in the literature are used as screening concentrations. These screening concentrations can be concentrations of pollutants in the ambient air, in soils, or in aerial plant tissues. They have been developed by searching the review literature; few original sources have been consulted. The analyst applying this procedure must read the material in Sec. 3 which lists these screening concentrations and provides background on them in order to apply and interpret them appropriately.

Section 5 describes a seven step process for screening a source. procedure begins by estimating the maximum ambient concentrations caused by the source for the averaging times specified for the screening concentrations. For some pollutants these maxima are compared directly to the screening values. For other pollutants (trace elements) estimates of deposition in the soil and subsequent uptake by plants are made based on an estimate of the maximum annual concentration. The estimated concentrations of the pollutant in the soil and aerial plant parts are then compared to appropriate screening concentrations. Concentrations in excess of any of the screening concentrations would indicate that the source might have adverse impacts on plants, soils, or animals and that the actions required by 40 CFR 52.21 (o) and (p) need to be taken. For situations where modeling results are not available for the source, significant emission levels corresponding to the various screening concentrations are developed in Sec. 5.2. In these cases, emissions in excess of the significance levels would trigger the additional actions.

The estimation of potential impacts on plants, animals, and soils is extremely difficult. The screening concentrations provided here are not necessarily safe levels nor are they levels above which concentrations will necessarily cause harm in a particular situation. Effects data for plants, animals, and soils are under constant revision and reevaluation. There is

good deal of controversy among experts. In addition, this procedure is based upon a simplistic view of extremely complex systems in which single value estimates are not possible and in which the number of variables is extremely large. Many simplifying assumptions have been involved in developing the procedure and are discussed in Sec. 3.

Ideally, the screening procedure should address the impacts of all the pollutants currently regulated under the Clean Air Act, but as shown in Table 2.1, screening concentrations were found for only half the regulated pollutants. Ozone and TSP are discussed in Sec. 3.1. Of the remaining substances for which screening concentrations were not found, methyl mercaptan, dimethyl sulfide, dimethyl disulfide, carbon disulfide, and carbonyl sulfide are regulated because of their odor potentials. Odor is an air quality related value and Sec. 52.21 (b)(23)(i) of the PSD regulations<sup>2</sup> gives "de minimis" emission levels for reduced sulfur (RS) and total reduced sulfur

Table 2.1 Regulated Pollutants

Screening	Concentrations
Available	Not Available
СО	TSPa
NO <sub>2</sub>	Asbestos
so <sub>2</sub>	Sulfuric Acid Mist
03 <sup>b</sup>	Vinyl chloride
Le ad	Methyl Mercaptan <sup>c</sup>
Mercury	Dimethyl Sulfide <sup>C</sup>
Beryllium	Dimethyl Disulfidec
Fluoride	Carbon Disulfide <sup>C</sup>
Hydrogen Sulfide	Carbonyl Sulfide <sup>C</sup>

<sup>&</sup>lt;sup>a</sup>Fraction of TSP present as trace elements treated through deposition and uptake by plants.

bScreening concentration available but no simple procedure for estimating the ozone impact of a single source is currently available.

<sup>&</sup>lt;sup>c</sup>Regulated indirectly as constituents of reduced sulfur or total reduced sulfur.

\_\_\_\_

(TRS) based on odor. RS and TRS include these sulfur compounds. Sources not emitting more than these "de minimis" levels (10 t/yr for both RS and TRS) are not expected to have a significant odor impact and hence should not require any additional review for impacts on air quality related values. If the 10 t/yr "de minimis" level is exceeded, the appropriate FLM might want to evaluate the potential for an odor problem. Whether or not these sulfurcontaining compounds might adversely affect plants, soils, or animals could not be determined. There was one questionable indication that methyl mercaptan might be toxic to plants at concentrations near 150,000  $\mu g/m^3$ , far above likely ambient concentrations.<sup>4</sup> Information for asbestos, sulfuric acid mist, and vinyl chloride was not available in the review literature consulted for this work.

Pollutants which can be screened by this procedure are listed in Table 2.2 according to whether they are screened for potential effects on plants or on animals and according to whether the potential effects are caused directly by concentrations of the pollutant in the ambient air or whether the potential effect is exerted indirectly through the soil or the diet. Absence of a pollutant from a particular column in the table does not necessarily mean that impacts can not result from the pollutant acting through the corresponding pathway. Such absence simply means that no data to provide a suitable screening concentration were found in the review literature consulted.

Table 2.2. Pollutants Screened

	Potential	Impacts on				
	Plants	Animals				
Direct Ambient Impact	Indirect through Deposition and Uptake	Direct Ambient Impact	Indirect through Plants in Diet			
so <sub>2</sub>	Arsenic	· · · · · · · · · · · · · · · · · · ·	Arsenic			
03	Boron	Beryllium				
NO <sub>2</sub>	Cadmium		Cadmium			
CO	Chromium					
H <sub>2</sub> s	Cobalt		Cobalt			
Ethylene	Copper		Copper			
Fluoride	Fluoride		Fluoride			
	Lead	Lead	Lead			
	Manganese		Manganese			
	Mercury					
	Nickel		Nickel			
	Selenium		Selenium			
	Vanadium		Vanadium			
	Zinc		Zinc			

<sup>&</sup>lt;sup>a</sup>The other five sulfur-containing compounds are screened for odor impacts during the "de minimis" determinaion for RS and TRS.

#### 3 AIR QUALITY RELATED IMPACT DATA

NOTE: In this chapter and throughout this work, a distinction is made between parts per million by volume (ppmv) and parts per million by weight (ppmw). The former, ppmv, is the unit more familiar to air quality analysts and is used, for example, to express ambient concentrations and standards. The latter, ppmw, or an equivalent (mg/kg, µg/g), is frequently used to express concentrations of elements in soils, plants, and animals. The air quality analyst should be aware of the difference, because the units are not equivalent. The unit ppmv is normally used only in expressing concentrations of components of gaseous mixtures.

#### 3.1 GENERAL

Data to be used in screening impacts on three air quality related values (vegetation and crops, soils, and fauna) are discussed in this section. Vegetation and crops receive the greatest amount of attention, reflecting the availability of data. No direct impacts on soils are defined, such impacts being screened through the potential impacts on vegetation growing in soils which have become contaminated by the deposition of air pollutants. Impacts on fauna are also addressed indirectly with effects being related to the ingestion of plants containing toxic elements taken up from pollutants deposited on soils. Thus, the information presented here represents a preliminary definition of air quality related values and impacts.

Perhaps as important as the areas addressed are several areas not addressed in this procedure. These areas are visibility, acid precipitation, a screen for TSP, and a screen for ozone. Consideration of visibility as an air quality related value is required by regulations (40 CFR 52.21 (o) and (p)). Addressing visibility was beyond the scope of this work. However, EPA has prepared a report to Congress on visibility<sup>6</sup> and draft regulations<sup>7</sup> have been published.

No simple procedure is currently available to deal with the impact of a single source on acid precipitation. Acid precipitation presents a regional problem involving long-range transport which makes the impact of a single-source difficult to isolate. Various adverse effects on vegetation have been noted in areas with low soil buffering capacities and subject to heavy

annual precipitation. Such areas appear to be most susceptible.8,9,10,11 Observed effects include reduced growth, reduced germination of seeds and pollen, accelerated leaching of nutrients, decrease in soil calcium and other bases, and reduced microbial activity, particularly that of nitrifiers and nitrogen-fixers. A major EPA initiative to study acid precipitation is currently underway. Policy and guidance will be formulated as part of this initiative.

Total suspended particulates (TSP) are not considered here. No useable information other than that used to develop the ambient standards (NAAQS) was found in the review literature. Thus, EPA's current procedure for TSP<sup>3</sup> should suffice for the review of generic TSP. However, the trace metals in TSP may have greater impacts on vegetation and soils than the total amount of particulates. This section provides information related to specific trace metals.

No simple models are currently available to estimate the impacts on ozone concentrations of emissions of volatile organic compounds (VOC) from a single source. EPA is currently developing means other than modeling to deal with VOC emissions and ozone. It appears likely that an emission management approach will be taken. When this approach has been completed it could probably be used to review new sources for impacts on air quality related values. Meanwhile, the minimum reported concentrations at which vegetative damage occurs are presented here but no method for their use is given and no significance levels for VOC emissions have been developed.

#### 3.2 NATURAL VEGETATION AND CROPS

## 3.2.1 General

Two pathways by which air pollutants can affect vegetation are considered here. The first is the direct exposure of a plant to a gaseous pollutant in the ambient air. The second involves indirect exposure to trace elements through deposition of the pollutant in the soil and later uptake by the plant. For each pathway certain qualifications and cautions should be kept in mind in order to avoid interpreting the values presented here either as absolutely safe levels for all plants or as levels which could never be exceeded without damaging vegetation. The following discussions are not intended to be exhaustive and details required by specialists are not given. The intent is to

provide the air quality analyst with a feeling for the difficulty of estimating screening concentrations for plants and the complexity of making detailed assessments of impacts on vegetation. References 8, 9, 12, and 13 may be consulted for additional details and guidance to primary source material.

Effects of pollutants can be classified as acute or chronic. Acute effects result from short-term (e.g., 3-hr) exposures to relatively high concentrations. Chronic effects result from exposures to lower concentrations for times of from months to several years. Most of the effects data for plants comes from experiments conducted under acute conditions of exposure with some limited information on chronic exposures. Thus, the data may not adequately reflect impacts which take years or decades to develop.

The values presented here represent the ambient levels at which visible damage or growth retardation may occur or the observed minimum levels at which injury and mortality to plants have been reported. These numbers are generally the lowest values consistently reported in the literature on plant response to controlled exposures of single pollutants. Both field and greenhouse studies have been used in developing the data. Experiments which demonstrated only physiological changes (e.g., a change in respiration rate) without associated visible damage or effects on growth, weight, or yield were not considered in this compilation.

The majority of the studies were performed on crops and other economically important species; for lack of sufficient data, it is assumed here that native plant species are affected at similar concentrations. In addition, assessment of the data on crops is difficult because of the number of horticultural varieties available for many of the species tested. In the process of selecting desirable attributes in different varieties, the species' original sensitivity or resistance to the element being tested may have been inadvertently altered, making general conclusions about the sensitivity of the species as a whole difficult.

Effects from simultaneous exposure to two or more pollutants have been ignored in the majority of the studies. Exposure to a single pollutant at a time is not the usual situation. Particular combinations and concentrations of pollutants may act either synergistically or antagonistically under certain conditions. Such situations are seldom clearly predictable with

current information and the screening procedure presented here does not deal with them. A limited discussion of synergisms is pesented in Sec. 3.2.3.

Each species exhibits a specific range of tolerance which may be higher, lower, broader, or narrower than another species'. In addition to the variation in tolerance between species, every individual of a given population has an intrinsic tolerance to environmental stress. Therefore, the population exhibits a characteristic range of tolerance so that all members of the population would not necessarily respond to pollutant levels that would adversely affect some members.

Species vary in the way they take up, metabolize, eliminate, and accumulate elements. Species also vary in the way they respond to different elemental forms. For example, As<sup>3+</sup> is generally thought to be more toxic to plants than As<sup>5+</sup>. The values presented here do not make such distinctions nor could they be made based on the review literature.

Finally, the response of species and individuals depends upon a number of uncontrolled variables. Changes in these variables might alter the sensitivity of the plant. These variables include: age (stage of development), health and vigor, season of year, temperature, light intensity, soil type, moisture content of soil, pH of soil, humidity, wind speed, and the presence of other elements.

## 3.2.2 Screening Concentrations for Ambient Exposures

Table 3.1 presents the suggested screening values for seven gaseous pollutants. These values represent the minimum concentrations at which adverse growth effects or tissue injury in exposed vegetation were reported in the literature. Data for some other gases could not be included because the critical specification of averaging time was missing. Where information was available, separate values are given for sensitive, intermediate, and resistant plants. Species belonging to each of these groupings are given in Appendix B for SO<sub>2</sub>, NO<sub>2</sub>, and ozone. Figure 3.1 displays graphically the variation in experimental determinations of the minimum SO<sub>2</sub> concentration at which effects occur. Figure 3.2 presents a similar display for NO<sub>2</sub>. For both pollutants there is reasonable but not perfect agreement between the graphical data and the screening concentrations recommended in Table 3.1. The use of the data from the table rather than interpolation from the curves is

Table 3.1 Screening Concentrations for Exposure to Ambient Air Concentrations<sup>a</sup>,<sup>b</sup>

		Minis	num Reported Level	(ppmv) <sup>c</sup>	
		<b>V</b>	egetation Sensitiv	rity	
Pollutant	Averaging Time	Sensitive <sup>d</sup>	Intermediate	Resistant	Reference
SO <sub>2</sub>	1 hr	.35(917)	•	•	14
_	3 hrs	.30(786)	.80(2096)	5.0(13100)	16
	l yr	<del></del>	<del></del> .007(18) <del></del>	<del></del>	17
03 <sup>e</sup>	l hr	.20(392)	.35(686)	.55(1078)	18
•	4 hrs	.10(196)	.15(294)	.35(686)	18
	8 hrs	.06(118)	.15(294)	.30(588)	18
NO <sub>2</sub>	4 hrs	2.0(3760)	5.0(9400)	9.0(16920)	19
_	8 hrs	2.0(3760)	4.0(7520)	8.0(15040)	19
	l mo	<del></del>	<b>—</b> .30(564) <del> </del>		£
	l y <del>r</del>		<del></del> .0510(94-188)	<del></del>	20
COS	l wk	1000 (1,800,000)	•	10,000 (18,000,000)	21
H <sub>2</sub> s	4 hrs	20.0-60.0 (28,000-84,000)	-	400 (560,000)	22
Ethylene <sup>h</sup>	3-4 hrs		—.04(47) <del>———</del>		24
<b>,</b>	24 hrs	<del></del>	—.001(1.2) <del>——</del>		25
Fluorine	10 days	<del></del>	<b>(0.5-10)</b>		26
Beryllium <sup>i</sup>	l mo		<del></del> (0.01)	<del></del>	27
Leadj	3 mo		<del>(1.5)</del>		28

All values except beryllium and lead refer to effects on vegetation.

bMinimum reported levels at which visible damage or growth effects to vegetation may occur.

CValues in parentheses are  $\mu g/m^3$  at 20°C and 1 atm.

dThese values should be used in the screening procedure unless it is known that only intermediate or resistant plants will be affected.

<sup>&</sup>lt;sup>e</sup>The values for 20% injury are reported here, since they correspond closely with other values in the literature.

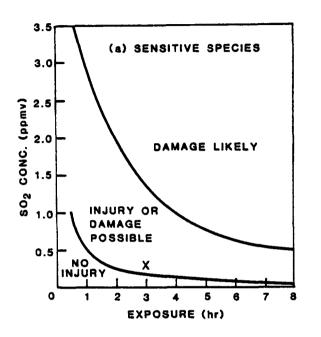
fBased on generalization of results of a number of studies.

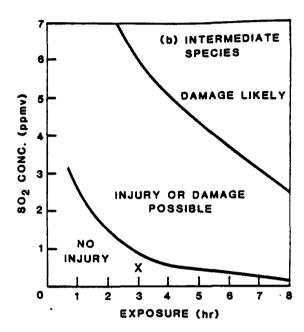
Skeversible decreases in photosynthetic rate have been shown to occur at significantly lower levels but effects on growth have not been demonstrated.

hEthylene " ... is the only hydrocarbon that should have adverse effects on vegetation at ambient concentration of 1 ppm or less." (Ref. 23).

ineshap value to protect public health. Very toxic to humans and presumably to some animals also.

JNAAQS value to protect public health.





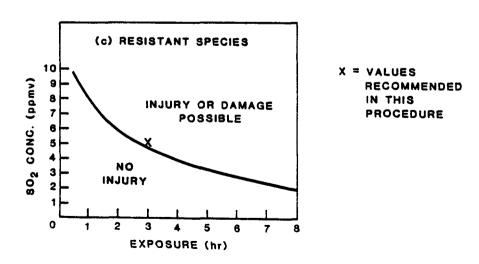


Fig. 3.1 SO<sub>2</sub> Dose-Injury Curves for Plant Species (From Ref. 8 as adapted from Ref. 29)

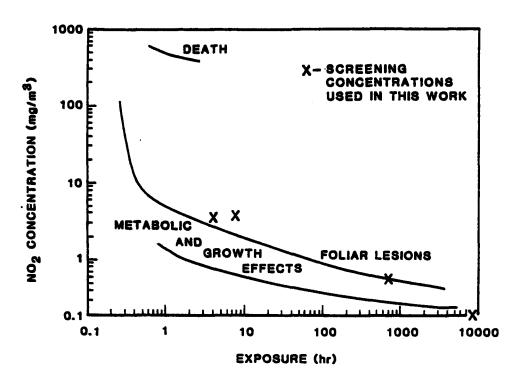


Fig. 3.2 NO<sub>2</sub> Dose-Injury Curves for Plant Species (From Ref. 8)

recommended, since the curves are based on attempts to fit theoretical doseresponse curves to experimental data whereas the tabulated screening concentrations are based directly on experimental results.

Several points are worth noting about the chosen screening concentrations. First, the significant variation between the values for the various sensitivity groupings should be noted. With this large variation it appears unlikely that use of any values but those for sensitive vegetation could be justified in a screening procedure, given the large number of species for which information is not available.

Second, the tabulated concentrations should be compared to NAAQS, PSD increments, and likely ambient concentrations. Table 3.2 summarizes these comparisons for the cases where they can be made. For pollutant/averaging times not tabulated, either no corresponding NAAQS or PSD increment exists or it appears that the screening concentration could be exceeded under certain circumstances. For the criteria pollutants, the NAAQS appear to protect against vegetative damage except possibly for 3-hr and annual SO<sub>2</sub> exposures. For the 3-hr exposure, the screening concentration exceeds the applicable PSD

Table 3.2 Screening Concentrations of Gaseous Pollutants
Compared to Ambient Criteria

	Averaging	Veg	etation Sensitiv	vity
Pollutant	Time	Sensitive	Intermediate	Resistant
so <sub>2</sub>	3 hr	< NAAQS <sup>a</sup> > PSD <sup>b</sup>	> NAAQSa > PSDb < NAAQSd	c c
	1 yr		> PSD Ie	
03	1 hr	> NAAQSf	> NAAQSf	> naaqsf
NO <sub>2</sub>	4 hr	-	c	C
	8 hr	-	C	c
	1 yr		—— ≈ NAAQSS——	
CO	l wk	c	-	c

 $a_{SO_2}$  3-hr NAAQS = .50ppmv (1300  $\mu g/m^3$ ).

increments and for the annual exposure, it exceeds the Class I increment. However, the screening concentration should be compared to the total SO<sub>2</sub> concentration including background whereas the PSD increment does not include background. Thus, a source could cause an SO<sub>2</sub> concentration less than the increment while the total SO<sub>2</sub> concentration (source plus background) could exceed the screening concentration. With the exception of the following it appears that possible adverse impacts to vegetation resulting from direct exposure to ambient concentrations of criteria pollutants are already covered by existing programs for NAAQS attainment:

- SO<sub>2</sub> exposures at 1 hour, 3 hours, and 1 year,
- Ozone exposures at 4 and 8 hours.

 $<sup>^{</sup>b}$ SO<sub>2</sub> 3-hr PSD increments ( $\mu g/m^{3}$ ) = 25(Class I), 512(Class II), 700(Class III), 325(Class I variance). These values do not include background.

<sup>&</sup>lt;sup>c</sup>Screening concentration unlikely to be reached under ambient conditions.

 $dSO_2$  annual NAAQS = .03 ppmv (80  $\mu g/m^3$ ).

 $<sup>^{</sup>e}$ SO<sub>2</sub> annual PSD increments ( $\mu$ g/m<sup>3</sup>) = 2(Class I), 20(Class II), 40(Class III), 20(Class I variance). These values do not include background.

 $f_{03}$  1-hr NAAQS = 0.12 ppmv (235  $\mu g/m^3$ ).

 $SNO_2$  annual NAAQS = 0.05ppmv (100  $\mu$ g/m<sup>3</sup>).

- NO<sub>2</sub> exposures of sensitive species at 4 and 8 hours, and
- Long-term NO2 exposures at 1 month and 1 year.

This observation does not preclude doing a review for impacts on plants, particularly where the minimum values at which effects have been reported are close to being exceeded. It does, however, indicate that the vegetative impact review can be done along with the review for NAAQS or PSD increments. Even in cases where review for NAAQS and PSD increments covers exposures to plants, there may still be the necessity of dealing with trace metal exposures through deposition in the soil or through concentration in plant tissues.

## 3.2.3 Synergisms

Only a very limited amount of information was available in the review literature consulted regarding synergisms. Three indications of synergism were found:

- SO<sub>2</sub> and NO<sub>2</sub>,
- SO<sub>2</sub> and O<sub>3</sub>, and
- SO<sub>2</sub>, O<sub>3</sub>, and NO<sub>2</sub>.

Table 3.3 presents values which could be used as screening concentrations based on the most restrictive values in the references. Where averaging times allow comparison, the screening concentrations for single pollutants in Table 3.1 are greater than the screening concentrations for mixed pollutants in Table 3.3. Given the problems with the data discussed in Secs. 3.1 and 3.2.1, this comparison should not be interpreted as clear evidence of synergism. An additional caution is also in order. Mixtures of gases may act synergistically on some species and antagonistically on others (see, for example, Ref. 18). Thus, the tabulated values should be used to indicate situations where the FLMs should be alerted so that the situation may be evaluated by them. There may be additional synergisms which are not noted in Table 3.3 but which could be added to the table and incorporated in the screening procedure at a later date.

Table 3.3 Synergisms of Gaseous Pollutants (Plants)<sup>a</sup>

Pollutants	Concentrations (ppmv)	Exposure	Reference
so <sub>2</sub>	.05	l hr	30
NO <sub>2</sub>	.05		
so <sub>2</sub> b	.30	l hr	31
03 .	.10		
	.05	4 hr	32
so <sub>2</sub> <sup>b</sup> o <sub>3</sub>	.05		
s0 <sub>2</sub>	.14	6 hr/day	33
03	.05	for 28	
NO <sub>2</sub>	.10	days	

<sup>&</sup>lt;sup>a</sup>The same criteria were used in selecting these values from Ref. 15 as were used in developing Table 3.1.

## 3.2.4 Screening Concentrations for Soil and Plant Tissue Exposures

Table 3.4 presents suggested screening concentrations for trace elements found to adversely affect plants. Two types of data are presented. gives a concentration which when present in the soil has been found harmful to The other gives a concentration found to be present in the tissues of plants which had been harmed. In considering these values, it should be remembered that most elements and compounds are not deleterious until they have been complexed in the soil and become suitable for uptake by plants. addition, many soil characteristics such as pH, composition (sand, clay, loam, organic matter, etc.), moisture content, and cation exchange capacity affect the amount of trace elements available for uptake. In developing the tabulated values, only data taken with the plants growing in soil were considered. Data developed in experiments in which plants were grown in aqueous nutrient solutions were ignored. Conditions of nutrient solution culture are likely to be sufficiently different from natural conditions as to render the results of the experiments misleading for the purposes of this work.

As with the ambient screening concentrations for gases, a great deal of variation is exhibited by the data as shown in Fig. 3.3. For comparison

bAntagonism, as well as synergism, has been reported for mixutes of SO2 and O3 (Ref. 18).

Table 3.4 Screening Concentrations for Exposure of Vegetation to Pollutant Concentrations in Soil and Tissue

Minim	um Report	ed Level (	pmw)
•	Polluta	nt Source	
Pollutant	Soil	Tissue	Reference
Arsenic	3	0.25	9
Boron	0.5	11	9
Cadmium	2.5	3	9
Chromium	8.4	1	9,35
Cobalt <sup>a</sup>	-	19	9
Copper	40	0.73	9
Fluoridea	400	310	9
Leada	1000	126	9
Manganese	2.5	400	9,36
Mercury	455		9
Nickel	500	60	9
Seleniuma	13	100	9,37
Vanadium	2.5	•••	38
Zinc	-	300	9

<sup>a</sup>Tissue concentrations may affect animals before affecting plants. Compare to toxic levels for animals in Table 3.7.

purposes, this figure includes results based on experiments in nutrient solutions and also shows the values chosen for screening concentrations in this work.

No standards or PSD increments currently apply to these trace elements so no comparisons with other review criteria can be made. It should be noted, however, that the heavy metals listed in Table 3.4 are emitted as particles and become TSP in the atmosphere. To the extent that they contribute to TSP levels, the NAAQS and PSD increments would apply to these trace elements. The connection between such ambient levels and the screening concentrations for soils and tissues is discussed in Sec. 5.

#### 3.3 SOILS

In contrast to the amount of published information on the effects of atmospheric pollutants on plants and animals, very little has been reported on their effects on soils. Research on trace elements in soils, often the same

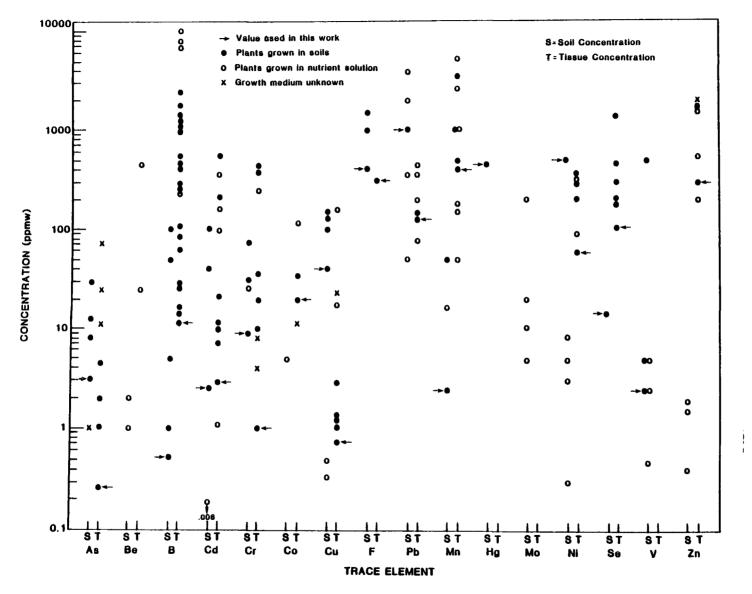


Fig. 3.3 Concentrations of Some Trace Elements Toxic to Terrestrial Plants

elements as atmospheric pollutants, has been directed to notable deficiencies or excesses that limit agricultural crop production. When the amount of an atmospheric pollutant entering a soil system is sufficiently small, the natural ecosystem can adapt to these small changes in much the same way as the ecosystem adapts to the natural weathering processes that occur in all soils. Cultural practices (e.g., liming, fertilization, use of insecticides and herbicides) add elements and modify a soil system more than a small amount of deposited atmospheric pollutant can. The secondary effects of the pollutant appear to impact the soil system more adversely than the addition of the pollutant itself to the soil. For instance, damaging or killing vegetative cover could lead to increased solar radiation, increased soil temperatures, and moisture stress. Increased runoff and erosion add to the problem. indirect action of the pollutant, through changes to the stability of the system, thus may be more significant than the direct effects on soil invertebrates and soil microorganisms. However, the lack of long-term historical data on both the type and amount of atmospheric pollutants as well as the lack of baseline data on soils has made difficult the task of determining the effect of pollutants on soils by monitoring changes associated with exposure to pollutants. A limited number of studies have been carried out on trace element contamination of soils. 39,40 Plant and animal communities appear to be affected before noticeable accumulations occur in the soils. approach used here in which the soil acts as an intermediary in the transfer of deposited trace elements to plants appears reasonable as a first attempt at identifying the air quality related values associated with soils.

When viewing soils in this way it is important to know the endogenous or background concentrations of elements already in the soil of interest, for these endogenous levels may be available for plant uptake. There is, however, a wide variation in the normal concentrations of various trace elements as shown in Table 3.5.8 If extremes in the concentrations are considered, the range of endogenous concentrations becomes even larger (see Fig. 3.4).41 Both references show relatively good agreement on the normal ranges. The tabulated values also provide "average concentrations" which can be used when specific information about the concentrations of trace elements in the region of interest is not available. One of the difficulties with screening for impacts on plants and soils becomes apparent when the endogenous concentrations in Table 3.5 are compared with the screening concentrations for soils in

Table 3.4: the screening values are exceeded for some part of the listed range for nine out of the twelve elements for which screening concentration are given. Fluorine, lead, and mercury are the only elements whose screening values lie above the corresponding endogenous ranges. The default average soil concentration exceeds the screening concentration for boron, manganese, vanadium, and chromium and, for the first three of these four, the entire listed normal range exceeds the screening value. In interpreting this indication, it must be remembered that the screening concentration value represents the lowest value found in the

Table 3.5 Range of Endogenous Soil
Contentrations of
Selected Elementsa

Element	Range (ppmw)	Average Soil Concentration (ppmw)
Arsenic	0.1-40	6.0
Beryllium	1-40	6.0
Boron	2-100	10.0
Cadmium	0.01-7.0	0.06
Chromium	5-3000	100
Cobalt	1-40	8
Copper	2-100	20
Fluoride	30-300	200
Lead	2-100	10
Manganese	100-4000	850
Mercury	0.01-4.0(?)	-
Nickel	10-1000	40
Selenium	0.01-80	0.5
Vanadium	20-500	100
Zinc	10-300	50

aBased on Ref. 8.

review literature (see Fig. 3.3) and that not all plant species are as sensitive as the one upon which the value is based. As outlined in Sec. 3.2.1, there are many additional reasons why there is no inherent conflict between screening concentrations and endogenous concentrations above these values. The chief among these are probably the variation in sensitivity between individuals, the variation in sensitivity between species, and the fraction of the endogenous concentration really available for uptake by plants. It should be noted, however, that endogenous concentrations of some elements can make soils toxic to some species. Thus, certain tolerant plants can act as indicator species for the element tolerated; they will be among the species present in soils where the endogenous concentrations of that element exceed levels toxic to more sensitive species. 12

The problem associated with the amount of an element in the soil which is actually taken up into plant tissues can be handled in an approximate fashion by using a plant:soil concentration ratio. Table 3.6 provides two sets of concentration ratios (CR's). One set is recommended for use in this work; the other is based on nonstandard methods using solution cultures

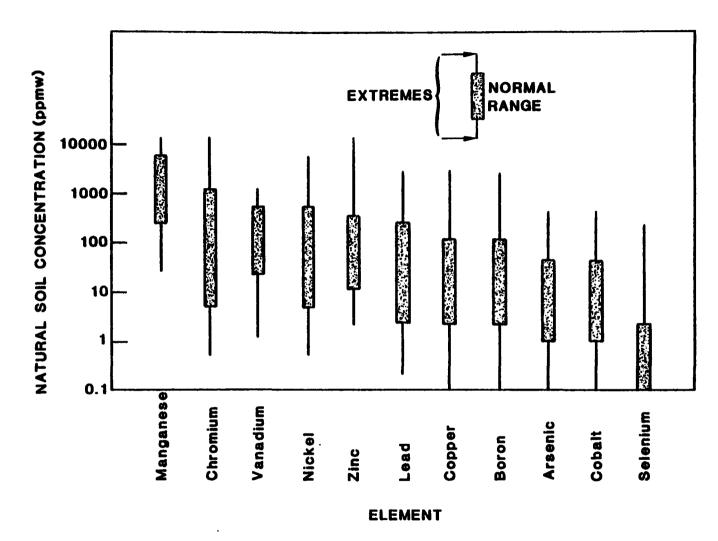


Fig. 3.4 Range of Endogenous Concentrations of Trace Elements (From Ref. 41)

but is given to provide some feeling for the large uncertainties associated with this type of work. comparison set of concentration ratios could be used in the screening procedure presented here to provide very conservative estimates of potential impacts. Some elements (boron and cadmium) tend to be concentrated by plants (ratios > 1), that is, concentrations in plant tissues exceed those found in the soil whereas the concentrations of most of the listed elements tend to be less in plant tissue than in the surrounding soil. In any case, these CR's represent ratios of averages and thus may give results quite different from the true ratio

Table 3.6. Plant: Soil Concentration Ratios

	•	
Element	Recommended Value <sup>a</sup>	Comparative Value <sup>b</sup>
Arsenic	0.14	4.2
Boron	5.3	-
Cadmium	10.7	222
Chromium	0.02	250
Cobalt	0.11	87
Copper	0.47	1000
Fluoride	0.03	-
Lead	0.45	2
Manganese	0.066	3000
Mercury	0.02-0.5	26
Nickel	0.045	331
Selenium	1.0	4
Vanad ium	0.01	1
Zinc	0.64	40

<sup>&</sup>lt;sup>a</sup>Based on Ref. 8.

between plant and soil concentrations in a particular case. However, they appear to be the best means available for estimating uptakes of various elements from the soil.

#### 3.4 FAUNA

The screening concentrations presented here are based on data for terrestrial vertebrates. Data for aquatic species, including fish, were not examined in the literature reviewed. Also, effects on aquatic and terrestrial microorganisms are not considered here. Table 3.7 presents the screening concentration values based on data summarized in Refs. 8 and 9. The tabulated values represent the lowest dietary concentrations found to be harmful. Several factors limited the usefulness of the available data. Some harmful levels were given in terms of average concentrations in the affected animals. Unfortunately no equivalents of the plant:soil CR's were available to go from dietary concentrations to concentrations per unit body weight. In addition, all the data on ambient exposures failed to give averaging times thus rendering it unuseable in this screening procedure. Even for the data upon which

bBased on Ref. 12. Based on nonstandard methods involving solution cultures. See discussion in text.

Table 3.7 is based, there were no indications as to how long the element needed to be ingested in the given concentration before causing the harmful effect. Comparison of the screening concentrations for animal effects (Table 3.7) with the values for plant tissue concentrations (Table 3.4) shows that the values for animals generally exceed those for plant tissue concentrations. However, for cobalt, fluoride, lead, and selenium, it appears that plants could accumulate concentrations that would be toxic to some animals before the plants themselves were harmed.

Table 3.7. Dietary Trace-Element Concentrations Toxic to Animals<sup>a</sup>

Trace Element	Dietary Concentration (ppmw)
Arsenicb	3
Cadmium <sup>b</sup>	15
Cobalt	1-3
Copperb	20-30
Fluoride	. 100-300
Lead	80-150
Manganeseb	500-5000
Nickel <sup>b</sup>	1000
Selenium	5-30
Vanad ium	10-500
Zinc	500-1000

<sup>&</sup>lt;sup>a</sup>Based on Ref. 8.

For beryllium and lead, data on ambient air exposures were available in terms of the NESHAP and NAAQS, respectively (see Table 3.1). These values relate to human exposures. Without other indications these same levels have been assumed to be potentially hazardous to at least some animals as well.

bTissue concentrations in plants may affect plants before affecting animals. Compare to plant screening concentrations in Table 3.4.

## 4 TRACE ELEMENT AIR QUALITY DATA

EPA's Storage and Retrieval of Aerometric Data (SAROAD) system was used as a data base to develop air quality information for trace elements. information was intended to serve primarily as an aid in estimating background concentrations so minimum concentrations were included. A secondary purpose of the information was to identify locations where high concentrations already For this purpose, maximum concentrations were included. of available data for all the pollutants discussed here with estimates for all relevant averaging times would not have been feasible so the data search was limited to trace elements including lead. It was also felt that more complete data for the gaseous criteria pollutants would be available locally than could be found in SAROAD. On the other hand, many localities probably lack estimates of trace element concentrations. Since only annual averages are used in screening for trace element impacts, the data search emphasized annual average data. Maximum and minimum short-term observations have been included in the data compilations for informational purposes.

In order to improve coverage, data for 1975-77 inclusive were used. Many locations had data for only one of the three years. As expected, all the data were based on high volume sampler data with 24-hour averaging times. It was also frequently the case that insufficient data was available to allow the calculation of a valid annual average. The available data is presented in Appendix C. No data was found for mercury, boron, cobalt, copper, and nickel. The data is presented by state and county for each pollutant. As can be seen from the tables, the spatial coverage is poor. For counties with data, only the minimum and maximum annual averages from all reporting stations are given. With multiple stations, it is unlikely that both values come from the same location.

In order to avoid possible misinterpretation of the data, it should be kept in mind that SAROAD routinely stores values below the limit of detectability as one-half the minimum detectable limit. In some cases, this will be the value which is listed as the minimum observation. These situations are usually fairly obvious, since the same minimum value will be recorded at a large number of stations.

#### 5 SCREENING PROCEDURE

#### 5.1 METHODOLOGY

## 5.1.1 Description

A simplified view of the pathways between sources and receptors is presented in Fig. 5.1. This simple view is used here as the basis for screening a source for potential adverse impacts on plants, soils, and animals. Emissions from the source are assumed to disperse in the atmosphere and add to whatever local background concentrations might exist to provide an estimate of the maximum ambient concentration for the averaging times of interest. These ambient concentrations may act along four different pathways. The first two are routes in which the ambient concentrations affect animals or plants directly without any intervening mechanisms. In the third, animals can ingest substances deposited on plants before the substances have been washed off by rain or blown off onto the soil. ingestion is a critical pathway. Appendix D provides a referenced discussion of the literature related to toxicity resulting from this pathway and the potential for harm to animals exists whenever heavy metals are deposited on materials which they ingest. Some start on dealing with this issue was made here in terms of estimating the amount of deposited material but a complete methodology was not developed. However, reviewers should be aware of this potentially critical pathway and the material in Appendix D may be useful in flagging critical situations. In the fourth, a certain amount of the dispersed material is deposited on the soil. As noted in Sec. 3, only the deposition of trace elements is considered here. The deposited trace elements as well as any endogenous concentration of the element are then available for uptake by plants in quantities which may be toxic to the plants themselves or to animals which feed upon the plants.

It is important to realize that this simplified picture leaves out many potentially important pathways and natural processes. For example, there is no provision for the uptake and concentration of substances by plants directly from the air; all such concentration is assumed to be through the soil with uptake by plant roots. No account is taken of removal of deposited substances from the soil by runoff, leaching, or erosion and the

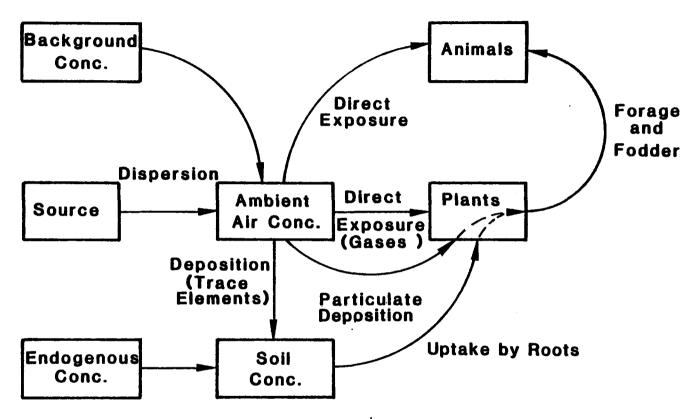


Fig. 5.1 Pollutant Pathways

subsequent deposition of such substances in bodies of water. Also, no account is taken of deposition directly from the air into water. Finally, the effects on animals of ingesting contaminated water have not been addressed.

Screening for a particular source is accomplished in a series of steps. Steps 1 and 2 apply to airborne pollutants; steps 1 and 3-7 apply for trace metals where deposition must be taken into account. Step 8 provides an alternative where modeling results for the source are unavailable.

- 1. Estimate the maximum ambient concentration for averaging times appropriate to the screening concentrations for pollutants emitted by the source and including any background concentrations.
- 2. For exposures to airborne pollutants, check the maxima from Step 1 against the corresponding screening concentrations in Table 3.1 or against the corresponding NAAQS, NESHAP or PSD increments, whichever applicable standard is most restrictive. In addition, the possibility of synergisms should be considered.
- 3. For trace metals, calculate the concentration deposited in the soil from the maximum annual average concentration assuming that all deposited material is soluable and available for uptake by plants.
- 4. Compare the increase in concentration in the soil to the existing endogenous concentration using the average values in Table 3.5 when local data is unavailable. (This provides a supportive indicator, not a primary decision parameter.)
- 5. Calculate the amount of trace element potentially taken up by plants using the CR's in Table 3.6.
- 6. Compare the concentrations from Steps 3 and 5 with the corresponding screening concentrations in Tables 3.4 and 3.7.
- 7. Reevaluate the results of the comparisons in Steps 4 and 6 using estimated solubilities of elements in the soil to provide supportive indications, recognizing that actual solubilities may vary significantly from the estimated values.
- 8. If modeling results are unavailable, the significance levels for emissions developed in Sec. 5.2 may be used to screen the source.

The discussion in Sec. 5.2 also provides an example of the application of the screening procedure. This example develops the significant emission levels for one of the trace elements from an estimate of a source's maximum

annual average concentration. Table 5.1 summarizes these steps and indexes them to the relevant sections, tables, and equations in the text. Figure 5.2 provides a flowchart of the screening procedure showing the more commonly used tables and equations.

## 5.1.2 Estimating Maximum Concentrations (Step 1)

To estimate the maximum concentration, the maximum air quality impact of the new source must be estimated and added to an appropriate background concentration.

## 5.1.2.1 Air Quality Modeling

The first step in the screening procedure for air quality related values is to estimate the maximum ambient concentrations of pollutants emitted from the new source for appropriate averaging times. Table 5.2 gives the correspondence between pollutants and the averaging times to be considered for each. Two cases need to be considered. The first arises when the required source-specific concentration estimates are available and the second arises when they are not.

Concentration Estimates Available. When source-specific estimates made by an approved model are available they should be used directly in making the calculations and comparisons called for in Steps 2-7 of Table 5.1. Such a situation would be ideal but such estimates may frequently be unavailable, particularly during early discussions of a permit application.

<u>Concentration Estimates Unavailable</u>. When source-specific estimates of concentrations are unavailable or when they are lacking for some critical averaging times, there are two courses of action:

- Use of a screening technique for air quality impacts if the emission rates and stack parameters are available or
- Use of the significance levels for emissions presented in Sec. 5.2.

If stack parameters are available, some simple techniques of dispersion modeling can be used to screen the source for its air quality impact, remembering that only a screen and not a definitive demonstration is required. Reference 42 provides such techniques developed by EPA for use in new source

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Table 5.1 Steps in Screening Procedure

Step	Description	Applicable Text		
		Section	Tables	Equat ion
1	Estimate ambient maxima			
	<ul><li>Modeling</li><li>Background</li></ul>	5.1.2 5.1.2, Appendix C	- C.1-C.10	- -
2	Screen for direct exposure	5.1.3	3.1,3.3,5.3	_
3	Calculate deposited concentration of trace elements <sup>a</sup>	5.1.3	-	5.1
4	Calculate percentage increases over endogenous concentrations <sup>b</sup>	5.1.3	3.5	5.4
5	Calculate tissue concentrations in plants	5.1.3	3.6	5.5
6	Screen for potential adverse impacts of trace elements	5.1.3	3.4,3.7,5.5	-
7	Consider effects of trace element solubility <sup>b</sup>	5.1.3	3.4,3.7,5.4	5.7,5.8
8	Apply significance emission levels <sup>c</sup>	5.2	5.6,5.7	

<sup>&</sup>lt;sup>a</sup>Reviewers may want to review the information in Appendix D to assess the potential for harm to animals from directly ingesting deposited materials.

<sup>&</sup>lt;sup>b</sup>Supportive indication only, not primary decision parameter.

<sup>&</sup>lt;sup>c</sup>Used only when source-specific modeling results are not available.

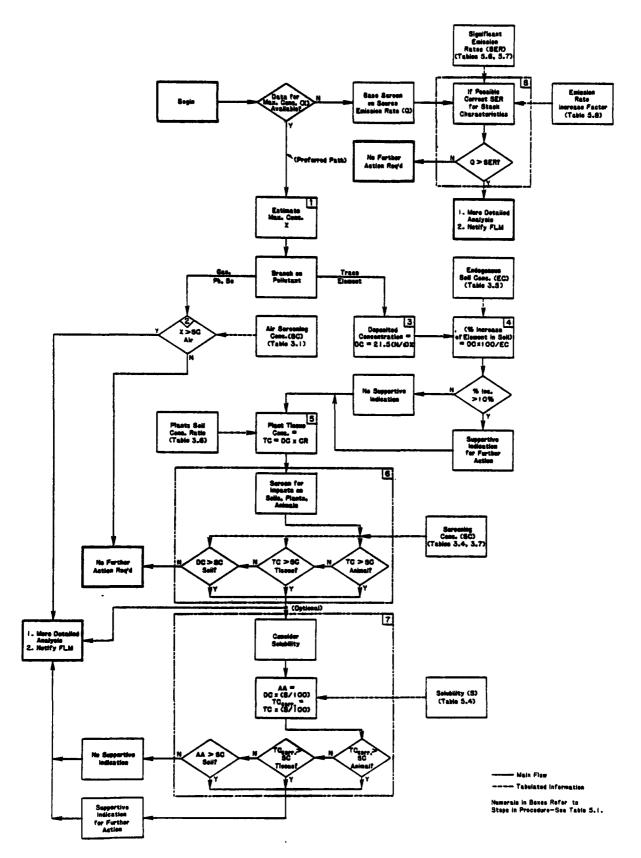


Fig. 5.2 Flowchart of Screening Procedure

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Table 5.2 Pollutants and Averaging Times

				Requ	ired Ave	eraging	Times			
Pollutant	l hr	3 hr	4 hr	8 hr	24 hr	1 wk	10 days	l mo	3 mo	l yr
so <sub>2</sub>	Х	Х			χa					Хр
NO <sub>2</sub>			x	x				X		$\mathbf{x}_{\mathbf{p}}$
со	χа			χa		X				
H <sub>2</sub> S	Хª		x							
Ethylene		x			x					
Fluoride					χa		X			
Ве								χa		
Pb									$\mathbf{x}^{\mathbf{b}}$	XC
Trace Elementsd										χe

<sup>&</sup>lt;sup>a</sup>For comparison with criteria not necessarily related to impacts on plants, animals, or soils (NAAQS, NESHAP's, PSD increments).

bApplies to both impacts on plants, animals, soils and other criteria.

<sup>&</sup>lt;sup>c</sup>Also included in trace element analysis.

dTrace elements: As, B, Cd, Cr, Co, Cu, F (as fluoride), Pb, Mn, Hg, Ni, Se, V, Zn.

eRequired for use in estimating amount of deposition.

review. These methods were used to develop EPA's significance levels for emissions 42 published as part of the proposed PSD regulations. 43,44

As an alternative, the procedure used in Ref. 45 to estimate air quality impacts can be used as presented in Appendix A. Some expansion of the original procedure was required to cover the range of averaging times needed for this screening procedure. The equations presented in Appendix A are suitable for hand calculation or the development of a simple computer code. The significance levels presented in Sec. 5.2 are based on this procedure.

# 5.1.2.2 Background Concentrations

The estimation of background concentrations is one of the perennially difficult problems of air quality analysis. Development of new approaches was beyond the scope of this work. The analyst should consult Ref. 46 for guidance on this subject. No attempt was made here to develop information for the gaseous criteria pollutants. For these gases, it was felt that local records would be likely to provide more timely and complete information. In addition, the sheer volume of data available precluded its inclusion in this procedure. No attempt was made to develop background estimates for other than annual averaging times.

For the 14 trace elements (including lead), EPA's SAROAD files were searched as described in Sec. 4. No information was found for mercury, boron, cobalt, copper, and nickel. The tables in Appendix C summarize the information found by state and county. To estimate a background value, the concentrations in the county of interest or nearby counties should be used and the minimum geometric mean picked. This minimum can then be added to the estimated maximum annual concentration from the source being screened. Values of the minimum geometric mean from other areas should be compared with the value chosen. It is possible that some of the tabulated minima may be too high to represent background levels because the monitor providing the data is impacted by a large source and thus is not representative of general background conditions.

It will not be possible to estimate background levels by this method for many locations. In such a situation, the minimum geometric mean may

be selected from among those tabulated in Appendix C and used in a sensitivity analysis to determine if the addition of a background level is likely to raise the predicted concentration above the screening concentration. If it does, then a determination of background will be necessary to allow a clear determination of the source's potential to cause adverse impacts due to trace element deposition.

# 5.1.3 Screening and Deposition (Steps 2-7)

Screening for Direct Impacts (Step 2). This screen applies to the pollutants listed in Table 3.1 for which data was available on direct impacts of airborne concentrations on plants and animals: SO<sub>2</sub>, NO<sub>2</sub>, CO, H<sub>2</sub>S, ethylene, flourides, Be, and Pb. After the maximum concentrations both with and without background have been calculated, screening is simple. The appropriate maxima are compared to the values given in Table 5.3. Values in excess of the screening concentrations indicate that additional detailed review is required and that the appropriate FLM should be notified. The possibility of synergisms should also be checked at this point. Consideration should be given to the synergisms listed in Table 3.3 but no screen on the values listed there is recommended here. Rather, the information could be used to alert the appropriate FLM to the possibility of a problem arising from synergisms.

Also included in Table 5.3 are the values used in reviewing new sources under other criteria. The value expected to be controlling for each pollutant has been circled in the table under the following assumptions:

- No background,
- Long averaging times result in lower concentrations than short averaging times, and
- For short averaging times, the concentration is proportional to averaging time raised to the power -0.17.

This observation is made only to give some feeling for what might be expected. It is possible, for example, for a new SO<sub>2</sub> source in a Class III area to be controlled by the 700  $\mu g/m^3$  PSD increment and still need to do a review for plant, soil, and animal impacts if 3-hour background levels are high enough to make the predicted ambient concentration likely to exceed 786  $\mu g/m^3$ . Completion of Step 2 would complete the screening for direct impacts from airborne pollutants.

Table 5.3 Ambient Screening Concentrations

	,				<del></del>				Amb i en	t Concent	ration (µg/	<b>"</b> 3)					-
						<del></del>			Pol lut	ent and A	veraging Tim	le <sup>4</sup>					
Screening	802				NO <sub>2</sub>			, co	H <sub>2</sub> 8	Ethylene		Pluoride	Beryllium	Lead			
Criterion	1	3	24	A	4	8	Н	A	1	8	W	4	3	24	240	м 3н	3M
AQRV Screening Concentrationb	917	786	_	18	3,760 <sup>b</sup>	3760b	564	100	-	-	1,800,000 <sup>b</sup>	28,000b	47	1.2	0.5	.01	1.5
NAAQSC, d	-	1,300	365	80	-	-	-	100	40,000	10,000	-	-	-	-	-	-	1.5
PSD Increment	-	25	<u>(3</u>	<b>)</b> 2	-	_	_	-	-	_	_	~	_	-	-	-	-
Ile't		512	91	20	-	-	-	-	-	-	-	-	-	-	-	-	-
III <sup>e</sup> ,f Variance <sup>e</sup> ,8	·-	700 325	182 91	40 20	-	-	-	-	-	-	-	-	-	-	-	-	-
NESHAPE,h	-	-	-	-	-	_	-	-	_	-	-	-	-	-	-	.01	-

Note: Circled values expected to be controlling; see text.

<sup>a</sup>Numerals: hours

W: 1 week

M: 1 month

A: Annual

bAmbient concentrations this high are unlikely.

C40 CFR 50.

dBased on maximum impact of source plus background.

eRef. 1.

fBased on maximum impact of source alone.

8Includes the source together with all other sources.

h40 CFR 61.

Calculating Deposited Soil Concentrations (Step 3). Deposition of trace elements is a long-term process extending over the lifetime of the source. The simple procedure used here depends upon an estimate of the maximum annual average concentration from the source as corrected by the addition of a background concentration if known. Reviewers may also want to review Appendix D at this point to assess the potential for harm to animals from direct ingestion of deposited heavy metals (see Sec 5.1.1). The following equation can be used to estimate the maximum concentration in the soil:

$$DC(ppmw) = 21.5 (N/d)\chi$$
 (5.1)

where:

DC = deposited concentration (ppmw),

N = expected lifetime of source (yr),

d = depth of soil through which deposited material
 is distributed (cm), and

 $X = maximum annual average ambient concentration from the source (<math>ug/m^3$ ).

The value generally recommended for d is 3 cm.8,9,12 Some work  $^{13}$  has assumed 20 cm for d, but the more conservative value of 3 should be adopted for use in this screening procedure unless site-specific data indicate that greater penetrations of deposited substances are more representative of local conditions. It should also be noted that an estimate of the source's lifetime must be made in order to use Eq. 5.1. In the absence of contrary indications, a value of N = 40 years should provide a reasonable and generally conservative estimate of source lifetimes based on lifetimes equal to twice the time allowed by the Internal Revenue Service for equipment depreciation.  $^{45}$ ,  $^{47}$  If the source is tied to a resource, the estimated resource lifetime might be used instead of 40 years. For example, a mine-mouth power plant might have a lifetime of N = 100 years based on the life expectancy of the mine or a gas plant might have a lifetime N = 15 years, the expected useful life of the gas field.

Equation 5.1 is simply derived. Consider a volume of soil 1  $m^2$  in area and d cm deep at the location of the source's annual maximum. The weight of material deposited on this area of 1  $m^2$  can be calculated as:

$$\begin{pmatrix}
\text{Weight} \\
\text{Deposited}
\end{pmatrix} = \begin{pmatrix}
\text{Ambient} \\
\text{Concentration}
\end{pmatrix} \times \begin{pmatrix}
\text{Deposition} \\
\text{Velocity}
\end{pmatrix} \times (1 \text{ m}^2) \times (\text{Time}).$$
(5.2)

The weight of the soil in the volume of interest is

$$\begin{pmatrix}
\text{Weight} \\
\text{of soil}
\end{pmatrix} = \begin{pmatrix}
\text{Volume} \\
\text{of soil}
\end{pmatrix} \times \begin{pmatrix}
\text{Bulk Density} \\
\text{of soil}
\end{pmatrix}$$

$$= (1 \text{ m}^2) \times (d) \times \begin{pmatrix}
\text{Bulk Density} \\
\text{of soil}
\end{pmatrix}.$$
(5.3)

Then the ratio of the weight deposited to the weight of the soil can be used to find the concentration of the deposited material by weight in the soil. Soil densities range from 1-2 gm/cm<sup>3</sup> and a value of 1.47 g/cm<sup>3</sup> is assumed here as a good average value.<sup>12</sup> If an average value of 1 cm/sec is assumed for the deposition velocity, Eqs. 5.2 and 5.3 can be combined to give

DC = (Weight deposited)/(Weight of soil)

$$= \chi \left(\frac{\mu g}{m^3}\right) \times 1 m^2 \times 1 \frac{cm}{sec} \times \frac{1 m}{.01 cm} \times 3.1558 \times 10^7 \left(\frac{sec}{yr}\right)$$

$$\times N(yr) \times \frac{1}{d(cm)} \times \frac{.01 cm}{m} \times \frac{1}{1 m^2} \times \frac{1}{1.47} \left(\frac{cm^3}{g}\right) \times \frac{1}{10^6} \left(\frac{m^3}{cm^3}\right)$$

$$= 21.5 (N/d) \times \left(\frac{\mu g}{g}\right)$$

$$= 21.5 (N/d) \times (ppmw)$$

where conversion factors have been used as appropriate to give consistent units. This result is simply Eq. 5.1. The principal assumptions in this derivation are:

- Deposition velocity of 1 cm/sec,
- Average bulk density of soil = 1.47 gm/cm<sup>3</sup>,
- Uniform distribution of deposited material throughout the soil volume, and
- All deposited material is retained by the soil, that is, no leaching, surface runoff, or erosion.

Calculate Increase over Endogenous Soil Concentration (Step 4). The purpose of this simple calculation is to provide a supportive indication

for the primary screen for deposition to be carried out in Step 6. As suggested in Ref. 13, an increase over the endogenous concentration of more than 10% over the lifetime of the source could be taken as a possible cause for concern. The percentage increase is simply calculated from

where the deposited concentration (DC) was calculated in Step 3. The average endogenous concentrations from Table 3.5 can be used but data for the area of interest is preferable given the wide range in natural concentrations.

It is not recommended at this time that a source be flagged for further actions based solely on the results of this calculation. The results of the screens in Step 6 are appropriate for that purpose. However, an indicated increase of more than 10% in this step would increase the assurance with which a finding that additional action was necessary could be made.

Calculate Potential Concentrations in Plant Tissue (Step 5). Once the deposited concentration in the soil has been calculated using Eq. 5.1, straightforward application of the plant:soil concentration ratios in Table 3.6 can be used to estimate the concentration in aerial plant parts (tissue concentration)

[Tissue concentration (ppmw)] =
[Deposited concentration (ppmw)] x [Concentration ratio]

$$TC (ppmw) = DC (ppmw) \times CR$$
 (5.5)

using TC for tissue concentration and other symbols introduced earlier. Equation 5.5 requires an additional conservative assumption:

 All the deposited material is soluable and available for uptake by plants.

OT

This assumption is almost always violated in practice. Table 5.4 gives the solubilities of some trace elements based on extraction of these elements from endogenous concentrations in the soil. 13 Of course, the solubilities of exogenous deposited elements could differ markedly from these values as could the solubilities of endogenous concentrations in different soils. The solubility of a trace element in the soil depends upon many factors. Among these

Table 5.4 Solubilities of Endogenous
Trace Elements<sup>a</sup>,<sup>b</sup>

Element	Solubility (%)c	Emission Rate Increase Factord
Arsenic	9	11
Boron		
Cadmium	40	2.5
Chromium	0.004	25,000
Cobalt	0.4	250
Copper	0.1	1,000
Fluoride		**
Lead		
Manganese	37	2.7
Mercury	0.8	120
Nickel	0.1	1,000
Selenium	21	4.8
Vanad ium		***
Zinc	8	12

aBased on Ref. 13.

are chemical form, temperature, presence of other elements, selective uptake by plants, soil pH, and soil moisture content. The composition of the soil is also an important determinant of solubility, especially the presence of organic matter and clays which can bind trace elements. The point is that a significant portion of the exogenous concentration may be unavailable for uptake by plants, making Eq. 5.5 a conservative estimator.

Screen for Potential Adverse Impacts from Trace Elements (Step 6). At this point the screen for adverse impacts from the deposition of trace elements is straightforward. The process is similar to that used in Step 2, that is, the comparison of calculated concentrations to tabulated screening concentrations. In this step, however, three comparisons need to be made:

bUsed in Step 7.

COnly soluable fraction would be available for uptake by plants.

dUsed when Step 8 is required.

- 1. The deposited concentration (DC) is compared to the soil screening concentration in Table 3.4,
- The tissue concentration (TC) is compared to the tissue screening concentration in Table 3.4, and
- 3. The tissue concentration (TC) is compared to the dietary screening concentration for animals in Table 3.7.

A calculated concentration in excess of any one of the three screening concentrations is an indication that a more detailed evaluation may be required for the new source and/or that the FLM should be notified, since there are indications of potential adverse impacts to plant, soils, or animals. In making these three comparisons, the following additional assumptions have been made:

- All deposited forms of an element have the same toxicity,
- The feeding or grazing range of animals is limited to the area exposed to the estimated maximum annual concentration, and
- Most importantly, it is the exogenous incremental burden which should be compared with the screening concentration values, not the burden which would result from both the exogenous and endogenous concentrations.

This last assumption is critical and follows the procedure used in Refs. 12 and 13. The assumption is implicit in Eq. 5.5 where only the deposited concentration (DC) is used to calculate the tissue concentration (TC) and in the three screens as defined above.

The three screens can be compared to see which is the most restrictive. The screening value for concentrations in aerial plant tissues and for concentrations toxic to animals can be converted into equivalent soil concentration values by use of the plant:soil concentration ratios. The dietary concentration potentially toxic to animals can be thought of as the concentration in aerial plant parts that may be toxic to animals. Thus, Eq. 5.5 can be rearranged to give the equivalent deposited concentration (EDC) corresponding to a particular screening tissue concentration (STC):

where the STC is either the plant tissue screening concentration from Table 3.4 or the animal screening concentration from Table 3.7. In fact, Eq. 5.6 provides an alternative approach to the screening procedure that is equivalent

to the one presented here. Table 5.5 gives the equivalent deposited concentrations (EDCs) for the trace elements. Based on the CR's and assumptions used here, animals appear to be the critical receptor for cobalt, lead, and selenium while tissue concentrations in plants appear to be critical for arsenic, cadmium, copper, and zinc. For the remaining seven elements, the soil concentration appears to be critical. As long as the screening concentrations and concentration ratios given here are used, Table 5.5 can be used to reduce the number of comparisons required for a screen. For example, cadmium sources need only be screened against the single screening value for plant tissue concentrations, since this screening concentration is shown to be controlling in the table.

Table 5.5 Equivalent Exogenous Soil Screening Concentrations

	Equivalent	Deposited Concent	ration (ppmw)
Trace Element	Soil <sup>a</sup>	Plant Tissue <sup>b</sup>	An ima ls <sup>c</sup>
Arsenic	3	1.8d	21
Boron	0.5d	2.1	_
Cadmium	2.5	0.28d	1.4
Chromium	8.4d	50	***
Cobalt	-	170	9.1 <sup>d</sup>
Copper	40	1.6d	43
Fluoride	400 <sup>d</sup>	10,300	3,300
Lead	1000	280	180d
Manganese	2.5 <sup>d</sup>	6,100	7,600
Mercury	455d	-	-
Nickel	500 <sup>d</sup>	1,300	22,000
Selenium	13	100	5 <b>d</b>
Vanadium	2.5d	-	1,000
Zinc	-	470d	780

aSame as soil value in Table 3.4.

bEDC = (STC for plants from Table 3.4)/CR.

CEDC = (STC for animals from Table 3.7)/CR.

dControlling value.

Since acute fluoride poisoning in various species of cattle has been well documated, 48 it is surprising that animals do not appear to be critical for fluorides. This may be due to the omission of the critical pollutant pathway involving ingestion by animals of materials deposited on plants prior to these materials being washed off or blown off the plants and carried into the soil. The same indication could be given of course, if the screening concentration value for the effects of soil fluorides on plants were based upon a very sensitive species. Further detailed investigation and more data would be required to determine whether the latter explanation is true or whether there is a deficiency in the procedure outlined here. In either case, the fluoride example serves to illustrate the potential problems involved in screening for impacts on air quality related values.

Consider Effects of Solubilities (Step 7). The assumption that all deposited material is soluable and available for uptake by plants is unlikely ever to be met in practice. If a screen indicates that a further action is needed on a source because its emissions will cause a trace element screening concentration to be exceeded, an attempt may be made to look at the possible effect of reduced solubility on that indication by considering the solubility of the deposited material. This additional consideration should only be used as a supportive indicator; it can only increase confidence in the decision to take further action; it can never reverse such a decision based on the screens in Step 6. That is, the conservative assumption of 100% solubility should be used in making the decision for further action on the source.

If the solubility of a particular trace element is S%, the amount actually available for uptake (AA) by plants is

or

$$AA = DC \times (S/100).$$
 (5.7)

This value for AA should be compared with the soil screening concentrations in Table 3.4. An equation similar to Eq. 5.5 can now be written reflecting

the assumption that only the fraction AA of the deposited concentration is available for uptake.

$$TC_{COTT} = AA \times CR = DC \times (S/100) \times CR = TC \times (S/100)$$
 (5.8)

where TC<sub>COTT</sub> stands for the tissue concentration corrected for the solubility of the deposited material. The new values of TC<sub>COTT</sub> could be compared with the screening concentrations for plant tissues and animals given in Tables 3.4 and 3.7, respectively.

## 5.2 EXAMPLE SCREEN AND SIGNIFICANT EMISSION RATES

Section 5.2.1 illustrates the use of Steps 1-7 of the screening procedure through application to a source of nitrogen dioxide and arsenic. Whenever source-specific estimates of maximum concentrations are available or can be generated. Steps 1-7 should be used. Step 8 provides an alternative screening procedure based on the concept of significant emission rates (SER). Section 5.2.2 illustrates the derivation of the SER for arsenic from the results for the example source and describes the use of the SER's for screening. Use of the SER's precludes any consideration of the emission characteristics of the source other than emission rate. Local conditions including background also cannot be taken into account. Application of Steps 1-7 is the preferred procedure.

## 5.2.1 Example Screen

The example source is assumed to have a plume release height of 30 m (physical stack plus plume rise). It is assumed that the source is subject to PSD review and that it is desired to screen the source for arsenic and nitrogen dioxide among other pollutants. An emission rate of 1 T/yr of arsenic is assumed for this example and estimates of maximum concentrations of  $NO_2$  are available for 4-hour and 8-hour averaging times. Following Table 5.1 or Fig. 5.2, the first step in the procedure is to estimate maximum concentrations for the times listed in Table 5.2. For arsenic, these estimates need to be made. Using the simple modeling procedure outlined in Appendix A, the maximum annual average ground level concentration is found to be  $X = 0.1051 \ \mu g/m^3$ . Other appropriate models or techniques could also be used. If an insignificant background is assumed for the example, this result completes Step 1 of the screening procedure for arsenic. For  $NO_2$ , the

available results show maximum ground level concentrations (including background) of  $\chi_4 = 51 \ \mu g/m^3$  and  $\chi_8 = 45 \ \mu g/m^3$  for averaging times of 4 and 8 hours, respectively. (A little foresight will show that estimates need not be made for 1 mo and 1 yr.) These results complete Step 1.

Then in Step 2 of the screening procedure, these maximum concentrations for  $NO_2$  would be compared to the appropriate screening concentrations in Table 3.1 or Table 5.3. For  $NO_2$ , the screening concentration at both 4 and 8 hours is  $3760~\mu\text{g/m}^3$ . The estimated maxima are for below this value. No calculation need be done for the one month and annual averaging times, since the modeled 4- and 8-hour maxima are already below the corresponding screening concentrations. There would thus be no indication that a more detailed review would be required for  $NO_2$  impacts on plants, soils, and animals.

Since the screen also involves a trace element, the next step is Step 3. If a 10-year lifetime (N=10) is assumed and the recommended value of 3 cm is used for the depth of soil throughout which the deposited arsenic is mixed, Eq. 5.1 gives

DC = 21.5 (N/d) X

= 21.5 (10/3) x (.1051) = 7.53 ppmw as the concentration of arsenic in the soil.

Following with Step 4 and Eq. 5.4,

[% Increase] =  $7.53 \times 100/6 = 126$ %

where 6.0 ppmw has been used as the average endogenous soil concentration of arsenic from Table 3.5. Thus, there is a supportive indication that the source should receive further review if Step 6 shows the potential for adverse impacts because the source may increase concentrations of arsenic in the soil by more than 10%. In Step 5, the plant tissue concentration would be calculated from Eq. 5.5:

 $TC = DC \times CR = 7.53 \times 0.14 = 1.05 \text{ ppmw}.$ 

Next the screening comparisons are made in Step 6. The DC (=7.53 ppmw) exceeds the soil screening concentration of 3 ppmw for arsenic given in Table 3.4. Similarly, the TC (1.05 ppmw) exceeds the tissue screening concentration of 0.25 ppmw given in Table 3.4. The TC does not exceed the animal-related screening concentration of 3 ppmw given in Table 3.7. There are thus

two indications that this source might adversely affect plants and that further actions need to be taken.

To look at the possible effect of arsenic solubility on these indications, the calculations in Step 7 can be done. For arsenic, Table 5.4 gives a solubility of 9% to account for the limited solubility of arsenic compounds. Equations 5.7 and 5.8 give AA =  $7.53 \times .09 = 0.68$  ppmw and  $TC_{COTT} = 1.05 \times .09 = 0.0945$  ppmw. AA does not exceed the soil screening concentration of 3 ppmw and  $TC_{COTT}$  does not exceed the tissue screening concentrations for plants and animals, 0.25 ppmw and 3 ppmw, respectively. Thus, no supportive indication has been found but the original indication that additional detailed work is required on the source is not altered and it is known that solubility effects might be important.

# 5.2.2 Significant Emission Rates

Basic Levels. This subsection discusses the development of a significant emission rate (SER) for arsenic based on the generic source discussed in Sec. 5.2.1 with a release height of 30 m and an expected lifetime of 10 years. An SER is defined as the minimum emission rate which would cause the source's impact to just equal the screening concentration. That is,

For arsenic in soils and the example source,

$$SER(Soils) = [3/7.53] \times (1 T/yr) = 0.40 T/yr.$$

Arsenic emissions from this source in excess of 0.40 T/yr might be expected to cause a soil concentration in excess of the screening concentration. Similarly, significant emission rates based on plant tissues (TC = 1.05 ppmw) and animal ingestion (TC = 3 ppmw) can also be calculated:

SER(Tissue) = 
$$[0.25/1.05] \times (1 \text{ T/yr}) = 0.24 \text{ T/yr}$$
 and  
SER(Animals) =  $[3/1.05] \times (1 \text{ T/yr}) = 2.8 \text{ T/yr}$ .

Such significant emission rates were calculated assuming a 30 m release height as in Ref. 43, a 10-year source lifetime, and the air quality model presented

in Appendix A. For pollutants acting along the direct pathways, Table 5.6 presents the significant emission rates. Table 5.7 presents such rates for trace elements. When no modeling results or stack parameters such as are required by simple air quality screening procedures are available, the source's emission rates can be compared directly with those given in these two tables. As already noted in the discussion of Table 5.3, other criteria may be controlling particularly when background is considered. Still, the significant emission rates presented in Table 5.6 can be used to screen for potential adverse impacts to plants, animals, and soils. Other criteria may apply to different stages of the new source review process. When applying the significant emission rates in Table 5.7, only the smallest value need be considered for each pollutant. The values based on exceeding ten percent of the average endogenous soil concentration should again only be used as supportive indicators; the primary decision is based upon exceeding the values based on the criteria for soils, plant tissues, and animals.

The values tabulated in Table 5.7 assume a source lifetime of 10 years. Significant emission rates for other lifetimes for trace elements acting through the deposition pathway are easily calculated:

Thus, for example, if the lifetime of the arsenic source in the above example had been 40 years instead of 10 years, the associated significant emission rate based on the plant tissue screening concentration would have been changed from 0.24 T/yr to

$$(0.24) \times (10/40) = 0.06 \text{ T/yr}.$$

Solubility. As in Step 7, additional supportive indications can be sought by considering the effects of solubility. A corrected significant emission rate can be found from

Table 5.6 Significant Emission Rates for Direct Acting Pollutantsa

									Bignific	ant Emise	sion Rate (1	(/yr)					
		-							Pollut	ant and A	veraging Ti	meb				· · · · · · · · · · · · · · · · · · ·	
Screening		802			NO <sub>2</sub>				co		H <sub>2</sub> 8	Ethylene		Fluoride	Beryllium	Lead	
	3	24	A	4	8	H		1	8	W	4	3	24	240	H	3M	
AQRV Screening Concentration	160	170	_	171	840	950	3,200	950	-	-	760,000	6,400	10.0	0.36	0.23	0.057	11
NAAQS	-	290	110	760	-	-	-	950	7,000	2,500	-	-	-	-	-	-	11
PSD Increment I	-	5.3	1.5	19	_	-	_	_	_	-	-	-	_	-	-	-	_
II	-	110	28	190	-	_	-	-	-	-	_	-	-	-	_	-	-
III	-	150	55	380	-	-	-	-	-		_	-	-	_	_	-	_
Variance	-	69	28	190	-	-	-	-	-	-	-	-	-	-	-	-	-
NESHAP	_	-	-	_	_	_	_	-	-	-		-	_	-	-	0.057	_

<sup>&</sup>lt;sup>a</sup>Based on 30 m release height and no background.

bNumerals: hours W: I week

M: 1 month A: Annual

Table 5.7 Significant Emission Rates for Trace Elements<sup>a</sup>

		Significant	Emission	Rate (T/yr)
			riterion	
Trace Element	Soils	Plant Tissue	Animals	10% of Endogenous Soil Concentration <sup>b</sup>
Arsenic	.40	. 24	2.8	.08
Boron	.067	.28	-	.13
Cadmium	.33	.037	.19	.00080
Chromium	1.1	6.7	-	1.3
Cobalt	-	23c	1.2	.11
Copper	5.3	.21	5.7	.27
Fluoride	53c	1400°	440 <sup>c</sup>	2.7
Lead	130d	37 <b>d</b>	24d	.13
Manganese	.33	810c	1000°	11 <sup>c</sup>
Mercury	61c	-	-	-
Nickel	67 <sup>c</sup>	170 <sup>c</sup>	3000°	.53
Selenium	1.7	13c	.67	.0067
Vanadium	.33	-	130°	1.3
Zinc	-	63c	100°	.67

aBased on a 30 m release height, no background, and a source lifetime of 10 years. For a lifetime of N years, divide the tabulated values by (N/10).

These emission rate increase factors are simply (100/S), the reciprocals of the solubilities in percent.

Other Stacks. Even though the stack parameters may not be known exactly, it may be known that the stack is hot or cold. Table 5.8 gives stack parameters for four stacks which might be useful if they are closer to the source's expected stack parameters than the 30 m release height assumed

bFor use as a supportive indicator only; based on a 10% increase over the average values in Table 3.5.

<sup>&</sup>lt;sup>C</sup>Exceeds the significant emission level for TSP of 10 T/yr established for PSD (Ref. 3).

dExceeds the significant emission level for lead of 1 T/yr established for PSD (Ref. 3).

Table 5.8. Summary of Representative Stacks

Stack	Height (m)	Temperature (°K)	Flow (m <sup>3</sup> /sec)	Emission Rate Increase Factor	
30 m release	30	293	0	1.00	
10 m cold	10	350	4	0.96	
l0 m hot	10	550	4	4.07	
30 m cold	30	350	4	3.43	
30 m hot	30	550	4	8.93	

in Tables 5.6 and 5.7. The volume flow rate of 4 m<sup>3</sup>/sec is felt to be conservative for major sources unless a large number of stacks are used. Also given in the table are emission rate increase factors for each model stack. A particular factor would be used to adjust the tabulated significant emission rates in Tables 5.6 and 5.7 to correspond more closely to concentrations expected from the proposed source:

# APPENDIX A

Estimates of Maximum
Ground Level Concentrations

#### APPENDIX A

## ESTIMATES OF MAXIMUM GROUND-LEVEL CONCENTRATIONS

This appendix develops the procedure used to estimate maximum ground-level concentrations (mglc's) from a single source for averaging times ranging from one hour to one year. The developments presented here follow the presentation in Ref. 45 which can be consulted for additional details. The procedure is useful for screening because the calculations can be done by hand or implemented in a simple computer program. The procedure accounts for stack parameters, plume rise, and meteorological conditions.

#### A.1 SHORT-TERM ESTIMATES

The familiar Gaussian plume model is the basis for estimating shortterm ground level concentrations.<sup>49</sup> According to this model the plume centerline concentration is given by

$$\chi(x) = \frac{Q \times 10^6}{\pi u \sigma_y(x) \sigma_z(x)} \exp \left[-1/2 \left(\frac{H}{\sigma_z(x)}\right)^2\right]$$
 (A.1)

where:

x = Downwind distance from source (m),

X(x) = Ground-level centerline concentration at  $x (\mu g/m^3)$ ,

Q = Source emission rate (g/sec),

u = Wind speed (m/sec),

ow(x) = Horizontal dispersion coefficient (m)

 $\sigma_{z}(x)$  = Vertical dispersion coefficient (m), and

H = Effective stack height (m) =  $h_s + \Delta h$  = (Physical stack height) + (Plume rise).

To derive an analytic expression for the mglc, the following commonly used representatives of the two dispersion coefficients are used:

$$\sigma_{\mathbf{v}}(\mathbf{x}) = \mathbf{a}\mathbf{x}^{\mathbf{b}} \tag{A.2}$$

and

$$\sigma_{\pi}(\mathbf{x}) = \mathbf{c}\mathbf{x}^{\mathbf{d}}.\tag{A.3}$$

The parameters a, b, c, and d depend upon atmospheric stability class and, for  $\sigma_z$ , the downwind distance x. The following expressions for the estimated mglc  $(\chi_m)$  and the corresponding downwind distance  $\chi_m$  may be derived. 50

$$\chi_{\underline{m}} = \frac{AQ \times 10^6}{\pi u} \times \frac{1}{\mu^{2\alpha}}$$
 (A.4)

and

$$x_{m} = \left[ \left( \frac{H^{2}}{c} \right) \frac{1}{2\alpha} \right]^{1/2d} \tag{A.5}$$

where:

$$\alpha = (b+d)/(2d) \tag{A.6}$$

and

$$A = \frac{c^{2\alpha-1}}{a} (2\alpha)^{\alpha} \exp(-\alpha)$$
 (A.7)

Values for a, b, c, d, and A are presented in Table A.1.

Both  $X_m$  and  $x_m$  depend on stability class and wind speed. To estimate these quantities, the plume rise must be estimated because both depend upon the effective stack height H. Plume rise can be estimated using the formulas of Briggs. 52,53

Setting

$$F = g\left(\frac{T - Ta}{T}\right)V \tag{A.8}$$

where:

g = Acceleration of gravity (9.8 m/sec<sup>2</sup>),

T = Exit gas temperature (°K),

 $T_a = Ambient temperature (°K), and$ 

V = Exist gas flow rate at temperature T (m<sup>3</sup>/sec),

it can be shown that

$$\Delta h(n/u) = C/u$$
 for neutral/unstable conditions (A.9)

and

$$\Delta h(s) = D/u^{1/3}$$
 for stable conditions. (A.10)

Table A.1 Dispersion Coefficient Parameters and Maximum Concentration Coefficient

Atmospheric			
Stability	Moderately Unstable	Neutral	Moderately Stable
Corresponding Pasquill-Gifford Stability Class	В	D	E-F (intermediate)
-	b		
a*	0.351	0.150	0.0853
P <b>*</b>	0.867	0.889	0.894
c**, <sup>†</sup>	0.139, 0.0494, 0.0494	0.0856, 0.259, 0.737	0.0682, 0.227, 1.437
d**, <sup>†</sup>	0.947, 1.114, 1.114	0.865, 0.687, 0.564	0.814, 0.618, 0.401
A <sup>†</sup>	0.335, 0.188, 0.188	0.396, 0.955, 3.85	0.468, 1.21, 34.7

<sup>\*</sup>Estimated from Fig. 3.2, Ref. 49.

<sup>\*\*</sup>Taken from Table 5, Ref. 51.

<sup>&</sup>lt;sup>†</sup>The first numbers given for each stability are appropriate at distances between 100 and 500 m, the second numbers at distances between 500 and 5000 m, and the third numbers at distances greater than 5000 m.

Assuming an ambient temperature of 293°K (20°C) and an ambient potential temperature lapse rate ( $\theta\theta/\theta z$ ) of 0.5°K/100 m, representative of moderately stable conditions,

$$F = 9.8 \left(\frac{T-293}{T}\right) V, \qquad (A.11)$$

$$C = 21.4F^{0.75} \text{ m}^2/\text{sec for } F<55 \text{ m}^4/\text{sec}^3,$$
 (A.12)

$$C = 38.7 \text{ F}^{0.6} \text{ m}^{2}/\text{sec} \text{ for F} > 55 \text{ m}^{4}/\text{sec}^{3}, \text{ and}$$
 (A.13)

$$D = 47.2F^{1/3} m^{4/3} sec^{-1/3}.$$
 (A.14)

A wind speed corresponding to the mglc can now be found. For neutral and unstable conditions,

$$u_{worst}(n/u) = \frac{b}{d} \frac{C}{h_s}, \qquad (A.15)$$

with a corresponding mglc

$$\chi_{\text{worst}}(n/u) = \frac{AQ \times 10^6}{\pi} \cdot \frac{1}{Ch_g^{b/d}} \cdot \frac{(b/d)^{b/d}}{(1+b/d)^{1+b/d}}$$
 (A.16)

For stable conditions

$$\chi_{\text{worst}}(s) = \frac{AQ \times 10^6}{\pi} \cdot \frac{u^{(b-2d)/3d}}{(u^{1/3}h_s + D)^{1+b/d}}$$
 (A.17)

Equation A.17 has no maximum unless b/d is greater than 2. Operationally, this difficulty is solved by setting u = 2 m/sec for the stable case in which case Eqs. A.10 and A.17 become

$$\Delta h(s) = 0.794 D$$
 (A.18)

and

$$\chi_{\text{worst}}(s) = \frac{AQ \times 10^6}{\pi} \cdot \frac{2^{(b-2d)/3d}}{(1.26 \text{ h}_s + \text{D})^{1+b/d}}.$$
 (A.19)

Equations A.15, A.16, and A.19 are the basic equations used to calculate the short-term mglc. The calculations need to be done separately for unstable, neutral, and stable conditions and the maximum value selected for the mglc. In addition, for each stability class, the calculations need to be done for three ranges of downwind distance because of the dependence of c, d, and A on x (see Table A.1). The value chosen for each stability class is the

maximum self-consistent value, that is, the maximum of the values for which the calculated  $\mathbf{x}_m$  falls within the range of downwind distances over which the particular c, d, and A values apply.

In implementing this procedure, high worst-case wind speeds are occasionally found which are unlikely to persist for periods of time on the order of hours to one day. On the other hand, low worst-case wind speeds are found which are small enough to render the Gaussian plume formulation inapplicable. To avoid both extremes and still retain a conservative estimate of the mglc, limits are placed on the worst-case wind speed for neutral/unstable conditions such that  $0.8 \le u_w \le 30$  m/sec.

Estimates made in this way are appropriate for averaging times of one hour. For averaging times out to about 24 hours, the one-hour estimates can be multiplied by an appropriate conversion factor from Table A.2. These factors represent a power law dependence of concentration on averaging time with an exponent of -0.17:

$$\chi(t) = \chi(1)t^{-0.17}$$
 (A.20)

For averaging times between 24 hours and about one month, a recognized simple procedure for estimating the concentration from a single source at one averaging time given the concentration at another averaging time appears to be lacking. Larsen<sup>54</sup> has developed a method which can be used in multi-source applications. For averaging times less than one month, he finds that for a year's data

$$\chi_{\max}(t) = \chi_{\max}(1 \text{ hr})t^{q}$$
 (A.21)

where q depends upon the geometric standard deviation of the concentration values. The form of Eq. A.21 with q = -0.17 is exactly the same as that of Eq. A.20. On the basis of this equivalence of mathematical form, the use of Eq. A.20 was extended beyond 24 hours to estimate conversion factors for 4 and 10 days as shown in Table A.2.

### A.2 LONG-TERM ESTIMATES

Expected monthly and annual mglc's from a single source are based upon the "sector-averaged" form of Eq.  $A.1:^{49,55}$ 

Table A.2 Averaging Time Conversion Factors

Averag Time		Conversion Factor
1		1.00ª
3		0.83ª
4		0.79ª
8		0.70ª
24		0.58ª
96	(4 da)	0.46 <sup>b</sup>
240	(10 da)	0.39 <sup>b</sup>

<sup>&</sup>lt;sup>a</sup>Based on Ref. 49.

$$\chi(\mathbf{x}) = \left(\frac{2}{\pi}\right)^{1/2} \frac{fQ \times 10^6}{u \sigma_{\mathbf{z}}(\mathbf{x}) \left(\frac{2\pi \mathbf{x}}{n}\right)} \exp \left[-1/2 \left(\frac{H}{\sigma_{\mathbf{z}}(\mathbf{x})}\right)^2\right]$$
(A.22)

where:

- n = the number of sectors into which the entire 360°
  range of wind directions is divided and
- f = the fraction of the time during which the wind direction lies in the sector of interest.

Using the same parameterization as above (Eq. A.3),

$$X_{\rm m} = \frac{BfQ \times 10^6}{\text{ub}^2 \beta}$$
 (A.23)

where:

$$\beta = (1+d)/2d$$
 (A.24)

and

$$B = \left(\frac{2}{\pi}\right)^{1/2} \frac{n}{2\pi} c^{2\beta-1} (2\beta)^{\beta} \exp(-\beta). \tag{A.25}$$

To estimate the expected long-term mglc, values of c and d for neutral atmospheric stability and distances between 500 and 5000 m are used and the plume rise is calculated using Eq. A.9. With these assumptions,

bSee discussion in text.

B = 0.256 and  $\beta = 1.23$ .

Examination of annual wind roses in Ref. 56 indicated that the maximum expected wind direction in a single  $22.5^{\circ}$  sector (n=16) is about 27% (f=0.27). For monthly wind roses, this maximum persistence is about 45% (f=0.45). The wind speed u used for both the annual and monthly calculations is u = 4.4 m/sec, corresponding to the nationwide annual mean wind speed based upon the speeds listed with the annual wind roses. For these conditions Eq. A.23 gives

$$\chi_{\rm m}({\rm yr}) = \frac{0.0157 \ {\rm Q} \times 10^6}{{\rm H}^2.46}$$
 for annual mglc's (A.26)

and

$$\chi_{\rm m}({\rm mo}) = \frac{0.0262 \, {\rm Q} \times 10^6}{{\rm H}^2.46}$$
 for monthly mglc's. (A.27)

## APPENDIX B

Pollutant Sensitivities of Plant Species

Table B.1. Sulfur Dioxide Sensitivity of Crop Speciesa

•	Sensitiv	ity			
Sensitive	2	Intermediate	Resistant		
Alfalfa	Leek	Cotton	Corn		
Apple	Lettuce		Sorghum		
Barley	Oats		Cantaloupe		
Bean, field	Okra		Citrus spp		
, lima	Onion				
Beet, sugar	Parsley				
, table	Parsnip				
Blackberry	Pea				
Blueberry	Pe ach				
Broccoli	Pear				
Brussels Sprouts	Pepper				
Cabbage	Plum, prune				
Carrot	Potato, Irish				
Celery	Potato, sweet				
Chard, Swiss	Pumpkin				
Cherry, sour	Radish				
, sweet	Raspberry				
Clover	Rye				
Clover, sweet	Safflower				
Cucumber	Soybean				
Currant, red	Spinach				
Eggplant	Squash				
Endive	Tobacco				
Gooseberry	Turnip				
Grapes	Wheat				
Kale					

<sup>&</sup>lt;sup>a</sup>Compiled from data in Ref. 16.

Table B.2. Sulfur Dioxide Sensitivity of Natural Vegetation<sup>2</sup>

#### Common Name

#### Scientific Name

## Sensitive

Alder, mountain
Aspen, large-toothed
, trembling

Ash, red (green)

, white

Birch, gray

, western paper

, white(paper)

, yellow

Blueberry, lowbush Cherry, bitter

Fir, subalpine

Grasses-bentgrass

-bluegrass

-desert grass

-Ky. bluegrass

-orchard grass

-red fescue

Hazel, beaked

, California

Hemlock, mountain

Larch, western

Maple, Manitoba

, Rocky Mt.

Mulberry, Texas

Pine, eastern white

, jack

, red

, Virginia

Rockspirea, creambush Serviceberry. low

, Saskatoon

, Utah

Sumac, staghorn

Tulip tree

Willow, black

Alnus tenuifolia Populus grandidentata

Populus tremuloides

Fraxinus pennsylvanica

Fraxinus americana

Betula populifolia

Betula papyrifera commutata

Betula papyrifera

Betula allegheniensis

Vaccinium angustifolium

Prunus emarginata Abies lasiocarpa

Agrostis palustris

Poa annua

Oryzopsis hymenoides

Poa pratensis

Dactylis glomerata

Festuca rubra

Corylus cornuta

Corylus cornuta californica

Tsuga mertensia Larix occidentalis

Acer negundo interius

Acer glabrum

Morus microphylla

Pinus strobus

Pinus banksiana

Pinus resinosa

Pinus virginiana

Holodiscus discolor

Amelanchier stolonifera

Amelanchier alnifolia

Amelanchier utahensis

Rhus typhina

Liriodendron tulipifera

Salix nigra

#### Scientific Name

### Intermediate

Basswood Birch, water Boxelder Chokecherry Cottonwood, black

, eastern

, narrowleaf Dogwood, red osier

Elm, American Fir, blasam , Douglas , grand

Grape, wild Hemlock, western Mahogany, mountain

Maple, Douglas , red

Mountain-ash, western

Oak, white Pine, lodgepole

, ponderosa , shortleaf

, western white

Poplar, balsam Sagebrush, big Snowberry, mountain

, Columbia

Spruce, Engelmann

, white

Witch hazel

Tilia americana Betula occidentalis Acer negundo Prunus virginiana Populus trichocarpa Populus deltoides Populus angustifolia Cornus stolonifera Ulmus americana Abies balsamea Pseudotsuga menziesii

Abies grandis Vitis riparia Tsuga heterophylla

Cercocarpus montanus Acer glabrum douglasii

Acer rubrum Sorbus scopulina Quercus alba Pinus contorta Pinus ponderosa Pinus echinata Pinus monticola Populus balsamifera Artemisia tridentata

Symphoricarpos oreophilus Symphoricarpos rivularis Picea engelmannii

Picea, glauca

Hamamelis virginiana

# Resistant

Black gum Buck-brush Buffalo-berry Ceanothus, redstem Cedar, western red , white(arborvitae) Dogwood, flowering Fir, silver , white Hawthorn, black

Nyssa sylvatica Ceanothus velutinus Shepherdia canadensis Ceanothus sanguineus Thuja plicata Thuja occidentalis Cornus florida Abies amabilis Abies concolor Crataegus douglasii

#### Table B.2. (Cont'd)

#### Common Name

## Scientific Name

# Resistant (cont'd)

Grape, Oregon Grasses-blue grama

-needle grass

-western wheatgrass

Juniper, common

, Rocky Mt.

, Utah

, Western

Kinnikinnick Locust, black

Mahogany, curl-leaf mt.

Maple, mountain

, silver

, sugar

Oak, gambel

, live

, northern red

, pin

Pine, limber

, pinyon

Poplar, Carolina

Sourwood

Spruce, blue

Squawbush

Sumac, smooth

Sycamore, American

Willow, shrubby

Yew, Pacific

aCompiled from lists in Refs. 9 and 16.

Odostemon aquifolium
Bouteloua gracilis
Stipa comata
Agropyron smithii
Juniperus communis
Juniperus scopulorum
Juniperus osteosperma
Juniperus occidentalis
Arctostaphylos uva-ursi
Robinia pseudoacacia
Cercocarpus ledifolius
Acer spicatum

Acer saccharinum
Acer saccharum
Quercus gambelii
Quercus virginiana
Quercus rubra
Quercus palustris
Pinus flexilis
Pinus edulis

Populus canadensis Oxydendron arboreum

Picea pungens Rhus trilobata Rhus glabra

Platanus occidentalis

Salix tristis Taxus brevifolia

Table B.3. Ozone Sensitivity of Crop Species<sup>a</sup>

	Sensitivity					
Sensitive	Intermediate	Resistant				
Alfalfab bean, pinto , white Broccoli Cloverb Corn, sweet Oatsb RadishC SafflowerC Soybeanb Spinachb Tobacco Tomatob	Bean, bush , lima Beet, table Cabbage Chard, swiss <sup>d</sup> Clover, white sweet Corn, field Cucumber <sup>d</sup> Potato, Irish Sorghum Squash, summer	Cotton Lettuce Onion				

aCompiled from data in Ref. 18.

bSome cultivars intermediate or resistant.

CSome cultivars intermediate.

dSome cultivars resistant.

Table B.4. Ozone Sensitivity of Natural Vegetationa

Common Name

#### Scientific Name

## Sensitive

Aspen, trembling Ash, red(green)

, white

Cottonwood, black Grasses-bent grass

-blue grass

-brome grass

Oak, gambel , white

Pine, Coulter

, eastern white

, jack , jeffrey

, loblolly , Monterey

, ponderosa

, Virginia

Serviceberry, Saskatoon Sycamore, American

Tulip tree `

Populus tremuloides Fraxinus pennsylvanica Fraxinus americana Populus trichocarpa Agrostis palustris Poa annua

Bromus tectorum Quercus gambelii

Quercus alba Pinus coulteri Pinus strobus Pinus banksiana

Pinus jeffreyi Pinus taeda Pinus radiata Pinus ponderosa

Pinus virginiana

Amelanchier alnifolia Platanus occidentalis Liriodendron tulipifera

# Intermediate

Boxelder

Cedar, incense

Grasses-Ky. bluegrass

-perennial rye -red fescue

Oak, black

, pin

, scarlet

Pine, lodgepole

, pitch

, shortleaf

, slash

, sugar

, Torrey

Redbud Sweetgum Acer negundo

Libocedrus decurrens

Poa pratensis Lolium perenne Festuca rubsa Quercus velutina Quercus palustris Quercus coccinea Pinus contorta

Pinus rigida Pinus echinata Pinus elliottii

Pinus lambertiana Pinus torreyana Cercis canadensis

Liquidambar styraciflua

Common Name

#### Scientific Name

# Resistant

Basswood

Birch, white (paper)

Black gum

Cedar, white (arborvitae)

Dogwood, flowering

Fir, balsalm

, Douglas

, white

Grasses-orchard grass

Hemlock

Juniper, western Locust, black

Maple, red

, sugar

Oak, mossy-cup

, northern red

, shingle

Pine, digger,

, red

Redwood Sequoia

Spruce, black

, blue

, white

Walnut, black

Tilia americana
Betula papyrifera
Nyesa sylvatica
Thuja occidentalis
Cornus florida
Abies balsamea

Pseudotsuga menziesii

Abies concolor Dactylis glomerata Tsuga canadensis

Juniperus occidentalis Robinia pseudoacacia

Acer rubrum Acer saccharum Quercus macrocarpa

Quercus rubra Quercus imbricaria Pinus sabiniana

Pinus resinosa Sequoia sempervirens

Sequoiadendron giganteum

Picea mariana Picea pungens Picea glaucaa Juglans nigra

<sup>&</sup>lt;sup>a</sup>Compiled from lists in Refs. 18 and 57.

Table B.5. Nitrogen Dioxide Sensitivity of Crop Species<sup>a</sup>

	Sensitivity			
Sensitive	Intermediate	Resistant		
Alfalfa	Bean, bush	Asparagus		
Barley	Celery	Cabbage, red		
Bean, pinto	Citrus spp.	, white		
Broccoli	Corn, sweet	Corn, field		
Carrot	Cotton	Cucumber		
Clover, crimson	Endive	Kale		
, red	Potato, Irish	Kohlrabi		
Leek	Rye	Onion		
Lettuce	Strawberry, pine	Sorghum		
Lucerne	Tomato			
Mustard, white	Wheat			
Oats				
Parsley				
Peas				
Radish				
Rhubarb				
Tobaccob				

aCompiled from lists in Refs. 19, 20, and 58.

 $<sup>^{\</sup>mathrm{b}}\mathrm{Some}$  cultivars intermediate or resistant.

Table B.6. Nitrogen Dioxide Sensitivity of Natural Vegetation<sup>a</sup>

Common Name

Scientific Name

### Sensitive

Grasses-Viper's grass

Scorzonera hispanica

## Intermediate

Fir, common silver

, white

Grasses-bluegrass Spruce, blue

, white

Abies pectinata

Abies alba

Poa annua

Picea pungens

Picea glauca

#### Resistant

Grasses-Ky. bluegrass

Poa pratensis

aCompiled from tables in Refs. 20 and 58.

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# APPENDIX C

Trace Element Air Quality Data

TABLE C-1. AIR QUALITY DATA FOR ARSENIC

	COUNTY		HINIHUM (UG/H3)			MAXIMUM (UG/M3)			
STATE		089	ARITH MEAN	GEO NEAN	089	ARITH HEAN	GEO MEAN		
но	BUCHANAN	0.8100			0.0100		~~~		
	CLAY	0.0100			0.0100				
	Jefferson	0.0100			0.0200	~~=			
TX	BEE	0.0200			0.0200				
	BEXAR	0.0200	6.0200A	0.0300A	0.0200	0.0200A	0.0300A		
	BOHIE	0.0200	0.0300A	0.0300A	0.1100	0.0300A	0.0300A		
	BRAZORIA	0.0200	0.0200	0.0300	0.0500	0.0300	0.0300		
	BRAZOS	0.0200	0.0300A	0.0300A	0.0700	0.0300A	0.030GA		
	BROUN	0.0200			0.0500				
	CALHOUN	6.0200			0.0200				
	CAHERON	0.0200	0.0200	0.0300	0.3000	0.0300	0.0300		
	CHAIBERS	0.0200			0.0200				
	DALLAS	0.0200	0.0200	0.0300	0.0600	0.0200	0.0300		
	DENTON	0.0200			0.0500				
	ECTOR	0.0200	0.020DA	0.03B0A	0.0600	0.0200A	0.0300A		
	ELLIS	0.0200			0.1000				
	EL PASO	0.0200	0.0200	0.0200	0.6600	0.0700	0.0500		
	GALVESTON	0.0200	D.0200A	0.0300A	0.4000	0.0200A	0.0300A		
	GRAY	0.0200			0.0200				
	GRAYSON	0.0200			0.0500				
	HALE	0.0200			0.0200				
	HARRIS	0.0200	0.0200	0.0300	<b>0</b> .1300	0.0300	0.0300		
	HAYS	0.0200			0.0700				
	HIDALGO	0.0200	8.0200	0.0300	0.1000	0.0300	0.0300		
	HOHARD	0.0200	0.0200A	0.0300A	0.0200	0.020GA	0.0300A		
	JEFF DAVIS	0.0200			0.0200				
	· JEFFERSON	0.0200	0.0200	0.0300	0.0500	0.0200	0.0300		
	LUBBOCK	0.0200			0.0700				
	MCLERRAN	0.0200			0.0200				
	HCHULLEN	0.0200			0.0200				
	HATAGORDA	0.0200			0.0200				
	MAVERICK	0.0200	0.0200A	0.0300A	0.0200	0.0200A	0.0300A		
	MIDLAID	0.0200	0.0200A	0.0300A	0.0300	0.0200A	0.0300A		
	HOUTGOLIERY	0.0200		***	0.0200				
	NOORE	0.0200			0.0200				
	HACOGDOCHES	0.0200			0.0200				
	HUECES	0.0200	0.0200	0.0300	0.2000	0.0300	0.0300		
	ORANGE	· 0.0200	0.0200A	0.0300A	0.0200	0.0200A	0.0300A		
	POTTER	0.0200			0.0200				
	SAN PATRICIO	0.0200			0.0200				
	SCURRY	0.0200			0.0200				

TABLE C-1. AIR QUALITY DATA FOR ARSENIC

			MAXIMUM (UG/M3)				
STATE	COUNTY	OB\$	ARITH HEAN	GEO HEAN	088	ARITH HEAN	GEO MEAN
	SHITH	0.0200	0.0300A	0.0300A	0.0700	0.0300A	0.0300A
	TARRANT	0.0200	~~~		0.1200		
	TAYLOR	0.0200	0.0300A	0.0300A	0.0500	0.0300A	0.0300A
	TITUS	0.0200	0.0200A	0.0300A	0.0500	0.0200A	0.0300A
	TON GREEN	0.0200			0.0500		
	TRAVIS	0.0200	0.0200	0.0300	0.0700	0.0300	0.0300
	VAL VERDE	0.0200			0.0200		
	VICTORIA	0.0200			0.0500		
	HALKER	0.8200			0.0200		
	HEBB	0.0200			0.0200		
	HICHITA	0.0200			0.0200		
	HISE	0.0200			0.0600		

TABLE C-2. AIR QUALITY DATA FOR CADMIUM

		t	IININUM (UG/	M3)	MAXINUM (UG/M		
STATE	COUNTY	OBS	ARITH MEAN	GEO HEAN	OBS	ARITH HEAN	6EO MEAN
AZ	APACHE COCONINO MARICOPA HONAVE NAVAJO PINA	0.0001 0.0001 0.0001 0.0001 0.0001	0.0001 0.0001 0.0003 0.0001 0.0002A 0.0006	0.0001 0.0001 0.0001 0.0001 0.0001A 0.0001	0.0100 0.0100 0.0700 0.0001 0.0100 3.0000	0.0002 0.0003 0.0040 0.0001 0.0002A 0.0037	0.0001 0.0001 0.0004 0.0001 0.0001A 0.0002
CO	LA PLATA HONTEZUNA	0.0001 0.0001	0.0001A 0.0001A	0.0001A 0.0001A	0.0100 0.0001	0.0002A 0.0001A	0.0001A 0.0001A
ID	SHOSHONE	0.0100	0.0095	0.0054	3.6800	0.4592	1.5670
IN	ALLEN BARTHOLOMEH CLARK DUGOIS ELKHART GRANT HOHARD JASPER JEFFERSON KNOX LAKE LA PORTE MARION HONROE ST. JOSEPH 'STEUBEN TIPPECAHOE VALDERBURGH VIGO HAYNE	0.0009 0.0002 0.0010 0.0009 0.0006 0.0010 0.0012 0.0003 0.0019 0.0002 0.0012 0.0012 0.0009 0.0006 0.0005 0.0005			0.0360 0.0016 0.0083 0.0024 0.0024 0.0042 0.0017 0.0011 0.0162 0.0031 0.0217 0.0217 0.0217 0.0217 0.0217 0.0250 0.0019 0.0048 0.0012 0.0056		
MN	BELTRANI BIG STONE BLUE EARTH CARLTON CLAY CROM HING DAKOTA GOODHUE NEMMEPIN	0.0010 0.0010 0.0010 0.0010 0.0010 0.0010 0.0010			0.0020 0.0010 0.0020 0.0020 0.0020 0.0020 0.0030 0.0050		

TABLE C-2. AIR QUALITY DATA FOR CADMIUM

			AININUM (UG	/H3)	MAXINUM (UG/M3)		
TATE	COUNTY	280	ARITH KEAN	GEO MEAN	OBS.	ARITH MEAN	GEO HEAI
	ITASCA	0.0010			0.0010		
	KANDIVOHI	0.0010			0.0020		
	Koochiching	0.0010			0.0110		
	LYON	0.0010			0.0040		
	HCLEOD	0.0010			0.0010		
	HILLE LACS	0.0010			0.0010		
	HOHER	0.0010			0.0020		
	NOBLES	0.0010			0.0020		
	OLMSTED	0.0010			0.0450		
	OTTERTAIL	0.0010			0.0020		
	POLK	0.0018			0.0050		
	ST. LOUIS	0.0010			0.0030		
	SCOTT	0.0010			0.0020		
	<b>S</b> TEARNS	0.6010			0.0180		
	Hashington	0.0010			0.0040		
	AMONIH	0.0010			6.0020		
)	ADAIR	0.0010			0.0042		
	AUDRAIN	0.0016			0.0088		
	BOONE	0.0007			0.0140		
	BUCHANAN	0.0016			0.0440		
	BUTLER	0.0003			0.0062		
	CALLAHAY	0.0012			0.0055		
	CAMDEN	0.0003			0.0046		
	CAPE GIRARDEAU	0.0020			0.0050		
	CLAY	0.0016			0.0150		
	COLE	0.0012			<b>0.0</b> 066		
	Jasper	0.0024			0.0079		
	· JEFFERSON	0.0027			1.4350		
	Livingston	0.0010			0.0052		
	MARION	0.0012			0.0050		
	NEH HADRID	0.0003			0.0045		
	NODAHAY	0.0010			0.0052		
	PETTIS	0.0008			0.0125		
	FIIELPS	0.0011			0.0053		
	PLATTE	0.0010			0.0109		
	ST. CHARLES	0.0050			0.0080		
	STE. GENEVIEVE	0.0031			0.0080		
	SCOTT	0.0010			0.0074		
	VERHON	0.0015			0.0041		
r	DEER LODGE	0.0100			0.0500		
1	RIO ARRIBA	0.0001	0.0001	0.0001	0.0100	0.0002	0.006

TABLE C-2. AIR QUALITY DATA FOR CADMIUM

	COUNTY		IINIHUH (UG	H3)	HAXIHUH (UG/H3)		
STATE		08\$	ARITH MEAN	GEO MEAN	OBS	ARITH MEAN	GEO MEAN
	SAN JUAN	0.0001	0.0002	0.0001	0.2000	0.0002	0.0001
OK	OKLAHOMA	0.0001	0.0008	0.0006	5.0000	0.2739	0.0012
SC	CHARLESTON	0.0020	~~~		0.0020		
TN	ANDERSON	0.0010			0.0030		
•••	BEDFORD	0.0010			0.0010		
	BLOURT	0.0010			0.0040		
	BRADLEY	0.0010			0.0010		
	CAMPDELL	0.0010			0.0040		
	CARTER	0.0010			0.0010		
	COFFEE	0.0010			0.0050		
	CURBERLAND	0.0010			0.0020		
	DYER	0.0810			0.0030		
	GIBSON	0.0010			0.0070		
	GREENE	0.0010			0.0010		
	HANBLEN	0.0010			0.0090		
	HENRY	0.0010		***	0.0030		
	HUNFHREYS	0.0010			0.8040		
	LINCOLH	0.0010			0.0010		
	MCHINA	0.0040			0.0040		
	MADISON	0.0010			0.0030		
	MARION	0.0010			0.0030		
	HAURY	0.0010			0.0030		
	HONTGOHERY	0.0010			0.0100		
	OBION	0.0010			0.0030		
	POLK	0.0010			0.0370		
	· PUTNAH	0.0010			0.0010		
	ROANE	0.0010			0.0090		
	ROBERTSON	0.0010			0.0050		
	RUTHERFORD	0.0010			0.0010		
	SULLIVAN	0.0010			0.0030		
	SUMMER	0.0010			0.0020		
	HARREN	0.0010	4-44 P		0.0010		
	HASHINGTON	0.0010			0.0010		
	HILLIANSON	0.0010			0.0010-		
	HILSON	0.0010	~		0.0010		
TX	BEE	0.0300			0.0300		
***	BEXAR	0.0300	0.0300A	0.0300A	0.0300	0.0300A	0.0300A
	BOLITE	0.0300	A00E0.0	0.0300A	0.0300	A0020.0	A0020.0
	BRAZORIE	0.0300	0.0300	0.0300	0.0300	0.0300A	0.0300A 0.0300
		2.2300	0.0300	4.0300	0.0300	4.0300	9.0200

TABLE C-2. AIR QUALITY DATA FOR CADMIUM

	COUNTY	1	MINININ (US/N3)			MAXIMUM (UG/N3)			
STATE		0B\$	ARITH HEAN	GEO MEAN	OBS	ARITH MEAN	GEO MEAN		
	BRAZOS	0.0300	0.0300A	0.0300A	0.0300	0.0300A	0.0300		
	BROWN	0.0300			0.0300				
	CALHOUN	0.0300			0.0300				
	CAHEROH	0.0300	0.0300	0.0300	0.0300	0.8300	0.0300		
	CHAIBERS	0.0300			0.0300				
	DALLAS	0.0300	0.0300	0.0300	0.1000	0.0300	0.0300		
	DENTON	0.0300			0.0300				
	ECTOR	0.0300	0.0300A	0.0300A	0.0300	0.0300A	0.0300		
	ELLIS	0.0300			0.0300				
	EL PASO	0.0300	0.0300	0.0300	0.1000	0.0300	0.0300		
	GALVESTON	0.0300	0.0300A	0.0300A	0.1000	0.0300A	0.0300		
	6RAY	0.0300			0.0300				
	GRAYSON	0.0300			0.0300				
	HALE	0.0300			0.0300				
	HARRIS	0.0001	0.0300	0.0300	0.0300	0.0300	0.0300		
	HAYS	0.0300			0.0300				
	HIDALGO	0.0300	0.0300	0.0300	0.0300	0.0300	0.0300		
	HOHARD	0.0300	0.0300A	0.0300A	0.0300	0.0300A	0.0300		
	JEFF DAVIS	0.0300			0.0300				
	JEFFERSON	0.0300	0.0300A	0.0300A	0.0300	0.0300A	0.0300		
	LUBBOCK	0.0300			0.0300				
	HCLEHHAN	0.0300			0.0300				
	NCHULLEN	0.0300			0.0300				
	HATAGORDA	0.0300			0.0300				
	MAVERICK	0.0300	O.0300A	0.0300A	0.0300	0.0300A	0.030		
	HIDLAND	0.0300	0.0300A	0.0300A	0.0380	0.0300A	0.0300		
	MONTGOHERY	0.0300			0.0300				
	NOORE	0.0300			0.0300				
	<ul> <li>NACOGDOCHES</li> </ul>	0.0300	~~~		0.0308				
	HUENCES	0.0300	0.0300	0.0300	0.2000	0.0300	0.030		
	ORANGE	0.0300	0.0300A	0.0300A	0.0300	0.0300A	0.030		
	POTTER	0.0300			0.0300				
	SAN PATRICIO	0.0300			0.0300				
	SCURRY	0.0300			0.0300				
	Shith	0.0300	0.0300A	0.0300A	0.0300	0.0300A	0.030		
	TARRANT	0.0300			0.0300				
	TAYLOR	0.0300	0.0300A	0.0300A	0.0300	O.0300A	0.030		
	TITUS	0.0300	0.0300A	0.0300A	0.0300	0.0300A	0.030		
	TOM GREEN	0.0300			0.0300				
	TRAVIS	0.0300	0.0300	0.0300	0.0300	0.0300	0.030		
	VAL VERDE	0.0300			0.0300				
	VICTORIA	0.0300			0.0300				
	HALKER	0.0300		~~~	0.0300				

TABLE C-2. AIR QUALITY DATA FOR CADMIUM

STATE	COUNTY	HININUM (UG/13)			MAXIMUM (UG/M3)			
		089	ARITH MEAN	GEO MEAN	008	ARITH HEAN	GEQ HEAN	
	HEBB	0.0200		***	0.0300			
	HICHITA	0.0300			0.0300			
	HISE	0.0300			0.0300			
UT	EMERY	0.0001			0.0001			
-	GARFIELD	0.0801	0.0001	0.0001	0.0200	0.0803	0.0001	
	KANE	0.0001	0.0002A	0.0081A	0.0100	0.0002A	0.0001A	
	SAN JUAN	0.0001	0.0001	0.0001	0.0100	0.0020	0.0001	
	HASHINGTON	0.0001	0.0001	0.0001	0.0100	0.0002	8.0001	

TABLE C-3. AIR QUALITY DATA FOR CHROHIUM

			HININUM (UG/	M3)	MAXIHUM (UG/M3)			
STATE	COUNTY	0B3	ARITH MEAN	GEO MEAN	OBS	ARITH HEAN	GEO HEAN	
AZ	APACHE	0.0010	0.0020	0.0010	0.0500	0.0030	0.0010	
	COCONINO	0.0010	0.0020	0.0010	0.0700	0.0050	0.0020	
	MARICOPA	0.0010			0.0010			
	HOHAVE	0.0010	0.0010	0.0010	0.0300	0.0040	0.0020	
	OLAVAN	0.0010	0.0030A	0.0010A	0.0500	0.0030A	0.0010A	
co	LA PLATA	0.0010	0.0020A	0.0010A	0.0500	0.0640A	0.0020A	
	MONTEZUMA	0.0010	0.0030A	0.0010A	0.0400	0.0030A	0.0010A	
МО	ADAIR	0.0060			0.0980	***		
	AUDRAIN	0.0060			0.0920			
	BOOHE	0.0040			0.0170			
	BUCHANAN	0.0060			0.0860			
	BUTLER	0.0050			0.0130			
	CALLAHAY	0.0650			0.0670			
	CANDEN	0.0040			0.0810			
	CAPE GIRARDEAU	0.0060			0.0180			
	CLAY	0.0070			0.0600			
	COLE	0.0060			0.0640			
	JASPER	0.0070			0.0150			
	JEFFERSON	0.6030	~~-		0.0640			
	LIVINGSTON	0.0040			0.0680			
	MARION	0.0060			0.0670			
	NEN MADRID	0.0050			0.2370			
	HODAHAY	0.0060			0.0610			
	PETTIS	0.0070			0.0110			
	PHELPS	0.0050			0.0840			
	· PLATTE	0.0050			0.0620			
	ST. CHARLES	0.0050			0.0080			
	STE. GEHEVIEVE	0.0090			0.0090			
	SCOTT	0.0050			0.0780			
	VERNON	0.0070			0.0130			
IN	ALLEN	0.0050			0.0270			
	BARTHOLOMEH	0.0010			0.0140			
	ELKHART	0.0020			0.0090			
	LAKE	0.0060			0.0160			
	MORROE	0.0050			0.0118			
	VANDERBURGH	0.0040		~	0.0100			
NM	RIO ARRIBA	0.0010	0.0020A	0.6010A	0.0500	0.0030A	0.0010A	
	HAUL HAS	0.0010	0.0020	0.0010	0.0500	0.0030	0.0010	

TABLE C-3. AIR QUALITY DATA FOR CHROHIUM

•		H	HINIMUN (UG/M3)			MAXINUM (UG/13)			
STATE	COUNTY	OBS	ARITH HEAN	GEO HEAN	08s	ARITH HEAN	GEO HEAN		
SC	CHARLESTON	0.1530			0.1530	***			
TN	ANDERSON	0.0010		~ * *	99.0000				
***	BEDFORD	0.0010			0.0050				
	BLOURIT	0.0010	***		0.0050				
	BRADLEY	0.0010			0.0050				
	CANFBELL	0.0010			0.0050				
	CARTER	0.0010			0.0010				
	COFFEE	0.0010			0.0050				
	CURBERLAND	0.0010			0.0090				
	DYER	0.0010			0.0050				
	6IBSON	0.0010			0.0050				
	GREENE	0.0010			0.0010				
	HANBLEN	0.0010			0.0050				
	HENRY	0.0010			0.0050				
	HARIPIREYS	8.0010			0.0050				
	LINCOLN	0.0010			0.0030				
	HCHINN	0.0050			0.0010				
	MADISON	0.0010			0.0050				
	MARTON	0.0010			0.0050				
	HAURY	0.0010			0.0050				
	MONTGOMERY	0.0010			0.0030				
	OBION	0.0010			0.0050				
	POLK	0.0010			0.0050				
	PUTNAM	0.0010			0.0020				
	ROANE	0.0010			0.0020				
	ROBERTSON	0.0010			0.0050				
	· RUTHERFORD	0.0010			0.0050				
	SULLIVAN	0.0010			0.0050				
	SURIER	0.0010							
	HARREN	0.0010			0.0050 0.0050				
	HASHINGTON	0.0010							
	HILLIAMSON	0.0010			0.0010 0.0010				
	HILSON	0.8010			0.0050				
TX	BEE	0.0200			0.0200				
	BEXAR	0.0200			0.0900				
	BOHTE	0.0200			0.0900				
	BRAZORIA	0.0200			0.0300				
	BRAZOS	0.0200			0.0200				
	BROLIN	0.0200			0.0200				
	CALHOUN	0.0200			0.0200				

TABLE C-3. AIR QUALITY DATA FOR CHRONIUN

			IINIMUM (UG/	H31	М	AXIMUM (UG/	113)
STATE	COUNTY	OBS	ARITH MEAN	GEO NEAN	028	ARITH HEAH	GEO MEAN
	CAMERON	0.0200			0.2400		
	CHATIBERS	0.0200	~~~		0.0200		
	DALLAS	0.0200			0.0800		
	DENTON	0.0200			0.0200		
	ECTOR	0.0200			0.0500		
	ELLIS	0.0200			0.0200		
	EL PASO	0.0200			0.1400		
	<b>GALVESTON</b>	0.0200			0.1000		
	GRAY	0.0200			0.0200	~	~~~
	6RAYSON	0.0200			0,0200		
	HALE	0.0200			0.0200		
	HARRIS	0.0010			0.5600		
	HAYS	0.0200			0.0200		
	HIDALGO	0.0280			0.0200		
	HOHARD	0.0200			0.0500		
	JEFF DAVIS	0.0200			0.0500		
	JEFFERSON	0.0200			0.0700		
	LUBBOCK	0.0280			0.0200		
	MCLENNAN	0.0200			0.0200		
	KCHULLEN	0.0200			0.0200		
	HATAGORDA	0.0200			0.0200		
	MAVERICK	0.0200			0.0200		
	MIGLAND	0.0200			0.0200		
	HONTGONERY	0.0200			0.0200		
	MOORE	0.0200			0.0230		
	NAGODOCHES	0.0200			0.0200		
	NUEHCES	0.0200			0.3100		
	ORANGE	0.0200			0.0200		
	· POTTER	0.0200			0.0200		
	SAN PATRICIO	0.0200			0.0200		
	SCURRY	0.0200			0.0200		
	SMITH	0.0200			0.0700		
	TARRANT	0.0200			0.0700		
	TAYLOR	0.0200			0.0700 0.0700		
	TAYLOR	0.0200			0.0200		
	TON GREEN	0.0200			0.0200		
	TRAVIS	0.0200			0.0200 0.0700		
	VAL VERDE	0.0200			0.0700		
	VICTORIA	0.0200					
	HALKER	0.0200 0.0200			0.0200		
	HEDB	0.0200			0.0200		
	HICHITA	0.0200			0.0200		
					0.0200		
	HISE	0.0200			0.0200		

TABLE C-3. AIR QUALITY DATA FOR CHROHIUM

STATE	COUNTY	HININIA (UG/H3)			MAXIMM (UG/H3)		
		OBS	ARITH HEAN	6EO MEAN	089	ARITH HEAH	GEO HEAN
ហ	EHERY GARFIELD KANE SAN JUAN MASHINGTON	0.0010 0.0018 0.0010 0.0010 0.0010	0.0020 0.0030A 0.0010 0.0020	0.0010 0.0010A 0.0010 0.0010	0.0300 0.0500 0.0300 0.0400 0.0400	0.0040 0.0030A 0.0046 0.0030	0.0010 0.0010A 0.0020 0.0010

TABLE C-4. AIR QUALITY DATA FOR FLUORIDE ION

	COUNTY		MINIMUN (UG/M3)			MAXINUM (UG/M3)		
STATE		OBS	ARITH MEAN	GEO HEAN	OBS	ARITH MEAN	GEO MEAN	
AZ	MARICOPA	0.0200	0.0300	0.0300	0.3700	0.0500	0.0400	
ND	BARNES	0.0200	0.0200A	0.0300A	0.0500	0.0200A	0.0300A	
	BILLINGS	0.0200	0.0200A	0.0300A	0.0800	0.0300A	0.0300A	
	BORHAN	0.0200	0.0200A	0.0300A	0.0200	0.0200A	0.0300A	
	BURLETGH	0.0200	0.0300A	0.0300A	0.3500	0.0400A	0.0300A	
	CASS	0.0200	0.0200	0.0300	0.0300	0.0300	0.0300	
	DUNI	0.0200	0.0200	0.0300	0.0200	0.0200	0.0300	
	GRAND FORKS	0.0200	0.0200	0.0300	0.1900	0.0300	0.0300	
	6RANT	0.0200	0.0300A	0.0300A	0.1200	0.0300A	0.0300A	
	HETTINGER	0.0200	0.0200	0.0300	0.0200	0.0200	0.0300	
	HCKEHZIE	0.0200	0.0200A	0.0300A	0.0200	0.0200A	0.0300A	
	HCLEAN	0.0200	0.0200A	0.0300A	0.1600	0.0200A	0.0300A	
	HERCER	0.0200	0.0200	0.0300	0.3000	0.0600	0.0400	
	HORTON	0.0200	0.0200	0.0300	0.0600	0.0300	0.0300	
	MOUNTRAIL	0.0200	0.0200	0.0300	0.0200	0.0200	0.0300	
	OLIVER	0.0200	0.0200	0.0300	0.1100	0.0300	0.0300	
	RANSEY	0.0200	0.0200A	0.0300A	0.0200	0.0200A	0.0300A	
	RICHLAND	0.0200	0.0200	0.0300	0.0200	0.0200	0.0300	
	SHERIDAN	0.0200	0.0200	0.0300	0.0200	0.0200	0.0300	
	STARK	0.0200	0.0200	0.0300	0.0200	0.0200	0.0300	
	STUTSHAN	0.0200	0.0200	0.0300	0.1400	0.0300	0.0300	
	HARD	0.0200	0.0200A	0.0300A	0.0200	0.0200A	0.0300A	
	HILLIANS	0.0200	0.0200	0.0300	0.0200	0.0200	0.0300	

TABLE C-5. AIR QUALITY DATA FOR LEAD

			INIHUH (UG/	H3)	MAXIMUM (UG/H3)			
STATE	COUNTY	OBS	ARITH HEAN	GEO MEAN	OBS	ARITH HEAH	GEO MEAN	
AL	ETOHAH	0.1700	Q.530QA	0.4400A	2.0600	0.5300A	0.4400A	
	JEFFERSON	0.2000	0.9400A	0.8400A	4.2300	0.9400A	0.8400A	
	Madison	0.2200	0.6500A	0.5500A	1.9600	0.6500A	0.5500A	
	MOBILE	0.3500			2.9400			
	HONTGONERY	0.2300			3.0900			
AZ	APACHE	0.0010	0.0140	0.0010	0.2000	0.0320	8.0160	
	COCHISE	0.0010	0.0190A	0.8090A	0.9300	0.0190A	0.0090A	
	COCONINO	0.0018	0.0120	0.0060	0.6200	0.0260	0.0130	
	GILA	0.0010			0.4000			
	GRAHAM	0.1000			0.4000			
	GREEHLEE	0.0010			0.4000		'	
	MARICOPA	0.0010	0.5530	0.0220	9.3570	2.3240	1.7560	
	HOHAVE	0.0010	0.0130	0.0050	0.3000	0.0180	0.0078	
	OLAVAM	0.0010	0.0150	0.0070	0.3000	0.0150	0.0078	
	PIHA	0.0010	0.3600	0.3210	2.1870	1.0010	0.8470	
	PINAL	0.1000			0.5000			
	YAVAPAI	0.0010			0.2000			
	YUNA	0.1000			0.4800		***	
AR	CRITTENDEN	0.3800			2.8700			
	HILLER	0.2300			1.0500			
	HONTGOHERY	0.0500			0.1500			
	PULASKI	0.3000	0.9100A	0.8390A	1.6300	0.9100A	0.8390A	
CA	ALAHEDA	1.4900	0.7210	0.6110	6.1100	1.2150	0.8800	
	FRES:10	0.2900	1.5100A	1.1200A	5.1300	1.5100A	1.1200A	
	· KERN	0.1960	1.4310A	1.0960A	5.5320	1.4310A	1.0960A	
	LOS ANGELES	0.5200	1.9100	1.6100	8.9400	2.5900	2.6500	
	MADERA	0.2450			2.0610			
	HERCED	0.2110			0.8700			
	HODOC HONTEREY	0.1860			0.1860			
	NAPA	0.0450 0.0600	0.5020A 0.7290A	0.4160A	1.5120	0.5020A	0.4160A	
	ORALIGE	0.4400	1.8000A	0.5460A 1.4200A	4.0500	0.7290A	0.5460A	
	RIVERSIDE	0.0900	0.6120	0.5310	6.4980 4.5490	1.8000A	1.4200A	
	SACRAIJENTO	0.2600	0.6310	0.5310	8.5000	1.8390	1.6740	
	SAN BERNARDINO	0.4100	1.5600	1.4808	4.5500	1.1000 1.6550	0.5150	
	SAN DIEGO	0.2500	0.9680	0.9900	4.2500 6.2600	1.5930 1.5930	1.4820	
	SAN FRANCISCO	0.3400	0.9000A	0.8080A	6.9100	1.0290A	1.1040	
	SAN JOAQUIN	0.2880		G.BUUUA	0.4630	1.UZYUA	A0888.0	
•	SAN HATEO	0.0100	0.6580A	0.4610A	4.2600	0.6580A	0.4610A	

TABLE C-5. AIR QUALITY DATA FOR LEAD

			IIIIIIUM (UG/	M3)		MAXIMUM (UG/	H3)
STATE	COUNTY	0BS	ARITH MEAN	GEO NEAN	OBS	ARITH MEAN	GEO HEAN
	SANTA BARBARA	0.0870	1.1050A	0.7940A	5.2470	1.1050A	0.7940A
	SANTA CLARA	0.3100	1.6200	0.8460	8.5000	1.4280	1.1180
	SISKIYOU	0.0620	0.2600A	0.2050A	0.8430	0.2600A	0.2050A
	SOLARO	0.0300			1.3000		
	SOHOHA	0.0800	0.4840	0.3480	2.0000	0.4840	0.3480
	TEHANA	0.0670			0.3460		
	VEHTURA	0.1980			1.9870		
CO	DENVER	0.4000			3.6400		
	LA PLATA	0.0010	0.0170	0.0060	0.1700	0.0220	0.0100
	HOHTEZUNA	0.0	0.0250A	0.0120A	0.1100	0.0250A	0.0120A
CT	FAIRFIELD	0.6300	1.3020A	1.2060A	2.4600	1.3020A	1.2060A
	HARTFORD	0.3400	1.0650A	0.9890A	2.2500	1.0650A	0.9890A
	NEH HAVEN	0.4680	1.1220	1.0100	4.1900	1.9160	1.7340
DE	KENT	0.0500	0.1600A	0.1300A	0.5200	0.1600A	0.1300A
	NEH CASTLE	0.0700	0.5300	0.3900	3.0700	1.5000	1.4200
20	HASHINGTON	0.4600			3.1800		
FL	DADE	0.1000	0.3460	0.2590	6.9000	2.0270	1.7460
	DUVAL	0.3000	0.8900A	<b>0</b> .8000A	2.7200	0.0390A	0.8000A
	HARDEE	0.0400			0.5200		
	HIGHLANDS	0.0			0.1100		
	HILLSBOROUGH	0.1500			2.5200		
	PINELLAS	0.1100			1.3100		
GA	· CHATHAM	0.1000	0.5590A	0.4600A	1.3700	0.5590A	0.4600A
	FULTON	0.4100	1.2000A	1.0700A	3.2800	1.2000A	1.0700A
	NUSCOGEE	0.1900	0.6100A	0.5500A	2.9400	0.6140A	0.4870A
ID	ADA	0.2800	0.7800A	0.7000A	2.6200	0.7800A	0.7000A
	BAHNOCK	0.1200			1.0500		
	BUTTE	0.0			0.1300		
	HEZ PERCE	0.2000		~	1.8300		
	SHOSHOHE	0.0200	0.5170	0.4390	82.0900	15.725 <b>0</b>	11.7850
IL	COOK	0.1500			4.3800		
	LAKE	0.1500			1.9200		
	PEORIA	0.2800			3.9000		
	ROCK ISLAND	0.1600			1.9900		
	ST. CLAIR	0.0300			1.4400		

TABLE C-5. AIR QUALITY DATA FOR LEAD

		4	inikum (UG/	M3)	MAXIMUM (UG/M3)		
STATE	COUNTY	OBS	ARITH HEAM	GEO MEAN	08\$	ARITH HEAN	GEO MEAN
	SANGAHON	0. 1800			0.8300		
	HILL	0.3100			1.2100		
	HIMHEBAGO	0.2000			2.0500		
IN	ALLEN	0.2940	0.6810A	0.6130A	1.8200 1.0460	0.6810A	0.6130A
	BARTHOLOMEH	0.1290	0.4300A	0.3660A	1.0460	0.4300A	0.3660A
	CLARK	0.3850	1.3370A	1.1770A	3.4040	1.3370A	1.1770A
	DELAHARE	0.3300			2.2900		
	DUBOIS	0.1630	0.4870A	0.4440A	1.0780	0.4876A	0.4440A
	ELKHART	0.2560	0.5080A	0.4770A	1.1020	0.5080A	0.4770A
	FLOYD	0.4600			2.0600		
	GRANT	0.1940	0.4850A	0.4420A	0.5970	0.4850A	0.4420A
	HOLIARD	<b>0.1</b> 400	0.5150A	0.4530A	1.5180	0.5150A	0.4580A
	JASPER	0.0310	0.1210A	0.1070A	0.2720	0.1210A	0.1070A
	Jefferson	0.1300	0.3770A	<b>0.3380</b> A	1.0410	0.3770A	0.3380A
	KNOX	0.2060	0.4990A	0.4620A	1.2810	0.4990A	0.4620A
	LAKE	Đ. 108O	0.5910	0.4910	6.8700	0.7320	0.6440
	LA PORTE	0.2380	0.4400A	0.4160A	0.7700	0.4400A	0.4160A
	MARION	0.0560	0.8300A	0.7070A	5.2550	0.8300A	0.7070A
	HOHROE	0.0400	0.7280A	0.6360A	1.8610	0.7280A	0.6360A
	PARKE	0.0600			0.2800		
	ST. JOSEPH	0.2350	0.6000A	0.5190A	1.3500	0.6000A	0.5190A
	STEUBEN	0.0440	<b>0</b> .1780A	0.1640A	0.3760	0.1780A	0.1640A
	TIPPECAHOE	0.2220	0.5300A	0.4930A	1.0500	0.5300A	0.4930A
	VANDERBURGH	0.1720	0.5250A	0.4670A	1.3620	0.5250A	0.4670A
	V160	0.1930	0.5120A	0.4420A	1.5840	0.5120A	0.4420A
	HAYNE	0.1960	0.5280	0.4600	1.6340	0.6270	0.5840
IA	• BLACKHAHK	0.1700	0.4300A·		2.0100	0.5840A	0.5090A
	DELAHARE	0.0520			0.2800		
	DUBUQUE	0.0200			99.0000		
	LEE	0.1100			0.3200		
	LIM	0.1500	0.5400A	0.4500A	1.8300	0.5180A	0.4680A
	POLK	0.3400	1.0070A	0.9030A	2.7500	1.0070A	0.9030A
	POTTAHATAMIÉ	0.4500			0.6500		
	SCOTT	0.3000			2.6700		
KS	SEDGHICK	0.2000			1.1400		
	SHALNEE	0.1600			1.6400		
	HYANDOTTE	0.0900	0.5100	0.4600	3.0200	9.4280	0.3840
KY	BOYD	0.1300			3.8900		
	FAYETTE	0.2900			3.5600		

TABLE C-5. AIR QUALITY DATA FOR LEAD

		i	MINIHUM (UG/	H3)		IAXIHUM (UG/	H3)
STATE	COUNTY	08\$	ARITH HEAN	GEO MEAN	ons	ARITH MEAN	GEO HEAN
	JEFFERSON	0.2700	0.9760A	0.9400A	3.3800	1.1800A	0.8910A
	KENTON	0.3200	0.7100A	A0083.0	1.2200	0.7100A	0.6800A
	HARREN	0.1900	A008P.0	0.4100A	1.1500	0.4800A	0.4100A
LA	CADDO PARRISH	0.2000	0.6720A	0.6100A	1.4400	0.6720A	0.6100A
	EAST BATON ROUGE PARRISH	0.2500	1.1500A	0.9000A	4.2600	1.1500A	0.9000A
	IBERVILLE PARRISH	0.0300	0.1300A	0.1100A	0.3800	0.1300A	0.1100A
	ORLEANS PARRISH	0.3900	0.8000	0.7700	1.6400	1.0300	0.9830
HE	CUMBERLAND	0.1000	0.4500A	0.4000A	1.4700	0.4500A	0.4000A
	HANCOCK	0.0	0.0600A	0.0400A	0.3900	0.0600A	0.0400A
MD	BALTIMORE (CITY)	0.4400	1.0700A	0.9800A	2.5000	1.0700A	0.9800A
	CALVERT	0.0300	0.1700A	0.1400A	0.3900	0.1700A	0.1400A
MA	CENTRAL HA. APCO	0.3000	0.8400A	0.7400A	1.8600	0.8400A	0.7400A
	HETROPOLITAN BOSTON APCO	0.4000			1.3900		
	PIONEER VALLEY APCO	0.6500			2.9000		
	SOUTHEASTERN NA. APCO	0.1900			1.1700		
HI	GEHESEE	0.3800			1.3400		
	INGHAM	0.2100			1.7100		
	KEHT	0.2700			2.2500		
	SAGINAH	0.1400	0.4000A	0.3700A	0.9100	A000A.0	0.3700A
	HAYNE	0.1800		***	2.7100		
101	BELTRANI	0.0860			0.1560	~~-	
	BIG STONE	0.0060			0.1130		
	· BLUE EARTH	0.2690			0.7580		
	CARLTON	0.0530			0.2280		
	CLAY	0.0530			0.1150		
	CROH HING	0.0130			0.2790		
	DAKOTA	0.0550 0.3130			0.4750		
	GOODHUE HEIRIEPIN	0.3130 0.1800	1.2700A	0.9400A	1.2540	4 0300	
	ITASCA	0.0220	1.270UA	U.74UUA	8.8100	1.2700A	0.9400A
	KANDIYOHI	0.0220			0.1190		
	KOOCHICHING	0.0320			0.2040 0.4450		
	LYON	0.1730			0.3670		
	HCLEOD	0.1180			0.3200		
	HILLE LACS	0.3330			1.2220		
	HOHER	0.2190			0.3910	*	
	HOBLES	0.0520			0.1720	***	

TABLE C-5. AIR QUALITY DATA FOR LEAD

			IINIHUH (UG/	M3)		MAXIMAM (UG/)	
STATE	COUNTY	OBS	ARITH MEAN	GEO MEAN	089	ARITH MEAN	GEO MEAN
	OLMSTED	0.2150			2.3930		
	OTTERTAIL	0.0840			0.3140		
	POLK	0.0530			0.1320		
	RANSEY	0.0720			3.8600		
	ST. LOUIS	0.0818			2.3100		
	SCOTT	0.0320			<b>0.3</b> 430		
	STEARNS	0.0100			1.1350		
	HASHINGTON	0.2180			0.5610		
	HINONA	0.1110			0.5770		
нз	HINDS	0.1400	0.7500A	0.6700A	2.7600	0.7500A	0.6700A
	JACKSON	0.0020			0.4200		
Ю	ST. LOUIS (CITY)	0.2900	0.89B0A	0.8100A	2.9200	8.8900A	0.8100A
	ADAIR	0.1200			0.5540		
	AUDRAIN	0.1100			4.2300		
	BOOHE	<b>0.11</b> 00			2.9400		
	BUCHANAN	0.0500			2.8100		
	BUTLER	0.0190			4.0600		
	CALLAHAY	0.1150			0.3920		
	CAMDEN	0.0100			0.7280		
	CAPE GIRARDEAU	0.0680			0.3100		
	CLAY	0.0010			2.1790		
	COLE	0.0900			0.7300		
	JACKSON	0.3500	~~~		1.3900		
	JASPER	0.1600			0.9600		
	JEFFERSON	0.2340			37.5300		
	LIVINGSTON	0.0900			0.6120		
	· MARION	0.1700			1.0800		
	NEN HADRID	0.1000			0.6360		
	HODALIAY	0.0460			0.3150		
	PETTIS	0.1970			0.3500		
	PHELPS	0.0690			0.2300		
	PLATTE	0.0900			1.1420		
	ST. CHARLES	0.2400			0.6510		
	STE GENEVIEVE	0.1400			0.3700		
	SCOTT	0.1400			2.1900		
	SHAHHON VERNON	0.0300 0.1490	0.0790A	0.0710A	0.1900 0.3130	0.0790A	0.0710A
				<u>_</u>	9.3130		
MT	GLACIER	0.0			0.0600		
	JEFFERSON	0.0600			10.9700		
	LEHIS AND CLARK	0.3300			<b>24.6200</b>		***

TABLE C-5. AIR QUALITY DATA FOR LEAD

			NINIMUM (UG/	/H3)	MAXINUM (UG/M3)		
STATE	COUNTY	OBS	ARITH MEAN	GEO NEAN	OBS	ARITH MEAN	GEO MEAN
	POHDER RIVER ROSEDUD	0.0 0.0		***	0.0400 0.0300		
NB	DOUGLAS LANCASTER THOMAS	0.2400 0.1200 0.0020	0.3800A	0.3400A	1.9900 0.9500 0.0500	0.3800A	0.3400A
W	CLARK HASHOE HHITE PINE	0.5300 0.3100 0.0			6.5200 5.2200 0.0400		
ИН	CODS MERRIMACK	0.0020 0.2000			<b>0</b> .1600 1.5300	* ** **	
И	CAMDEN ESSEX GLOUCESTER HUBSON MERCER MIDDLESEX PASSAIC UNION	0.3100 0.2200 0.1300 0.2600 0.4000 0.4800 0.6100 0.6500			2.5200 2.1100 1.1200 2.4900 4.4200 1.4400 3.0500 4.8600		
191	BERNALILLO RIO ARRIBA SAN JUAN	0.4800 0.8616 0.0010	1.2700A 0.0130A 0.0130	1.0600A 0.0040A 0.0060	4.3200 0.1600 0.1100	1.2700A 0.0140A 0.0230	1.0600A 0.0070A 0.0120
нү	ALBANY FRIE JEFFERSON MOHROE NEH YORK NIAGARA OHEIDA OHORDAGA HESTCHESTER	0.1700 0.4300 0.0100 0.5700 0.2700 0.2300 0.4600 0.2700			1.2700 1.3800 0.2600 1.2700 2.3700 0.7500 1.9000 2.7000 2.1400		
NC	DARE DURIIAM FORSYTH GUIL FORD MECKLENBURG	0.0200 0.3600 0.2900 0.3300 0.2300	0.8000  0.8000	 0.7400  0.5900	0.2500 4.0300 2.2200 3.0600 3.8100	0.9230 0.7550	0.8300 0.6270
HD	BURLEIGH	0.3500			0.5700		

TABLE C-5. AIR QUALITY DATA FOR LEAD

	COUNTY	H	HINIHUM (UB/H3)			HAXIHUM (UG/H3)			
STATE		089	ARITH HEAH	BEO HEAN	098	ARITH MEAN	GEO HEAN		
OH	CUYAHOGA FRAIXLIN HANILTON JEFFERSON LUCAS	0.4300 0.2700 0.3600 0.1300 6.2600	0.7900A 0.8200A	0.7100A 0.7500A	1.8700 2.3100 1.8000 1.4200 1.7200	0.7900A 0.8200A	0.7100A 0.7500A		
	MAIONING HONTGONERY SCIOTO SUINIT	0.2500 0.4300 0.1300 0.2800	0.6100 0.9800 0.4100 0.5700A	0.5500 0.8600 0.3600 0.5500A	1.4700 2.7500 1.0400 1.2000	0.6100 0.9800 0.4100 0.5700A	0.5500 0.8600 0.3600 0.5500A		
OK	CHEROKEE OKLAHOHA TULSA	0.0500 0.0100 0.0005	0.1810 0.5300A	0.0650 0.4700A	<b>e</b> .2100 38.0996 1.4200	1.9120 0.5300A	1.5170 0.4700A		
OR	CURRY MULTNOMAH	0.0020 0.0020	0.0300A 0.8300A	0.0100A 6.6600A	0.0700 4.2300	0.0300A 0.8300A	0.0100A 0.6600A		
PA	ALLEGHENY BEAVER BERKS BLAIR BUCKS CAIBRIA CHESTER CLARION DAUPHIN ERIE INDIANA LACKANANNA LANCASTER LEHIGH LUZERNE LYCONING MERCER HONTGONERY NORTHANPTON HORTHUNDERLAND YORK	0.5240 0.5860 0.2900 0.0010 0.1600 0.0020 0.1600 0.2400 0.0200 0.1600 0.1700 0.2800 0.3230 0.2340 0.2500 0.1680 0.2600 0.1660	0.8100A  0.5100 1.0400 0.6000A 2.0500A 0.7900A 0.7900  0.6600A 1.2480	0.7400A  0.4700 0.9000 0.3000A 1.8500A 0.7100A 0.7200  0.6100A 1.1790 0.6600A	3.1100 2.8420 6.4380 2.8250 2.2600 3.2430 1.3800 0.4400 2.6000 6.6100 2.7500 2.6600 1.7820 1.5640 2.1700 1.4550 0.3300 2.7200 2.5890 2.2630	0.8100A  0.5670 1.0400 0.6000A 2.0500A 0.7900A 0.8270  0.6600A 1.3280	0.7400A  0.4860  0.9000 0.3000A  1.8500A 0.7100A 0.7820  0.6100A 1.2100		
RI	PROVIDENCE	0.3800		~~~	2.0300				

TABLE C-5. AIR QUALITY DATA FOR LEAD

		<u> </u>	IININUH (UG/	rt3 )	HAXIHUH (UG/H3)			
STATE	COUNTY	CBS	ARITH HEAN	GEO MEAN	QBS	ARITH MEAN	GEO MEAN	
	HASHINGTON	0.0300			0.6900			
SC	CHARLESTON	0.8450			0.8450			
	GREENVILLE	0.3500	1.0500	0.8600	3.4500	1.1320	0.9380	
	RICHLAND	0.0500			4.1500			
SD	CUSTER	0.0			0.0500			
	HIHHHHHA	0.0200			1.6200			
TH	ANDERSON	0.5800			1.3100		*	
•••	BEDFORD	0.2800			0.6400			
	BLOURT	0.6500			1.5000			
	BRADLEY	0.2000			0.4900			
	CAMPBELL	0.4500			0.8100			
	CARTER	0.5300			0.5300			
	COFFEE	0.1600			1.6200			
	CUMBERLAND	0.0100	0.0900A	A0080.0	0.5000	0.0900A	0.0800A	
	DAVIDSON	0.2500			2.4600			
	DYER	0.1600			0.7700	~~~		
	GIBSON	0.1400			0.4500			
	GREEN	0.4300			0.6500			
	HANBLEN	0.1300			2.7400			
	HEHRY	0.0200			0.7100			
	HUHPHREYS	<b>0</b> .0300			0.2100			
	KKOX	0.3700 0.2800			3.9000 0.2300			
	LINCOLN	0.2800 0.1800			0.2300 0.1800			
	NCMINN MADISON	0. 1300 0. 1300			2.3300			
	MARION	0.1500 0.1500			0.2208			
	MAURY	0.4000			2.4700			
	MONTGOMERY	0.2300			0.9200			
	OBION	0.2400			0.5700			
	POLK	0.3700			1.4200			
	PUTHAM	0.1700			0.7800			
	ROANE	0.3300			1.9000			
	ROBERTSON	0.2300	~~~		0.4500			
	RUTHERFORD	0.2400			0.8200			
	SHELBY	0.2500	~~~		5.5700			
	SULLIVAN	0.2300			2.2900			
	SUITIER	0.2100			0.4900			
	HARREN	0.2200			0.4200			
	HASHINGTON	0.6300			0.6300			
	HILLIANSON	0.2700			0.2700			

TABLE C-5. AIR QUALITY DATA FOR LEAD

			HININUN (UG/H3)			HAXIHUH (UG/H3)			
STATE	COUNTY	OB\$	ARITH HEAN	GEO MEAN	088	ARITH MEAN	GEO HEAN		
	MIT20M	0.2500		***	0.2900				
τx	MILSON  BEE BEXAR BOHIE BRAZORIA BRAZOS BROWN CALHOUN CAHERON CHANBERS DALLAS DENTON ECTOR ELLIS EL PASO GALVESTON GRAY GRAYSON HALE HARRIS HAYS HIDALGO HOHARD JEFF DAVIS	0.2500 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200	0.4400A 0.4200A 0.1400 0.3700A  0.0500  0.2300 0.4400A  0.0900 0.4500A  0.1500  0.0300 0.0700	0.2800A 0.3300A 0.0900 0.3000A  0.0300 0.1700 0.3900A 0.0500 0.3800A  0.0900 0.0500	0.2900 0.1300 4.3300 1.9500 0.7400 0.8700 1.8100 0.1300 1.2300 8.0200 1.0700 1.5000 4.0900 1.2000 0.1300 1.2300 0.1300 1.2500 0.12500 0.0200	0.4400A 0.4200A 0.1900 0.3700A  0.1700  2.9280 0.4400A 1.1100 0.4500A  0.8500  0.2800 0.0700	0.2800A 0.3300A 0.1200 0.3000A  0.1000  2.6310 0.3900A  1.0200 0.3800A  0.6700 0.2100 0.0500		
	JEFFERSON LUDBOCK MCLEMIAN MCHULLEN MATAGORDA MAVERICK MIDLAND MONTSONERY MOORE NACOGDOCHES NAIECES ORANGE POTTER SAN PATRICIO SCURRY SMITH TARRANT TAYLOR	0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200 0.0200	0.1300  0.4900A 0.1200A  0.3600 0.1100A  0.5000A 0.6910A 0.1700A	0.0800  0.3900A 0.0900A  0.2400 0.0700A  0.3900A 0.5340A 0.1300A	0.9200 1.9600 0.7900 0.3300 0.2100 1.9000 0.6000 0.7000 0.1400 0.5500 17.3000 0.7200 1.4800 0.1100 0.2700 1.4000 3.8000 0.7400	0.5200  0.4900A 0.1200A  0.6100 0.1100A  0.5000A 0.8650A 0.1700A	0.5000  0.3900A 0.0900A  0.5300 0.0700A  0.3900A 0.8060A 0.1300A		

TABLE C-5. AIR QUALITY DATA FOR LEAD

			4ININUM (UG/	M3)	HAXIHUM (US/H3)		
STATE	COUNTY	OBS	ARITH MEAN	GEO MEAN	ocs	ARITH MEAN	GEO HEAN
	TITUS	0.0200	G. 1100A	0.0700A	0.5800	0.1100A	0.0700A
	TON GREEN	0.0200	0.1000A	0.0700A	0.6500	0.1000A	0.0700A
	TRAVIS	0.0200	0.0400	0.0300	2.6000	0.7300	0.6400
	VAL VERDE	0.0200			0.6000		
	VICTORIA	0.0200			0.8300		
	HALKER	0.0200			0.3300		
	HEBB	0.0208			0.7100		
	HICHITA	0.0200	0.4600A	0.4000A	1.2100	0.4600A	0.4000A
	HISE	0.0200			0.6500		
UT	EHERY	0.0010			0.1600		
	GARFIELD	0.0010	0.0150	0.0070	0.1700	0.0170	0.0080
	KANE	0.0010	0.0170A	0.0630A	0.0900	0.0170A	0.0080A
	SALT LAKE	0.3500			4.9100		
	SAN JUAN	0.0010	0.0110	0.0050	0.1400	0.0200	0.0080
	HASHINGTON	0.0010	0.0190	0.0070	0.2400	0.0300	0.0160
	HEBER	0.2300			3.5500		
VT	CHITTENDEN	0.0020	0.6860	0.4710	1,2600	0.7900	0.7600
	ORANGE	0.0500			1.1800		
VA	0000	0.1300	0.5200	0.4400	3.7500	0.9700	0.8600
	FAIRFAX	0.2100			2.1400		
	PAGE	0.0300	0.2400A	0.1900A	0.8100	0.2400A	0.1900A
	PITTSYLVANIA	0.1800	0.5700A	0.4800A	1.9300	0.5700A	0.4800A
	HYTHE	0.0200	0.0900	0.0200	0.1900	0.0900	0.0800
НА	KING	0.0700	1.4600A	1.3100A	4.4800	1.4600A	1.3100A
	• PIERCE	0.1600	0.9500A	0.8200A	2.1900	0.9500A	0.8200A
	SPOKANE	0.1800			1.2900		
HV	CABELL	0.2200			2.2900		
	KANANHA	0.1200	0.5200	0.4400	2.6900	0.6240	0.7200
HI	DANE	0.1400	0.6000A	0.5200A	1.3000	0.6000A	0.5200A
	DOOR	0.0			0.5500		
	DOUGLAS	0.0300	0.2300A	0.1900A	1.0700	0.2300A	0.1900A
	EAU CLAIRE	0.1500			0.9800		
	KENOSHA	0.0800			1.0300		
	MILHAUKEE	0.3200			1.6100		
	RACINE	0.1300	0.4200A	0.3500A	1.4700	0.4200A	0.3500A
нү	LARAHIE	0.1100			0.6600		

TABLE C-5. AIR QUALITY DATA FOR LEAD

		· ·	HINIHUH (UG/H3)			HAXINUM (UG/M3)		
STATE	COUNTY	OBS	ARITH HEAN	GEO HEAN	089	ARITH HEAH	GED NEAH	
	HATRONA	0.800			0.4400		***	
	PARK	0.0			0.0300			

TABLE C-6. AIR QUALITY DATA FOR MANGANESE

	COUNTY		Minimum (UG/H3)			MAXIMUN (UG/M3)		
STATE		OBS	ARITH MEAN	GEO MEAN	GBS	ARITH MEAN	GEO MEAN	
AZ	APACHE COCONINO HARICOPA HOHAVE NAVAJO	0.0 0.0 0.0 0.0	0.0060 0.0040 0.0310 0.0050 0.0080A	0.0020 0.0010 0.0200 0.0010 0.0020A	0.1200 0.1300 3.0840 0.1500 0.1900	0.0130 0.0160 0.1480 0.0090 0.0080A	0.0040 0.0080 2.7652 0.0030 0.0020A	
CO	LA PLATA MONTEZUNA -	0.0 0.0	0.0060A 0.0070A	0.0020A 0.0020A	0.0700 0.0400	0.0210A 0.0070A	0.0100A 0.0020A	
IN	ALLEN BARTHOLOMEH CLARK DUBDIS ELKHART GRANT HOMARD JASPER JEFFERSON KNOX LAKE LA PORTE MARION HOMROE ST. JOSEPH STEUBEN TIPPECAMOE VANDERBURGH VIGO HAYNE	0.0170 0.0060 0.0230 0.0130 0.0120 0.0150 0.0250 0.0060 0.0070 0.0080 0.0050 0.0180 0.0090 0.0150 0.0180 0.0180 0.0180			0.0440 0.0530 0.1360 0.0570 0.0530 0.0400 0.2500 0.0170 0.0230 0.0400 0.1010 0.0770 0.0510 0.0450 0.0450 0.0450 0.0330 0.1150 1.1320			
но	ADAIR AUDRAIN BOOHE BUCHANAN BUTLER CALLANAY CANDEN CAPE GIRARDEAU CLAY COLE JASPER JEFFERSON	0.0260 0.0290 0.0210 0.0040 0.0190 0.0190 0.0130 0.0160 0.0300 0.0210 0.0330			0.1020 0.0600 0.1120 0.1220 0.1150 0.0700 0.0620 0.0540 0.1250 0.0510 0.1410			

TABLE C-6. AIR QUALITY DATA FOR MANGAMESE

	COUNTY	HINIMUM (NB/H2)			HAXIMUM (UG/H3)			
STATE		08\$	ARITH HEAN	GEO MEAN	083	ARITH MEAN	GEO NEAN	
	LIVINGSTON	0.0310			0.8280			
	MARION	0.0190			0.0690			
	NEH HADRID	0.0190			0.2720			
	HODAHAY	0.0200			0.0710			
	PETTIS	0.0310			0.0610			
	Phelps	0.0110			0.0730			
	PLATTE	0.0230			0.0960			
	ST. CHARLES	0.0230			0.1120			
	STE. GENEVIEVE	0.0610		***	0.0610			
	SCOTT	0.0020			0.3640			
	VERNON	0.0340			0.0820			
HH	RIO ARRIBA	0.0	0.0040A	0.0010A	0.0400	0.0040A	0.0010A	
SC	CHARLESTON	0.0170			0.0170			
TN	ANDERSON	0.0320			0.0420			
	BEDFORD	0.0160			0.0240			
	BLOUNT	0.0310			0.0450			
	BRADLEY	0.0100			0.1290			
	CAMPBELL	0.0200			0.0340			
	CARTER	0.0200			0.0200			
	COFFEE	0.0030			0.0330			
	CUMBERLAND	0.0040			0.0250			
	DYER	0.0060			0.1260			
	GIBSON	0.0040			0.0300			
	GREENE	0.0280			0.0340			
	HAMBLEN	0.0160			0.0700			
	· HENRY	0.0060		'	0.0260			
	HUMPREYS	0.0120			0.6960			
	LINCOLN	0.0360			0.0360			
	HCHIHI	0.0190			0.0190			
	MADISON	0.0160			0.0350			
	MARION	0.0270			0.0290			
	HAURY MONTGONERY	0.0310			0.0810			
	OBION	0.0080			0.0410			
	POLK	0.0160 0.0370		~	0.1030		~~~	
	PUTNAM	0.0370			0.0440			
	ROAHE	0.0610			0.0180			
	RODERTSON	0.2500			0.9720			
	RUTHERFORD	0.2300 0.0100			0.0270			
	SULL IVAN	0.0060			0.0240			
	· ·	v. 000 <b>0</b>			0.1670			

TABLE C-6. AIR QUALITY DATA FOR MANGANESE

	COUNTY	H	INIMUM (UG/	M3)	MAXINUM (UG/H3)			
STATE		OBS	ARITH HEAH	GEO MEAN	OBS	ARITH MEAN	GEO MEAN	
	SUNKER	0.0180			0.0250			
	HARREN	0.0070			0.0250			
	HASHINGTON	0.0350			0.0350			
	HILLIAMSON '	0.0370			0.0370			
	Hilson	0.0100			0.0230			
TX	BEE	0.0200			0.0900			
•	BEXAR	0.0200			0.1400			
	BOHIE	0.0200			0.1000			
	BRAZORIA	0.0200			0.1600			
	BRAZOS	0.0200			0.0900			
	BROUN	0.0200			0.0700			
	CALHOUN	0.0200			0.0900			
	CAMERON	0.0200			0.3100			
	CHANBERS	0.0200			0.1000			
	DALLAS	0.0200			0.1900			
	DENTON	0.0200			0.0700			
	ECTOR	0.0200			0.3100			
	ELLIS	0.0200			0.0200		~~-	
	EL PASO	0.0200			0.2000			
	GALVESTON	0.0200			0.1300			
	GRAY	0.0200			0.1500			
	6RAYSON	0.0200			0.0700			
	HALE	0.0200			0.0300			
	HALE	0.0			1.1500			
	HAYS	0.0200			0.1400			
	HIDALGO	0.0200			0.1500			
	HORARD	0.0200	~~~		0.0900			
	· JEFF DAVIS	0.0200			0.1400			
	JEFFESON	. 0.0200			0.1400			
	LUBBOCK	0.0200			0.1000			
	HCLENNAN	0.0200			0.0800			
	HCHULLEN	0.0200			0.0300			
	MATAGORDA	0.0200			0.0200			
	MAVERICK	0.0200			0.0700			
	HIDLAND	0.0200			0.3700			
	HOLITGOLIERY	0.0200			0.1000			
	MODRE	0.0200			0.1600			
	NACOGDOCHES	0.0200			0.0200			
	HUECES	0.0200			2.6600			
	ORANGE	0.0200			0.2200-			
	POTTER	0.0200			0.1600			
	SAN PATRICIO	0.0200			0.0200			

TABLE C-6. AIR QUALITY DATA FOR MANGANESE

STATE			HININUN (UG/H3)			HAXIMUM (UG/H3)		
	COUNTY	083	ARITH MEAN	GEO MEAN	039	ARITH MEAN	GEO ITEAN	
	SCURRY	0.0200			0.0600			
	SHITH	0.0200			0.0900			
	TARRANT	0.0200			0.1000			
	TAYLOR	0.0200		~	0.0500			
	TITUS	0.0200			0.0700			
	TCH GREEN	0.0200			0.1700			
	TRAVIS	0.0200			0.1000			
	VAL VERDE	0.0200			0.0200			
	VICTORIA	0.0200			0.0700			
	HALKER	0.0200			0.0600			
	HEBB	0.0200			0.0600			
	HICHITA	0.0200			0.0300			
	HISE	0.0200			0.0600			
บา	EMERY	0.8			0.0400			
	GARFIELD	0.0	0.0070	0.0020	0.0500	0.0140	8.0870	
	KANE	0.0	0.0050A	0.0010A	0.0400	0.0050A	0.0010A	
	San Juan	0.0	0.0040	0.0010	0.0300	0.0100	0.0030	
	HASHINGTON	0.0	0.0070	0.0030	0.0800	0.0078	0.0030	

TABLE C-7. AIR QUALITY DATA FOR SELENIUM

	COUNTY		MINITHUM (UG/H3)			MAXIMUM (UG/M3)			
STATE		088	ARITH HEAN	GEO MEAN	OBS	ARITH MEAN	GEO MEAN		
ΤX	BEE	0.0100			0.0100				
•••	BEXAR	0.0100	0.0100A	0.0200A	8.0400	0.0100A	0.0200A		
	BOHIE	0.1000	0.0100A	0.0200A	0.0300	0.0100A	0.02004		
	BRAZORIA	0.0100	0.0100	0.0200	0.0100	0.0100	0.0200		
	BRAZOS	0.0100	0.0100A	0.0200A	0.0100	0.0100A	0.0200/		
	BRONII	0.0180			0.0300				
	CALHOUN	0.0100			0.0100				
	CAMERON	0.0100	0.0100	0.0200	0.0100	0.0100	0.0200		
	CHALBERS	0.0100			0.0300				
	DALLAS	0.0100	0.0100	0.0200	0.0100	0.0100	0.0200		
	DENTON	0.0100			0.0500				
	ECTOR	0.0100	0.0100A	0.0200A	0.0100	0.0100A	0.0200		
	ELLIS	0.0100			0.0400				
	EL PASO	0.0100	0.0100	0.0200	0.0100	99.0000	99.0000		
	GALVESTON	0.0100			0.1300				
	GRAY	0.0100			0.0100				
	6RAYSON	0.0100			0.0300				
	HALE	0.0100			0.0100				
	HARRIS	0.0100	0.0100	0.0200	0.0300	0.0100	0.0200		
	HAYS	0.0100			0.0300				
	HIDALGO	0.0100	0.0100	0.0200	0.0100	0.0100	0.0200		
	HOHARD	0.0100	0.0100A	0.0200A	0.0100	A0010.0	0.0200		
	JEFF DAVIS	0.0100			0.0100				
	JEFFERSON	0.0100	0.0100A	0.0200A	0.0100	0.0100A	0.0200		
	LUBBOCK	0.0100			0.0100				
	MCLENHAN	0.0100			0.0400				
	HCHULLEH	0.0100			0.0100				
	<ul> <li>NATAGORDA</li> </ul>	0.0100			0.0100				
	MAVERICK	0.0100	0.0100A	0.0200A	0.0100	0.0100A	0.0200		
	MIDLAHF	0.0100	0.0100A	0.0200A	0.0100	0.0100A	0.0200		
	HONTGONERY	0.0100			0.0100				
	MOORE	0.0100			0.0100				
	NACOGDOCHES	0.0100			0.0100				
	HUECES	0.0100	0.0100	0.0200	0.0300	0.0100	0.0200		
	ORANGE	0.0100	0.0100A	0.0200A	0.0100	Q.0100A	0.0200		
	POTTER	0.0100			0.0100				
	SAN PATRICIO	0.0100			0.0100				
	SCURRY	0.0100			0.0100				
	SHITH	0.0100	0.0100A	0.0200A	0.0300	0.0100A	0.0200		
	TARRANT	0.0100	9.0000	99.0000	0.0300	99.0000	99.0000		
	TAYLOR	0.0100	0.0100A	0.2000A	0.0100	0.0100A	0.2000		
	TITUS	0.0100	0.0100A	0.0200A	0.0100	0.0100A	0.0200		

TABLE C-7. AIR QUALITY DATA FOR SELENIUM

STATE		HININUM (UG/H3)			HAXIHUH (UG/II3)		
	COUNTY	083	ARITH HEAN	GEO HEAN	OBS	ARITH HEAM	GEO MEAN
	TOH GREEN	0.0100			0.0100		
	TRAVIS	0.0100	0.0100	0.0200	0.0500	8.0200	0.0200
	VAL VERDE	8.0100			0.0100		
	VICTORIA	0.0100	~~~		0.0400		
	HALKER	0.6100			0.0100		
	HEBB	0.0100			0.0100		
	HICHITA	0.0100			0.0300	~~~	
	HISE	0.0100			0.0100		

TABLE C-8. AIR QUALITY DATA FOR VAHADIUM

	COUNTY	н	HINIPUM (UG/H3)			HAXIHUN (UG/H3)			
STATE		OBS	ARITH HEAN	6EO MEAN	0B <b>S</b>	ARITH MEAN	GEO MEAN		
TN	AHDERSON	0.0200		***	0.0200				
***	BEDFORD	0.0200			0.0200				
	BLOUNT	0.0200			0.0200				
	BRADLEY	0.0200			0.0200				
	CAMPBELL	0.0200	~		0.0200				
	CARTER	0.0200			0.0200				
	COFFEE	0.0200			0.0200				
	CUMBERLAND	0.0200			0.0200				
	DYER	0.0200	~		0.0200				
	GIBSON	0.0200		~~~	0.0200		~		
	GREENE	<b>0</b> .020 <b>0</b>			0.0200				
	HANBLEN	0.0200			0.0200				
	HEIRY	0.0200			0.0200				
	HUNPREYS	0.0200			0.0200				
	LINCOLN	0.0200			0.0200				
	MCHINN	0.0200			0.0200				
	MADISON	0.0200			0.0200				
	MARION	0.0200			0.0200				
	HAURY	0.0200			0.0200				
	HONTGOMERY	0.0200			0.0200				
	OBION	0.0200			0.0200				
	POLK	6.0200			0.0200				
	PUTNAM	0.0200			0.0200				
	ROAME	0.0200			0.0200				
	ROBERTSON	0.0200			0.0200				
	RUTHERFORD	0.0200			0.0200				
	SULLIVAN	0.0200			0.0200				
	· SUITHER	0.0200			0.0200				
	HARREN	0.0200			0.0200				
	Hashirigtoh	0.0200			0.0200				
	Hillianson	0.0200			0.0200				
	HILAON	0.0200			0.0200				
TX	BEE	0.0010			0.0020				
	BEXAR	0.0010			0.0010				
	BONIE	0.0010			0.0010				
	BRAZORIA	0.0010			0.0020				
	BRAZOS	0.0010			0.0030				
	BROUN	0.0010			0.0040				
	CALHOUN	0.0010			0.0010				
	CAHEROH	0.0010			0.0020				
	CHANDERS	0.0010			0.0010				

		HIHIHUH (UG/H3)			MAXIMUN (UG/H3)		
STATE	COUNTY	OBS	ARITH HEAH	GEO MEAN	OBS	ARITH HEAN	GEO MEAN
	DALLAS	0.0010			0.0020		
	DENTON	0.0010			0.0010		
	ECTOR	0.0010			8.0010		
	ELLIS	0.0010			0.0010		
	EL PASO	0.0010			0.0020		
	GALVESTON	0.0010			0.0070		
	GRAY	0.0010			0.0010		
	GRAYSON	0.0010			0.0020		
	HALE	0.0010			0.0010		
	HARRIS	0.0010			0.0030		
	HAYS	0.0010			0.0010		
	HIDALGO	0.0010			0.0010		
	HO!IARD	0.0010			0.0020		
	JEFF DAVIS	0.0010			0.0010		
	JEFFERSON	0.0010			0.0230		
	LUBBOCK	0.0010			0.0020		
	HCLEHNAN	0.0010			0.0020		
	HCHULLEN	0.0010			0.0010		
	HATAGORDA	0.0010			0.0010		
	MAVERICK	0.0010			0.0010		
	HIDLAND	0.0010			0.0010		
	MONTGONERY	0.0010			0.0010		
	HOORE	0.0010					
	NACOGDOCHES	0.0010			0.0010		
	NUECES	0.0010			0.0010		
	ORANGE				0.0010		
	POTTER	0.0010 0.0010			0.0020		
	SAN PATRICIO	0.0010			0.0010		
	· SCURRY	0.0010			0.0010		
					0.0010		
	SHITH	0.0010			0.0030		
	TARRANT	0.0010			0.0100		
	TAYLOR	0.0010			0.0010		
	TITUS	0.0010			0.0010		
	TON GREENE	0.0010			0.0010		
	TRAVIS	0.0010			0.0030		
	VAL VERDE	0.0010			0.0010		
	VICTORIA	0.0010			0.0010		
	HALKER	0.0010			0.0010		
	HEGB	0.0010	·		0.0010		
	HICHITA	0.0010			0.0010		
	HISE	0.0010			0.0020		

TABLE C-9. AIR QUALITY DATA FOR ZINC

		HINIMUM (UG/M3)			MAXINUN (UG/M3)		
STATE	COUNTY	085	ARITH HEAN	6EO HEAN	085	ARITH HEAH	GEO NEAN
AZ	APACHE	0.0100			0.2800		
	COCONINO	0.0200			0.1500		
	GILA	0.0100			0.1900		
	GRAHAH	0.0200			0.0980		
	GREENLEE	0.0300			0.2300		
	HARICOPA	0.0001			0.8410		
	MOHAVE	0.0001	~~-		0.1700		
	OLAVAH	0.0100			0.1400		
	PIHA	0.0001			0.3000		
	PINAL	0.0300			0.2200		
	SANTA CRUZ	0.1100			0.1100		
	YAVAPAI	0.0100			0.1800		
	YUMA	0.0200			0.1000		
0	SHOSHOHE	0.0100			8.9000		
[N	ALLEN	0.0550			0.1590		
	BARTHOLOMEH	0.0480			0.1140		
	CLARK	0.0924			0.6551		
	DUBOIS	0.0478			0.2774		
	ELKHART	0.0590			0.0780		~
	GRANT	0.0772			0.2285		
	HORARD	0.2010			1.4440		
	JASPER	0.0543			<b>0.1</b> 36 <b>8</b>		
	JEFFERSON	0.0215			0.0976		
	KNOX	0.0410			0.0851		
	LAKE	0.0410		*	0.3990		
	· LA PORTE	0.0634			0.1971		
	HARIOH	0.1000			0.4960		
	MONROE	0.0430			0.0960		
	ST. JOSEPH	0.0645			0.4068		
	STEUBEN	0.0602			0.2873		
	TIPPECANOE	0.0687			0.0995		
	VANDERBURGH	0.0520			0.1720		
	A1e0	0.1391			0.4640		
	HAYHE	0.0736			0.3921	+	
C	CHARLESTON	0.3500			0.3500		
X	BEE	0.0			0.0300		
	BEXAR	0.0	0.0200A	0.0100A	0.2100	0.0200A	0.0100
	BOHIE	0.0	Q.0500A	0.0300A	0.7400	0.0500A	0.0300

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MAXIMUM (UG/H3)

STATE	COUNTY	OBS	ARITH NEAN	GEO HEAN	ODS	ARITH HEAR	GEO MEAN
	BRAZORIA	0.0	0.0300	0.0200	0.5300	0.0400	0.0300
	BRAZOS	6.0	0.0400A	0.0300A	0.0900	0.0400A	0.0300A
	BROLIN	0.0			0.4500		~
	CALHOUN	0.0			0.0600		
	CAHERON	0.0	0.0100	0.0100	0.4200	0.0408	0.0200
	CHAMBERS	0.0			0.7400		
	DALLAS	0.0	0.0200	<b>6</b> .0100	0.2400	0.0200	0.0200
	DENTON	0.0			0.1000		
	ECTOR	0.0	0.0600A	0.0400A	<b>0</b> .1600	0.0600A	0.0400A
	ELLIS	0.0			0.1400		
	EL PASO	0.0	0.0500	0.0200	2.3400	0.1500	<b>G.</b> 0700
	GALVESTON	0.0	0.0800A	0.0700A	6.7000	0.080GA	0.0780A
	GRAY	0.0100			0.0500		
	6RAYSON	0.0			<b>0.1</b> 600		
	HALE	0.0100			0.0300		
	Harris	0.0	0.6400	0.0200	2.0600	<b>0</b> .1300	0.0800
	HAYS	0.0			0.1000		
	HIOAL60	0.0	0.0100	0.6100	0.2200	0.0200	0.0100
	HOHARD	0.0	0.0200A	0.0100A	0.1700	0.0200A	0.0100A
	JEFF DAVIS	8.6			0.1000		
	JEFFERSON	9.0	0.0400A	0.0200A	0.4800	0.0400A	0.0200A
	ŁUBBOCK	0.0			0.3100		
	HCLENNAN	0.0			0.1200		
	HCHULLEH	0.0			0.1800		'
	HATAGORDA	0.0			0.0600		
	MAVERICK	0.0	0.0500A	0.0300A	0.3600	0.0500A	0.0300A
	HIDLAND	0.0	0.0400A	0.0200A	0.1300	0.0400A	0.0200A
	HONTGOHERY	0.0200	'		0.1000		
	HOORE	0.0			0.0900		
	NACOGDOCHES	0.0300			0.1000		
	NUECES	0.0	0.0300	0.0200	28.6100	2.5200	1.1500
	ORANGE	0.0	0.0400A	<b>0.</b> 0300A	0.4700	0.0400A	A0020.0
	POTTER	0.0			0.1800		
	, SAN PATRICIO	0.0			0.0400		
	SCURRY	0.0			0.1200		
	Shith	0.0	0.0400A	O.0300A	0.1700	0.0400A	A0020.0
	TARRANT	0.0			1.5000		
	TAYLOR	0.0	0.0300	0.0200	0.1600	0.0300	0.0200
	TITUS	0.0	0.0400A	0.0200A	0.3600	0.0400A	0.0200A
	TOH GREEN	0.0			1.1400		
	TRAVIS	0.0	0.0100	0.0100	1.8300	0.0500	0.0400
	VAL VERDE	0.0			0.0900		
	VICTORIA	0.0			0.0700		

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# APPENDIX D

Effects of Deposited Particulate Matter

## APPENDIX D

## EFFECTS OF DEPOSITED PARTICULATE MATTER

Most evidence for particulate toxicity is derived from studies of domestic animals. It is often not clear if the symptoms of toxicity are the result of ingestion, inhalation, or both. Only those studies which clearly indicated ingestion of dust-covered vegetation are summarized here. There appears to be a definite relationship between deposition of fine particles of arsenic, fluoride, lead, and copper on vegetation; their ingestion by animals; and chronic or acute injury to animals. 59,60 Other metals which may also be implicated are zinc and cadmium. The surfaces of vegetation, especially those covered with fine hairs (stems, leaf petioles, and blades), provide a major filtration and reaction surface for metal-laden particles of 1-5 µm and less. 61

Fluorides are reported to cause more damage to domestic animals than any other air pollutant.<sup>62</sup> Dietary fluoride is generally accepted as the major source of fluorosis in animals.<sup>9</sup> Fluorosis has been noted in most domestic livestock, presumably resulting from particulate fluoride deposited on vegetation and ingested by animals.<sup>63,64</sup> For cattle, the most susceptible domestic animal,<sup>26,65,66</sup> diets containing concentrations exceeding 40 ppmw fluoride may have severe toxic effects.<sup>67</sup> The safe range for soluble and insoluble fluorides has been specified at 30-50 ppmw and 60-100 ppmw, respectively, for cattle.<sup>68</sup> Sheep and swine (70-100 ppmw), chickens (150-300 ppmw), and turkeys (300-400 ppmw) are less sensitive to dietary fluoride levels.<sup>68</sup>

Arsenic deposited on vegetation from smelting operations has been known to kill livestock if enough was ingested. 62,69-72 Ingestion of arsenic-contaminated dust/soil on forage presents the greatest dangers to grazing animals. 73 However, a wide range of toxicity for arsenic compounds exists and is correlated to animal excretion rates. 9 The reported biological half-life of arsenic compounds ranges from 30-60 hours. 74,75 Those compounds excreted most rapidly tend to be least toxic.

Lead poisoning of cattle, horses, and other grazing animals as a result of ingestion of contaminated forage has been reported often. 76-80 Fodder contaminated by lead and zinc by atmospheric deposition from a foundry was responsible for the death and slaughter of 140 cows. 81 Ingestion of surface

deposits of airborne lead on forage, especially adjacent to heavily traveled highways, 82 and inhalation of lead can be significant additions to the total body burden of animals. 83,84 Zoo animals (cats, primates, and snakes) are susceptible to lead aerosol poisoning. 85

Several other elements have been implicated in the illness or death of grazing animals when deposited on forage. For example, soot containing vanadium at a concentration of 1 µg/g was dumped near a pasture and subsequently spread by wind. The pasture grass was covered with a film of soot and, when ingested by cattle, caused sickness and death.86,87 Iron particles from a magnesium plant in Austria adversely affected the digestive tracts of grazing cattle.88 Domestic animals grazing near specialized steel and alloy plants have been poisoned by ingesting dust containing molybdenum with vegetation.62

No specific studies are known which address the intake of trace elements by wildlife through ingestion of dust-coated vegetation. Fluorosis has been identified in wild animals, especially deer and elk. Honey bees, red deer, and wild hares are known to be especially sensitive to arsenic poisoning. Newman poisoning a state-of-knowledge review of the effects of industrial air pollution on wildlife. However, specific information dealing with surface-contaminated forage could not be identified. Ingestion of plant material which has concentrated heavy metals through uptake and inhalation of airborne particulates represents the majority of case histories.

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#### 15. SUPPLEMENTARY NOTES

#### 16. ABSTRACT

Section 165 of the Clean Air Act requires preconstruction review of major emitting facilities to provide for the prevention of significant deterioration (PSD) and charges Federal Land Managers (FLMs) with an affirmative responsibility to protect the air quality-related values of Class I areas. Regulations implementing these provisions require an analysis of the impairment to visibility, soils, and vegetation (52.21(o)).

The information and screening procedure presented here provide interim guidance: (1) to aid in determining whether emissions are significant or whether there are significant air quality impacts under Section 52.21(o). and (2) to aid in flagging sources which should be brought to the attention of an FLM under Section 52.21(p).

Impacts on vegetation and soils are the principal areas addressed by the procedure which thus takes a limited view of the possibly broad scope of air quality-related values. A selected review of impacts on fauna has also been included and the odor potential of regulated pollutants is addressed. This procedure is intended for use by air quality engineers and is not a manual for the assessment of impacts on plants, soils, and other air quality-related values such

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